

# TOWARD A SHARED UNDERSTANDING OF CLIMATE-SMART RESTORATION ON AMERICA'S NATIONAL FORESTS

A Science Review and Synthesis



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AMERICAN FORESTS  
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## A Science Review and Synthesis

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**Cover image:** Ensuring the sustainability and resilience of national forests in the face of rapid climate change will require a dramatic increase in the pace, scale, and quality of ecologically appropriate and climate-smart forest restoration (Shasta-Trinity National Forest, California). Photo: Rick Hanover/USFS.

*Toward a Shared Understanding of Climate-Smart Restoration on America's National Forests* is available online at [www.nwf.org/ClimateSmartRestoration](http://www.nwf.org/ClimateSmartRestoration)

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1200 G Street, NW  
Washington, DC 20005  
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# TABLE OF CONTENTS

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<b>EXECUTIVE SUMMARY</b> .....	<b>ii</b>
<b>1. INTRODUCTION</b> .....	<b>1</b>
1.1 Purpose of this Review .....	3
<b>2. AMERICA’S FORESTS FACE A CHANGING FUTURE</b> .....	<b>4</b>
2.1. Understanding Climate Change and Its Impacts on Forest Ecosystems .....	5
2.2. Addressing Climate-related Uncertainty .....	14
<b>3. IMPLICATIONS OF CLIMATE CHANGE FOR NATIONAL FOREST MANAGEMENT</b> .....	<b>16</b>
3.1 U.S. Forest Service Response to Climate Change .....	18
3.2 Proposed Principles for Climate-Smart Forest Restoration .....	19
<b>4. RESTORATION IN THE CONTEXT OF FUTURE CHANGE</b> .....	<b>22</b>
4.1 Looking to the Future While Learning from the Past .....	23
4.2 Embracing Functional Restoration of Ecological Integrity .....	25
4.3 Restoring and Managing Forests in the Context of Larger Landscapes and Longer Time Frames .....	26
4.4 Adopting Agile Planning and Management Approaches that Accommodate and Address Uncertainty .....	28
<b>5. MANAGING NATIONAL FORESTS FOR ADAPTATION AND RESILIENCE</b> .....	<b>30</b>
5.1 Addressing Climate Risks by Linking Adaptation Strategies to Key Climate-related Impacts .....	31
5.2 Managing for Change, Not just Persistence .....	31
<b>6. MANAGING NATIONAL FORESTS TO ACHIEVE NATURAL CLIMATE SOLUTIONS</b> .....	<b>41</b>
6.1 Optimizing, Rather than Maximizing, Carbon Sequestration Opportunities .....	43
<b>7. BALANCING TRADE-OFFS AMONG ECOSYSTEM SERVICES</b> .....	<b>45</b>
7.1 Examples of Trade-offs in National Forest Management .....	46
7.2 Enhancing Collaboration to Identify Shared Values, Navigate Trade-offs, and Maximize Synergies .....	50
<b>8. A PATH FORWARD</b> .....	<b>52</b>
<b>9. REFERENCES</b> .....	<b>53</b>



*Assuring a sustainable and resilient future for America's national forests requires that restoration practices look to the future while learning from the past (Deschutes National Forest, Oregon). Photo: Buddy Mays/Alamy.*

## EXECUTIVE SUMMARY

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**A** rapidly changing climate, including rising temperatures, changing precipitation patterns, and more extreme storms, is having profound consequences for America's national forests. Climate-related impacts on forest systems include larger and more severe disturbances (e.g., wildfires, drought, and insect outbreaks), shifts in tree species ranges and forest composition, and changes in forest dynamics and regeneration capacity.

Many of our national forests have been significantly modified by past management and land use, and forest managers are contending with ongoing threats from invasive species, disease outbreaks, and other challenges. With the added impacts on forest systems from climate change, an enormous mismatch exists between the level of restoration work currently underway and the scale of the challenge. As a result, there is a need to substantially increase the pace, scale,

and quality of restoration on our national forests, and to ensure that this restoration is carried out in an ecologically appropriate and climate-smart manner.

Continuing and accelerating climatic changes, and their associated impacts, have significant implications for the effectiveness of traditional forest restoration efforts, including reliance on historical conditions as benchmarks for restoration outcomes. Drawing on a growing body of evidence, research, and experimentation, this science review and synthesis looks at how climate change is inspiring an important evolution in approaches for national forest restoration and management. Over the past decade, the U.S. Forest Service has made considerable progress in understanding the effects of a changing climate on forest ecosystems and working to incorporate climate considerations into its planning and management. Nonetheless, varying perspectives on what climate change means for ecological restoration in practice and how to navigate potential trade-offs continue to pose challenges to integrating climate adaptation and mitigation in national forest planning and management.

Addressing this challenge would benefit from a shared understanding among agency staff and stakeholders of what constitutes a forward-looking and climate-smart approach to national forest restoration. To this end, this report reviews and summarizes recent advances and ongoing evolution in how the concepts and principles of climate adaptation and mitigation can help promote the development and application of climate-smart forest restoration.

To inform further discussion and help advance such a shared understanding, we propose the following overarching principles for climate-smart restoration on America's national forests.

*This report reviews and summarizes recent advances and ongoing evolution in how the concepts and principles of climate adaptation and mitigation can help promote the development and application of climate-smart forest restoration.*

- ***Look to the future while learning from the past.*** Forest planners and their partners should develop forward-looking goals for management that build on an understanding of the historical range of variability and past responses to disturbance, but account for and anticipate future climate-related changes.
- ***Embrace functional restoration of ecological integrity.*** As climatic conditions continue to change, it will become increasingly difficult to restore the ecological integrity of forest systems based on historical species compositions and structures. Rather, goals for ecological integrity should emphasize the capacity of forest systems to adapt and adjust, including through enhancing functional diversity and habitat complexity.
- ***Restore and manage forests in the context of larger landscapes and longer time frames.*** Climate change necessitates that planners and managers consider larger spatial scales (e.g., watersheds, landscapes, and regions) and longer time frames to ensure that localized and near-term actions do not compromise the capacity of forests to accommodate and adjust to changing conditions.
- ***Adopt agile planning and management approaches that accommodate and address uncertainty.*** Restoring and managing forests in the face of continuous climatic change requires decision-making under uncertainty, underscoring the importance of adaptive planning and management, including the consideration of multiple plausible scenarios of future conditions.
- ***Address climate risks by linking adaptation strategies to key climate-related impacts.*** Understanding climate vulnerabilities and risks to priority forest resources and values serves as the



*In designing restoration projects forest managers will need to determine when and where to manage for persistence—as in this boreal conifer stronghold—and when to facilitate forest transitions (Superior National Forest, Minnesota). Photo: Eli Sagor/UMN.*

basis for developing and implementing adaptation strategies that are capable of reducing risks and sustaining the ecological, social, and economic systems associated with national forests.

- **Manage for change, not just persistence.** As climatic conditions exceed historical ranges of variability, national forest planners will need to consider how to reconcile “desired” future conditions with climatically achievable future conditions. Planners and managers increasingly will need to determine when and where

it may be possible to manage for the persistence of current/historical forest conditions, and when it may be necessary to manage for change by accepting or even facilitating ecological transitions.

- **Optimize, rather than maximize, carbon sequestration opportunities.** National forests will play an increasingly important role in achieving the nation’s climate mitigation goals. Attempting to maximize carbon sequestration and storage, however, can undermine other important ecosystem services and national forest values. Managers should instead seek to optimize sequestration opportunities by balancing carbon goals with other important forest restoration, management, and resilience outcomes.
- **Enhance collaboration to identify shared values, navigate trade-offs, and maximize synergies in the context of changing conditions.** Managing forests for multiple, sustained ecosystem services will necessarily entail trade-offs, particularly given the challenges and uncertainties associated with changing climatic conditions. Engaging local communities and diverse constituencies as early as possible in the forest planning process helps gain buy-in and identify opportunities to minimize trade-offs and maximize synergies, including acknowledgment and discussion of the potential for fundamental changes in national forest conditions.

Restoring our national forests in the face of rapid climate change increasingly will require the deployment of climate-smart approaches to planning and management. This report—and the above principles—are intended to advance development of a shared understanding of climate-smart forest restoration and to promote dialogue among national forest managers and their partners about how to ensure a sustainable and resilient future for America’s national forests.



*Covering 193 million acres across the United States, the National Forest System provides a wide array of ecosystem services, such as clean water, wildlife habitat, and recreational opportunities (White Mountain National Forest, New Hampshire). Photo: Mattia Panciroli/Flickr.*

## 1. INTRODUCTION

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**A**merica's national forests, which represent approximately 20% of all U.S. forest lands, are a treasured resource (Oswalt et al. 2019). From the rain-drenched evergreens of the Pacific Northwest to the brilliant fall displays of New England's deciduous trees, national forests are a core element of our regional identities and economies. Administered by the U.S. Forest Service, the National Forest System includes 175 national forests and grasslands spread across 193 million acres. These lands and waters offer a wide array of ecological services and economic benefits, including provision of timber and fiber, clean water, wildlife habitat, and recreational opportunities. As society grapples with the challenge of a rapidly changing climate, our national forests are increasingly recognized as providing yet another key service: carbon uptake and storage.

Yet, national forests themselves are being profoundly affected by climate change (Millar and Stephenson 2015, Vose et al. 2018, Anderegg et al. 2020). Rising temperatures, increasingly erratic precipitation patterns, and more powerful storms are exacerbating the impacts of existing forest threats and disrupting key ecological processes. Climatic changes are increasing the frequency and/or severity of droughts, wildfires, insect epidemics and other forest disturbances, and climate-induced tree mortality events are becoming widespread, especially in the West (van Mantgem et al. 2013, Millar and Stephenson 2015, Vose et al. 2018, Anderegg et al. 2019). Further, changing climatic conditions can complicate—or even prevent—the recovery and regeneration of forests following such disturbances (e.g., Coop et al. 2020, Rodman et al. 2020, Meyer et al. 2021).

Many of our national forests have been modified by past management and land use, including timber harvest, livestock grazing, long-term fire suppression, energy development, and expansion of housing in nearby wildland areas (e.g., Naficy et al. 2010, Radeloff et al. 2010, Haugo et al. 2015, Battaglia et al. 2018, BLM 2018, D'Amato et al. 2018, Franklin et al. 2018). National forest managers are also contending with ongoing threats from invasive species, disease outbreaks, and other challenges (e.g., Poland et al. 2021). Given the added impacts on forest systems from a rapidly changing climate, an enormous mismatch exists between the current level of forest restoration work and the scale of the challenge (USFS 2012a). The Forest Service estimates that

65–82 million acres in the National Forest System need proactive restoration and management to reduce the risks from these multiple and interconnected stressors (Buford et al. 2015).

To address these growing concerns, there is a need to substantially increase the pace, scale, and quality of restoration on our national forests, an endeavor that will require significant additional funding and capacity (Gandhi et al. 2019). To ensure that the benefits of those investments will endure for generations to come, restoration must account for current and future climatic conditions, taking into consideration both *climate adaptation*, which seeks to reduce climate-

*There is a need to substantially increase the pace, scale, and quality of restoration on our national forests, an endeavor that will require significant additional funding and capacity.*



For much of its history, the Forest Service sought to suppress virtually all wildland fires, a legacy that has contributed to increasingly dense forests and higher fuel loads. Photo: laffertyryan/Flickr.

related impacts on forest systems, as well as *climate mitigation*, which focuses on stabilizing the climate through improved carbon management. National forest restoration and management must also be carried out in ways that support the broad array of environmental benefits and services that Americans now expect from these landscapes—a challenge that necessitates engaging diverse stakeholders that may have divergent and sometimes conflicting values. Thus, in addition to requiring the application of the best available ecological and climate science, restoring America’s forests will depend on strengthening societal understanding, engagement, and support for action. Despite the considerable challenges facing the National Forest System, we can help ensure a sustainable and resilient future for our national forests by embracing ecologically appropriate and climate-smart approaches to restoration.

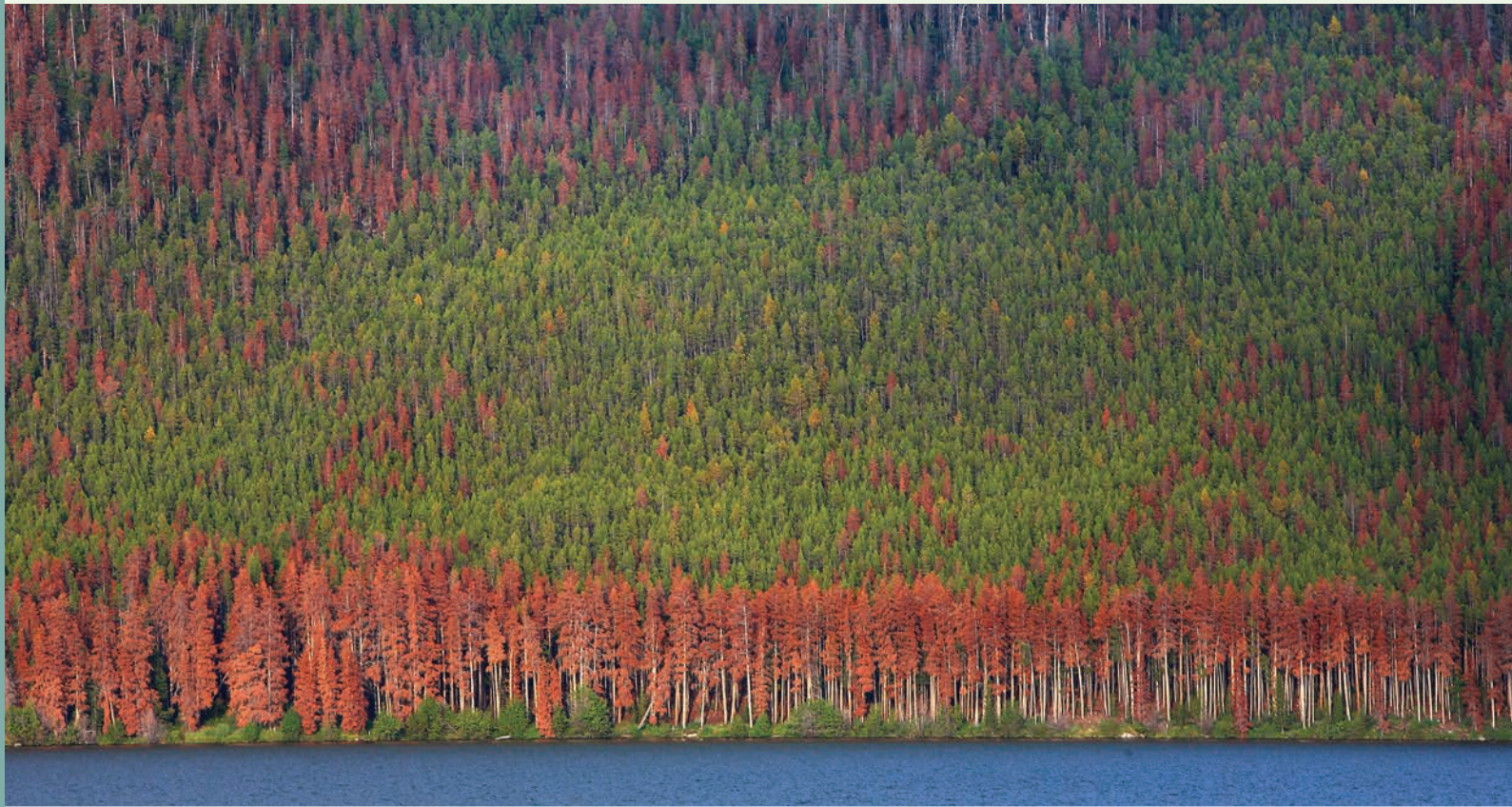
## 1.1. PURPOSE OF THIS REVIEW

Key to realizing this vision is the development of a collective, science-based understanding of what it means to do forest “restoration” in an era of accelerating climate change. This review and

synthesis of the science underlying climate-smart forest restoration is intended to help advance such a shared understanding. In particular, it is designed to help organizations and individuals who are engaged in national forest–related conservation and restoration understand the implications of rapid climatic changes for achieving desired forest restoration outcomes and enhance their ability to productively collaborate with the Forest Service in achieving common goals.

Indeed, the Forest Service has made considerable progress over the past decade in understanding the impacts of climate change on forest ecosystems and identifying approaches for better incorporating climate adaptation and mitigation principles into its work. To scale up the application of these practices, federal land managers need committed, informed, and climate-savvy partners in the state, tribal, nonprofit, and private sectors. Drawing on a growing body of evidence, research, and experimentation, this review looks at how rapid and accelerating climate change is inspiring an important evolution in policies and approaches for national forest restoration and management. It is intended to serve as a foundation for continued dialogue and collaboration to enhance the application of such approaches in practice.





*Warming winter temperatures have contributed to increases in mountain pine beetle outbreaks and associated tree mortality across much of the Rocky Mountains (Arapaho National Recreation Area, Colorado). Photo: A.J. Schroetlin/Flickr.*

## 2. AMERICA'S FORESTS FACE A CHANGING FUTURE

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**A**s Gifford Pinchot, the first chief of the Forest Service, wrote, “Next to the earth itself the forest is the most useful servant of man. Not only does it sustain and regulate the streams, moderate the winds, and beautify the land, but it also supplies wood, the most widely used of all materials ... The object of practical forestry is precisely to make the forest render its best service to man in such a way as to increase rather than to diminish its usefulness in the future” (Pinchot 1905). Nevertheless, for much of the 20th century, timber harvest was the primary emphasis of national forest management (D’Amato et al. 2018, Franklin et al. 2018), and activities such as roadbuilding, logging, and fire suppression resulted in considerable alteration of natural forest conditions.

By the middle of the 20th century, concerns about the continued sustainability of timber resources, along with growing public interest in ecological and noneconomic social values, led to enactment of the Multiple-Use Sustained-Yield Act, which explicitly called for national forests and grasslands to be managed for a range of purposes, including “outdoor recreation, range, timber, watershed, wildlife and fish” (MUSYA 1960). Subsequent laws and policies, including the National Environmental Policy Act, the Endangered Species Act, and especially the National Forest Management Act of 1976, have further guided the agency’s planning and management toward the broader goal of sustaining ecosystem health and diversity. By the turn of the 21st century, the concepts of ecosystem management and ecological restoration had gained prominence, with a

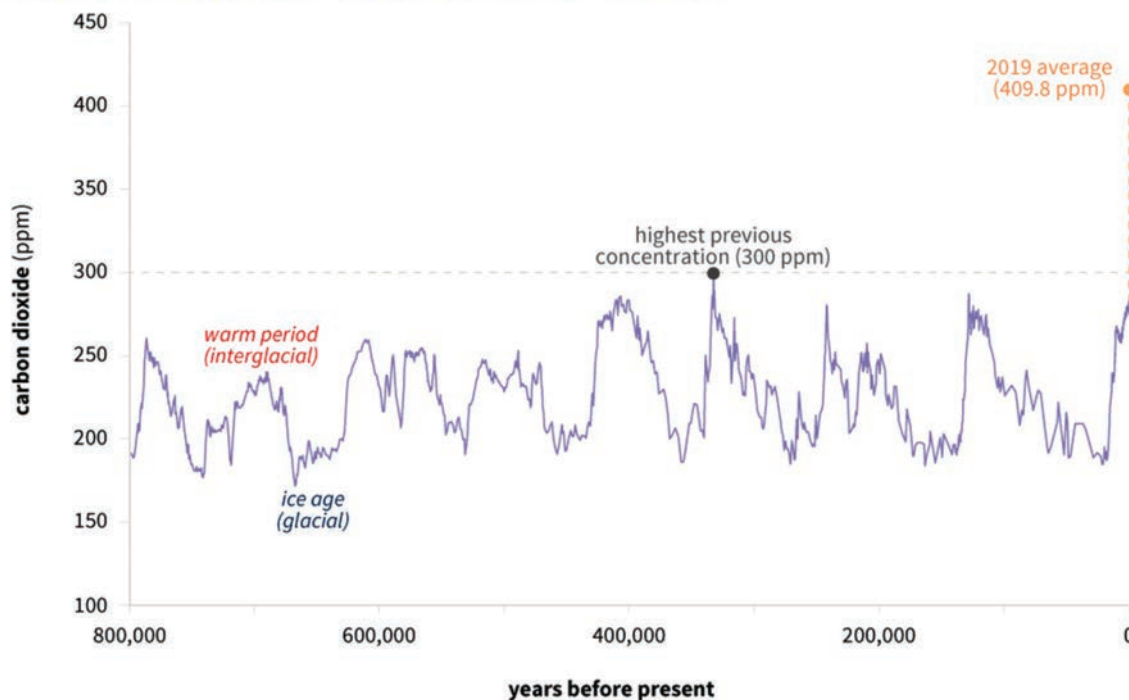
view toward achieving a mosaic of forest conditions on the landscape, often benchmarked by estimates of the historical (or “natural”) range of variability (Millar 2014). This shift reflected a recognition of the key role of natural processes like succession and disturbance in supporting forest health and biodiversity goals.

Yet, on top of pervasive threats from invasive species, land use change, and other major stressors, the growing risks from a rapidly changing climate have compelled updated forest planning and management approaches across all of the nation’s forest systems, including those managed by the Forest Service (USFS 2011a, Vose et al. 2018). Indeed, the impacts and risks from climate change were explicitly addressed in the 2012 Forest Service Planning Rule (2012 Planning Rule), which reflects the first substantive update to National Forest System planning in 30 years (USFS 2012b).

## 2.1. UNDERSTANDING CLIMATE CHANGE AND ITS IMPACTS ON FOREST ECOSYSTEMS

The relationship between atmospheric carbon dioxide (CO<sub>2</sub>) levels and warming of the Earth through the “greenhouse effect” has been well established scientifically since the late 1800s, based on fundamental principles of physics and chemistry (Anderson et al. 2016). Human activities, especially the burning of fossil fuels, have increased atmospheric concentrations of CO<sub>2</sub> from about 280 parts per million (ppm) in the mid-1800s to about 410 ppm today—far higher than at any time in at least the last 800,000 years (Lüthi et al. 2008, Kopp et al. 2017) (see Figure 1). Higher atmospheric CO<sub>2</sub> has already contributed to a notable increase in the observed temperature in the

### CARBON DIOXIDE OVER 800,000 YEARS



**Figure 1. Atmospheric carbon dioxide.** Atmospheric carbon dioxide (CO<sub>2</sub>) concentrations (in parts per million [ppm]) are now higher than at any point in the past 800,000 years. On a geologic timescale, the increase since 1850 (shown in orange) looks virtually instantaneous. Source: NOAA (2020).

United States. Nationally, average air temperatures are now more than 1.8°F warmer than the historical record, with Alaska warming at about twice the rate of the lower 48 states (Vose et al. 2017).

Further demonstrating this warming trend, the past decade (2011–2020) was the warmest on record, and the world’s seven hottest years have occurred since 2014 (NOAA 2021). These temperature increases are evidence of rapid climate change, and without substantial reductions in atmospheric CO<sub>2</sub> concentrations, average temperatures, and the frequency of multiday heat waves, will continue to rise. The annual average temperature over the contiguous United States is projected to increase by about 2.5°F by mid-century relative to 1976–2005, which suggests that recent record-setting temperatures may become “common” within a few decades (Vose et al. 2017). Rising temperatures, of course, are just one aspect of our changing climate. We are also experiencing altered precipitation patterns across the country, including shifts in types, seasonality, intensity, and variability (Easterling et al. 2017). Of particular concern is the increasing frequency and severity of extreme weather events, including hurricanes, heavy rainstorms, and droughts.

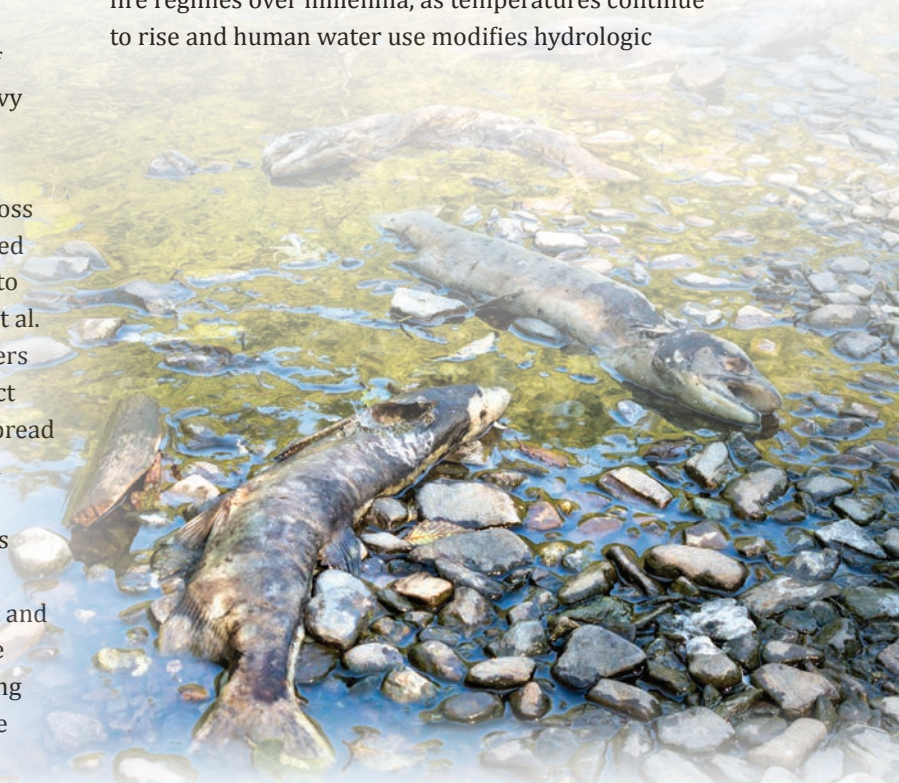
Over the past several decades, forest managers across the country have observed escalating climate-related impacts, and scientists project significant changes to continue in the years and decades to come (Joyce et al. 2014, Vose et al. 2018). For example, warmer winters and extended growing seasons are promoting insect outbreaks and facilitating the establishment and spread of invasive species. Temperature increases coupled with drought are associated with larger wildfires and a longer fire season. Warmer, wetter conditions favor more extreme flooding and can promote the establishment of certain forest pathogens. As plant and animal species across the country experience these and other changing conditions, many are responding through shifts in phenology (timing of key life-cycle events like leaf emergence or flowering), changes in growth, reproduction, and mortality rates, and altered abundance patterns and distributional ranges.

These species-level responses are also leading to ecosystem-level shifts, including changes in ecosystem composition, structure, and function.

The following overview represents just a sampling of the extensive and growing body of scientific literature on the known and potential future impacts of climate change on North America’s forests, including those within the National Forest System.

### 2.1.1. More Severe Drought and Heat Waves

Increasing temperatures and altered precipitation patterns are contributing to changes in drought conditions across the United States, and scientists project that the frequency and severity of drought in many areas will increase as the climate continues to change (Strzepek et al. 2010, Chiang et al. 2018, Vose et al. 2018). Although droughts have been an important factor shaping forest characteristics and fire regimes over millennia, as temperatures continue to rise and human water use modifies hydrologic



*Low water levels and warmer temperatures are putting additional stress on salmon in the Pacific Northwest and Alaska (Chinook salmon, Ketchikan Creek, Alaska). Photo: VC Images/Alamy.*

systems, scientists are seeing shifts in key drought characteristics, such as faster onset (flash drought), higher drought intensities, and impacts in new locations (Crausbay et al. 2017, 2020; Vose et al. 2018).

The specific causes, types, and impacts of drought on forest ecosystems vary considerably, both between and within ecological system types (Hanson and Weltzin 2000, Klos et al. 2009, Seager et al. 2009, Anderegg et al. 2013, Vose et al. 2016). A key challenge in anticipating the location and intensity of drought impacts now and in the future is that climate change and human water use are influencing many different drought drivers at the same time (e.g., precipitation, temperature, storage of water as snow, and hydrologic influences of vegetation and water withdrawals). In addition, factors like soil types, topography, and certain management practices (e.g., thinning of forest stands) can also influence the water available for ecosystems (Van Loon et al. 2016; Crausbay et al. 2017, 2020; Cartwright 2018).

In the western United States, many forest systems are adapted to annual, seasonal droughts. However, warmer conditions have contributed to an increase in the duration and severity of droughts in parts of the region over the last two decades, affecting both terrestrial and aquatic systems (e.g., van Mantgem et al. 2009, Hicke et al. 2016, Mote et al. 2016, Young et al. 2017, VerWey et al. 2018, Elias et al. 2021). For example, extreme drought associated with low snowpack has significantly altered hydrologic conditions across much of the West, contributing to higher winter streamflow, earlier and reduced peak spring flow, reduced summer flow, and, in combination with higher air temperatures, increased stream temperatures (Mote et al. 2018, Peterson et al. 2019). In addition to affecting fish and other aquatic species, shifts in river hydrographs due to reduced snowpack can have a significant impact on riparian areas and the diverse plants and animals they support

(Dwire et al. 2018, Elias et al. 2021). In the coming decades, average annual snowpack across much of the West is projected to decline considerably, with broad implications for forest ecosystems and the services they provide, including provision of water resources to human communities (Fyfe et al. 2017, Gergel et al. 2017).

In the Midwest and East, where drought historically has been more sporadic and less intense than in the West, there is limited current scientific understanding of how forest systems will respond to what are projected to be uncharacteristically severe drought conditions in the future (Hanson and Weltzin 2000, Brooks 2009, Asbjornsen et al. 2019, Druckenbrod et al. 2019). Given that some tree species are known to be more tolerant of drought conditions than others, however, research suggests that more prolonged and frequent drought resulting from a changing climate could lead to a change in forest composition and structure in some areas (Klos et al. 2009, Vose et al. 2016). Further, many eastern forests have embedded wetlands or support species that are highly adapted to waterlogged soils. In those systems, changes in moisture regimes are likely to drive loss of species and contribute to forest type transitions, even without water becoming limiting for plant growth (Brooks 2009, Vose et al. 2018).

An increase in the occurrence of compound extreme events (e.g., extreme heat and drought) has the potential for even greater impacts on terrestrial and aquatic forest systems than individual events (Kopp et al. 2017). Over the past 50 years, there has been a substantial increase in concurrence between heat waves and low precipitation across much of the country, leading to hotter droughts and increases in evaporative demand (i.e., low moisture levels in the air pulls moisture from soils and vegetation) (Mazdiyasn and AghaKouchak 2015). Such combined extreme events can have significant social and ecological

*Average annual snowpack across much of the West is projected to decline considerably, with broad implications for forest ecosystems and the services they provide, including provision of water resources to human communities.*



*Long-term drought is affecting even large, mature trees. In 2020, up to 10% of the world's giant sequoias died from the combined stress of drought and an exceptionally severe wildfire (Sequoia National Forest, California). Photo: Nathan Stephenson/USGS.*

implications (Kopp et al. 2017). For example, the 2011–2016 drought in California, which was characterized by both low precipitation and high temperatures, killed more than 125 million trees (AghaKouchak et al. 2015, Diffenbaugh et al. 2015, USFS 2019). Further, a combination of extreme heat and drought has had a significant impact on salmon and other cold-water fish across much of the West, contributing to extensive die-offs in some areas (e.g., Crozier et al. 2020, Westley et al. 2020, Bowerman et al. 2021). As temperatures continue to rise, the impacts of drought are likely to worsen (Herrera-Estrada and Sheffield 2017). For instance, scientists project a significant increase in the risk of drought-induced stress and tree mortality in many areas over the coming decades, which is likely to significantly affect forest structure, composition, and function, potentially driving transformation of existing forested systems to other ecological types (Anderegg et al. 2015, Batllori et al. 2020).

### 2.1.2. Worsening Insect and Pathogen Outbreaks

How climate change will affect forest insect and pathogen outbreaks in the future is uncertain, in part due to variability in species-specific responses to climate change (Pureswaran et al. 2018). However, recent climatic changes have been exacerbating outbreaks of both native and introduced insects and pathogens across the country (Dukes et al. 2009, Weed et al. 2013, Hatfield et al. 2015, Berner et al. 2017, Loehman et al. 2017, Swanston et al. 2018). For example, higher temperatures are accelerating the development of some insects, such as the hemlock woolly adelgid, whose eggs have been hatching up to 3 months ahead of their normal emergence in parts of the Southeast (Leppanen and Simberloff 2017). Warmer winters have also enabled the establishment of cold-limited insects and diseases, such as southern pine

beetle (Trân et al. 2007), spruce budworm (Régnière et al. 2012), and beech bark disease (Stephanson and Ribarik Coe 2017).

Longer growing seasons can also mean more generations of insects in a season, increasing the duration and magnitude of insect damage (DeLucia et al. 2012). Mountain pine beetle in the western United States, for instance, has become a dominant disturbance agent as changing climatic conditions have altered the life cycle of some broods to two generations per year (Mitton and Ferrenberg 2012). Recent pine beetle outbreaks have resulted in tree die-offs across more than 240,000 acres per year (Breshears et al. 2005, Klapwijk et al. 2012, Hicke et al. 2016). Climate change may also affect the severity of outbreaks via changes in host species, such as through shifts in host distribution and a decline in tree health (Dukes et al. 2009, Weed et al. 2013). For example, drought stress has been found to make pinyon pines more susceptible to outbreaks of the pinyon ips bark beetle (Kleinman et al. 2012). Together, these stressors have contributed to significant declines in certain types of pinyon–juniper woodlands in the Southwest (Friggens et al. 2020). On the whole, insect outbreaks have killed hundreds of millions of trees across the United States over the past 20 years (Hicke et al. 2016). As noted by Vose et al. (2018), this appears to be far outside of the historical context, and it is likely associated with a changing climate.

Outbreaks of forest diseases are also projected to increase under warmer and wetter conditions in some areas (Sturrock et al. 2011, Weed et al. 2013). For example, epidemics of certain foliar, vascular, and root pathogens, such as Swiss needle cast and sudden oak death, have been linked to higher temperatures and increasing precipitation (Weed et al. 2013). In the Coast Range of the Pacific Northwest, Swiss needle cast has been a persistent disease of Douglas-fir since the discovery of the fungal pathogen in the mid-1990s. Research suggests that the disease will increase at higher elevations and higher latitudes under a scenario of warmer winters and wetter summers (Lee et al. 2017). Higher temperatures and an increase in

spring precipitation are also projected to increase the prevalence and range of the pathogen responsible for sudden oak death across California, Oregon, and Washington (Kliejunas 2011).

### 2.1.3. Larger Wildfires and a Longer Fire Season

Climate change is increasing wildfire potential across much of the United States (e.g., Cattau et al. 2020, Goss et al. 2020, Parks and Abatzoglou 2020, Higuera and Abatzoglou 2021). In the West, for instance, higher temperatures, earlier snowmelt, and reduced summer precipitation have contributed to a longer wildfire season and an increase in the total area burned, including the size of patches burned by high-severity fire (McKenzie et al. 2004, Running 2006, Westerling et al. 2006, Miller and Safford 2012, Hicke et al. 2016, Westerling 2016, Singleton et al. 2019). Extreme drought has exacerbated fire conditions across much of the region, and the occurrence of so-



*Wildfire season has been lengthening across the western United States, and forest managers are documenting an increase in the acreage burned by high-severity fires (2016 Roaring Lion Fire, Bitterroot National Forest, Montana). Photo: USDA.*

called “megafires”—those with areal extent greater than 100,000 acres—has risen significantly (Davis and Weber 2018). In addition to posing risks to human communities, these fires can emit considerable amounts of carbon into the atmosphere (e.g., Sommers et al. 2014, Urbanski 2014), and they have the potential to permanently transform forest ecosystems that are not adapted to high-severity fire regimes (Hicke et al. 2016, Jones et al. 2016, Berner et al. 2017, Heyck-Williams et al. 2017, Haffey et al. 2018, Halofsky et al. 2020).

In interior Alaska, record-breaking heat and drought in recent years have led to a significant increase in both the severity of wildfires and the total area burned. This has altered dynamics between vegetation and permafrost and is contributing to a shift in boreal forest composition from spruce-dominated to deciduous-dominated systems (Johnstone et al. 2010, Wolken et al. 2011, Sanford et al. 2015, Foster et al. 2019). Within the next few decades, climate change is projected to lead to functional and structural changes in Alaska’s boreal forest that are unprecedented in the last 6,000 years (Chapin et al. 2010). Indeed, a number of studies suggest that the western United States will experience even larger, more frequent wildfires and a longer fire season as the climate continues to change. For example, scientists project that by mid-century total annual area burned across the West could increase by 2-6 times from the present (Litschert et al. 2012, Ojima et al. 2014, Vose et al. 2018).

Drought has also exacerbated wildfires in the East. In the Southern Appalachian region, for example, an exceptional drought in 2016 contributed to an above-normal fall wildfire season, resulting in loss of life and property (Vose et al. 2018, James et al. 2020). As in the West, many southeastern forest systems are adapted to

*Scientists project that by mid-century total annual area burned across the West could increase by 2-6 times from the present.*

specific fire regimes. Longleaf pine forests, for instance, rely on relatively frequent, low-intensity fires. Changing climatic conditions are projected to increase both the intensity and severity of wildfires across much of the region (Liu et al. 2013, Terando et al. 2016). This could lead to tree mortality as well as limit the window of opportunity for management options such as prescribed burns (Mitchell et al. 2014, Kupfer et al. 2020).

### **2.1.4. Heavier Rainstorms and More Extreme Flooding**

Floods are another important ecological disturbance factor that promote a mosaic of forest conditions on the landscape and help sustain biodiversity associated with all stages of forest succession (Hayes et al. 2018). However, an increase in the frequency and intensity of heavy rainfall events, changes in the timing and extent of snowmelt, and other climate-related factors are exacerbating flooding in forests across the country, affecting aquatic and terrestrial systems as well as roads and other built infrastructure. In the Pacific Northwest, for example, intense and persistent rainfall and associated flooding have led to erosion and landslides in the Mt. Baker-Snoqualmie National Forest in Washington, destroying roads and disrupting access (Raymond et al. 2014). A combination of higher average temperatures and increased precipitation in fall, winter, and spring are projected to contribute to increased flooding for most of the region’s river basins toward the latter half of this century, particularly in areas where precipitation has been historically dominated by rain or mixed rain and snow (Tohver et al. 2014).

In the eastern United States, the frequency of extreme precipitation events has increased since the mid-1990s, and further increases are projected to contribute to higher total runoff and peak streamflow across much of the region (Easterling et al. 2017, Swanston et al. 2018). In addition to posing risks to downstream communities, an increase in the magnitude and frequency of flooding associated with such events could have significant implications for the composition and function of

ecosystems not adapted to the altered hydrologic regime. For example, studies suggest that an increase in the depth and duration of early-season flooding in some areas could reduce the success of seedling regeneration within bottomland hardwood forests (e.g., McCurry et al. 2010, Kabrick et al. 2012).

Extreme rainfall associated with tropical storms and hurricanes (tropical cyclones) has also become more common over the past few decades, and recent storms have contributed to significant flooding and oversaturation of soil in areas of the Atlantic and Gulf coasts and U.S. Caribbean (e.g., Risser and Wehner 2017, Patricola and Wehner 2018, Keellings and Alaya 2019, Case et al. 2021a). Although hurricane-strength winds can cause significant damage to forests, research suggests that climate-induced increases in the duration and intensity of rainfall associated with tropical cyclones are also of concern (Van Beusekom et al. 2018, Keellings and Alaya 2019, Gang et al. 2020). In Puerto Rico, for instance, heavy rain during Hurricanes María and Irma in 2017 was found to have caused greater damage to the island's tropical forests than did wind (Hall et al. 2020).

### 2.1.5. Altered Phenology and Growth Rates

Studies suggest that changes in temperatures and precipitation patterns and a continued rise in atmospheric CO<sub>2</sub> will affect the growing environment of many tree species throughout the United States, altering processes such as tree phenology, water use, primary productivity, recruitment, and mortality (Vose et al. 2012, Fisichelli et al. 2014, Harvey et al. 2020). Responses among species will vary both spatially and temporally and will depend on factors such as age class, competition, and other biotic and abiotic conditions (Restaino et al. 2016, Ford et al. 2017, Dixit et al. 2020).

One of the primary ways that climate change affects forests is through changes in the timing of life-cycle events (phenology) of trees and other species. Indeed, a trend toward earlier leaf emergence in the spring and shifts in leaf senescence among a number of plant species in fall has been an important indicator of the effects of recent climatic changes (Gill et al. 2015, Panchen et al. 2015, Monahan et al. 2016). For example, Melaas et al. (2018) found that leaf emergence has trended earlier in the eastern temperate forests of



More extreme rainfall events and flooding are causing damage to roads and other infrastructure on national forests (Umatilla National Forest, Oregon). Photo: USFS.

North America over the 30-year period between 1984 and 2013. Research suggests that earlier spring warm-up, later fall cooldown, and an associated increase in the length of the growing season in the Northeast in the coming decades are likely to affect productivity of a number of tree species (Rustad et al. 2012). These changes, in turn, can affect nutrient proficiency, photosynthesis, and other physiological processes, as well as ecological processes such as water fluxes and carbon uptake (see section 2.1.1.7, below) (Richardson et al. 2013, Keenan et al. 2014, Estiarte and Peñuelas 2015, Way and Montgomery 2015, Fang et al. 2020).

A trend toward earlier spring budburst is also occurring among some tree species in the West, which is attributed in part to higher average temperatures (Ault et al. 2011). As the climate continues to warm, scientists expect further changes in the timing of budburst, although differences will vary by species and location of a population within its range (Harrington and Gould 2015). For example, some tree species, such as Douglas-fir, require a period of chilling before warmer spring temperatures force budburst, although the required chilling period varies across different genotypes. Projections for temperature-induced changes in budburst among Douglas-fir across parts of its range in western North America suggest that a 3.2–5.5°C increase in average winter temperatures by 2080 could result in significantly earlier budburst for the trees across much of the study area, but later budburst in some southern portions due to a lack of chilling (Harrington and Gould 2015). Delayed budburst could lead to poor growth, particularly if climate change also leads to earlier drought in those areas.

In addition to contributing to a warmer climate, a higher concentration of atmospheric CO<sub>2</sub> can have a direct effect on plant growth. Experimental studies have shown that trees grow faster when exposed to enriched CO<sub>2</sub> levels (e.g., Norby et al. 2005, Kim et al. 2020).

Evidence suggests that higher levels of CO<sub>2</sub> have also contributed to an increase in forest water use efficiency in some areas (Keenan et al. 2013, Fernández-de-Uña et al. 2016). In addition, the combination of a reduction in acidic deposition, warmer springs, and an increase in atmospheric CO<sub>2</sub> was found to contribute to recent increases in the growth of red spruce trees in the Central Appalachian Mountains (Mathias and Thomas 2018). However, the net effect of CO<sub>2</sub> fertilization on forest ecosystems remains uncertain, as potential benefits may be offset by water stress and other climate-related impacts (e.g., Morin et al. 2018, Belmecheri et al. 2021).

### 2.1.6. Shifts in Species' Ranges

A number of trees and other forest plant and animal species (both native and nonnative) are shifting their distributions in response to changing climatic conditions (e.g., Mamet et al. 2019; Knott et al. 2019, 2020; Terskaia et al. 2020). In some instances, species have moved to higher elevations or latitudes, largely in response to changing temperatures; in others, species have shifted longitudinally or to lower elevations, tracking shifts in precipitation patterns and moisture availability (e.g., Crimmins et al. 2011, Bell et al. 2014, Lenoir and Svenning 2015, Fei et al. 2017). Taking a variety of factors into consideration, model-based projections suggest that further range shifts among tree species are likely in forests across North America (e.g., Iverson et al. 2008, 2019; Woodall et al. 2009; Iverson and McKenzie 2013; Clark et al. 2014; Matthews et al. 2014; Prasad et al. 2020; Rehfeldt et al. 2020; Toot et al. 2020).

Importantly, range shifts may result from a variety of site- and species-specific processes, and the potential for and rate of such shifts depend on factors such as reproduction, dispersal, and establishment. For example, a geographic shift in species may occur

*Given the considerable variability in the rates and degree to which tree range shifts will occur over time, the composition of many forest systems in the future will differ from both historical and current conditions.*

because of increased recruitment at the leading edge of a species' range, reproductive failure or die-off at the trailing edge, or a combination of the two (Lenoir and Svenning 2015, Boisvert-Marsh et al. 2019). However, the ability of trees to expand into newly suitable habitat in the future will depend on whether they can keep pace with the increasing velocity of climate change (Zhu et al. 2012, Sittaro et al. 2017, Liang et al. 2018, Solarik et al. 2018). Further, successful establishment in new areas depends not just on a favorable climate, but also on whether other, site-specific biotic and abiotic conditions (e.g., soil type) will support recruitment and survival (Etterson et al. 2020, MacKenzie and Mahony 2021). Given the considerable variability in the rates and degree to which tree range shifts will occur over time, and the high rates of land conversion in many parts of the United States that limit tree species dispersal, the composition of many forest systems in the future will differ from both historical and current conditions (Fei et al. 2017).

Another potential concern for national forest management is the potential for climate change to contribute to shifts in the range of invasive species (i.e., introduced, nonnative plants and other organisms that do or can cause environmental or economic harm to humans) (Iannone et al. 2020). As noted in the discussion on insect and pathogen outbreaks (which may include both native and nonnative species), studies suggest that climate change may enable invasive species with established populations to move into new areas as constraints such as temperature thresholds are removed (Hellmann et al. 2008, Poland et al. 2021). Further, climate-related changes in disturbances, such as uncharacteristically severe or frequent fires, drought, or increased flooding and erosion, may enable invasive species to gain a foothold in disturbed areas (Hellmann et al. 2008, Poland et al. 2021). Although climate may become less suitable for invasive species in some areas (e.g., Bradley 2009), some of the very characteristics that make some species “invasive,” such as broad habitat requirements, high dispersal ability, and short time to maturity, are likely to put many invasive species at a competitive advantage (Hellmann et al. 2008, Duker et al. 2009).

### 2.1.7. Changes in Ecosystem Composition, Structure, and Function

Both direct and indirect effects of climate change can alter the composition, structure, and function of forest ecosystems (Vose et al. 2018). For example, changes in tree species composition, which can occur rapidly following disturbances or more gradually through changes in recruitment or survival, can lead to disruption of interspecific relationships, such as competition, predator–prey interactions, and pollinator–plant associations. In some forests of the Northeast, for instance, wildflower leaf-out and bloom times in the understory are strongly limited by light availability. Higher temperatures have contributed to earlier leaf-out among the overstory trees compared



*Freshwater swamp forests, such as these bald cypresses in Louisiana, are suffering widespread die-offs as sea-level rise increases salinity in estuaries, coastal wetlands, and tidal rivers. Photo: Southern Forests/Flickr.*

*The impacts of climate change could lead some forest systems to reach critical ecological thresholds, where conversion to different forest types, or even to non-forested systems, is possible.*

to the understory vegetation, which has resulted in a shorter period of light availability for understory herbs (Heberling et al. 2019). If this trend continues, the region could see a decline in wildflower species abundances and a shift in community composition, which could alter forest energy flow and nutrient cycling (Gilliam 2007), as well as affect pollinator visitation (Cho et al. 2017).

In addition, shorter and warmer winters are creating significant changes in winter belowground processes, which can have a wide range of cascading impacts. In the Northern Forest region, for example, a decline in the duration and depth of insulating snowpack has, paradoxically, contributed to freezing soils, which can lead to reductions in soil microbial biomass and affect the ability of trees to retain nutrients and carbon (Sorensen et al. 2018, Contosta et al. 2019, Harrison et al. 2020, Wilson et al. 2020). And in parts of coastal Alaska, milder winter conditions and an associated decline in snowpack has contributed to significant mortality among yellow-cedar due to freeze damage to fine roots, particularly in areas with poorly draining soils (Hennon et al. 2016).

Lastly, the impacts of climate change could lead some forest systems to reach critical ecological thresholds, where conversion to different forest types (e.g., the conversion of black spruce to deciduous forest in interior Alaska following severe wildfire [Hansen et al. 2020]), or even to non-forested systems, is possible (Koch et al. 2009, Adams 2013, Anderson-Teixeira et al. 2013, Allen et al. 2015, Teskey et al. 2015, Walker et al. 2018, Busby et al. 2020, Harris and Taylor 2020). Research by Parks et al. (2019a), for instance, suggests that a significant percentage of

forest systems at the trailing edge of their distribution (i.e., those areas expected to experience a range contraction under changing climatic conditions) in the intermountain western United States are at risk of fire-facilitated conversion to non-forest conditions by mid-21st century. And along the Atlantic and Gulf coasts, accelerating sea-level rise has contributed to the conversion of coastal forests to intertidal vegetation, affecting both the composition and function of coastal systems (Kirwan and Geden 2019). For example, Ury et al. (2021) calculated that more than 21,000 acres of trees on North Carolina's Alligator River National Wildlife Refuge were converted to "ghost forest" (i.e., areas characterized by standing dead trees and fallen tree trunks) between 1985 and 2019 due to rising sea levels and increasing salinity.

## 2.2. ADDRESSING CLIMATE-RELATED UNCERTAINTY

Despite tremendous progress in scientific understanding of known and potential climate-related impacts on forest ecosystems, it is important to recognize that key areas of uncertainty remain regarding how the climate may change in the future, how species and ecological systems may respond to those changes, and how people may react to those climatic and ecological shifts (Park et al. 2014). Although some of these uncertainties can be reduced over time—for instance, through the development of new research and modeling techniques and ongoing monitoring—other forms of uncertainty may never be resolved. National forest managers, however, have always confronted uncertainty in their work, and managing in the face of such uncertainty is not

*Key areas of uncertainty remain regarding how the climate may change in the future, how species and ecological systems may respond to those changes, and how people may react to those climatic and ecological shifts.*



*Adaptive management has long been recognized as essential to effective forest restoration, as has the importance of promoting landscape heterogeneity. The uncertainties of future climatic conditions and ecological responses make these approaches even more imperative (Willamette National Forest, Oregon). Photo: Matthew Tharp/USFS.*

new. Thus, existing experience and approaches for managing dynamic and unpredictable systems can be brought to bear on the new challenges of climate change-related uncertainty. For example, adopting a more agile, learning-based approach to management (i.e., *adaptive management*) has been recognized by the Forest Service as an important principle for addressing climate change (Millar et al. 2007; Joyce et al. 2008, 2009). The 2012 Planning Rule (USFS 2012b) provides for an “adaptive planning cycle” that allows the Forest Service to “adapt to changing conditions and improve management based on new information.” In addition, the Forest Service has embraced tools such as scenario planning to identify robust management strategies (i.e., those that are likely to be effective

across a range of potential future conditions) and facilitate long-term decision-making under uncertain future conditions (e.g., Nydick and Sydoriak 2011, Joyce and Coulson 2020).

Nevertheless, concerns about the inherent uncertainties and complexity associated with climate change remain a barrier to implementation of climate adaptation among some forest practitioners (Halofsky et al. 2018). Enhancing engagement between scientists and practitioners through workshops, trainings, collaborative working groups, and other means will be essential to address the considerable additional challenges that the complexities and uncertainties of a changing climate bring to national forest management.



*Over the past two decades, Forest Service spending on wildfire suppression grew from under 20% of total budget to more than 50%, eroding the agency's ability to fund needed forest restoration and management activities. Photo: Kari Greer/USFS.*

### 3. IMPLICATIONS OF CLIMATE CHANGE FOR NATIONAL FOREST MANAGEMENT

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Climate change will challenge the ability of the Forest Service to achieve its mission of sustaining the health, diversity, and productivity of our national forests and grasslands to meet the needs of present and future generations. Indeed, the impacts of climate change have already had significant implications for Forest Service activities. For instance, the increasing cost of fighting worsening wildfires has stressed the agency's budget and contributed to the massive backlog in forest restoration (USFS 2017, Dumroese et al. 2019, Nave et al. 2019). Over the last two decades, Forest Service spending on wildfire suppression grew from under 20% of total budget to more than 50%, eroding the agency's ability to fund restoration and management activities to improve forest health and resilience, despite the growing need for such activities (USFS 2015b).

Fortunately, recent policies provide an opportunity to accelerate ecologically appropriate and climate-smart forest restoration. New authorities included in the 2018 Consolidated Appropriations Act, for instance, creates a disaster fund separate from the agency's operating budget that can be tapped for extraordinary wildfire suppression activities. This has the potential to free up much needed funding for restoration and reforestation activities and enable the Forest Service to expand work on collaborative, landscape-scale approaches to fire management (Taylor 2019). Recent Forest Service directives have also encouraged forest planners and managers to place greater emphasis on fostering forest resilience under current and future climatic conditions, rather than requiring strict adherence to historical ranges of variation (USFS 2015a, Bone et al. 2016).

Addressing the growing threats from climate change on the nation's forests requires a two-pronged approach. *Climate adaptation* refers to efforts designed to prepare for and address the impacts of climate change, with an emphasis on reducing key climate-related vulnerabilities and risks (Stein et al. 2013). *Climate mitigation*, in contrast, focuses on efforts to reduce the atmospheric concentrations of carbon dioxide and other greenhouse gases that are the underlying drivers of rapid climate change (IPCC 2014). Over the past two decades there has been considerable progress in developing the science and practice of both climate adaptation and mitigation in the United States and globally, although there is an urgent need for much broader adoption and deployment of these approaches. The Forest Service has been an early adopter and leader among the nation's resource management agencies in recognizing the importance of integrating consideration of climate change into its planning and management activities, including both adaptation and mitigation (e.g., USFS 2008, 2011a, 2011b, 2012b, 2015a, 2016, 2018).

However, recent studies suggest that broad-scale progress on climate-related activities at the Forest Service has often been hampered by several factors, including: 1) uncertainties about future climatic conditions, the potential ecological response to those changes, and appropriate management and adaptation options (e.g., Archie et al. 2014; Janowiak et al. 2017, 2020; Halofsky et al. 2018; Scheller and Parajuli 2018); 2) inconsistencies in agency guidance and adaptation-related terminology (e.g., Kemp et al. 2015, Bone et al. 2016, Timberlake and Schultz 2017, Greiner et al. 2020); and 3) barriers to planning and implementation of climate adaptation and mitigation strategies, including limited staff and funding capacity, jurisdictional complexities, and political and economic concerns (e.g., Jantarasami et al. 2010, Kemp et al. 2015, Timberlake and Schultz 2017, Halofsky et al. 2018, Gandhi et al. 2019). Progress has also been hampered by continuing challenges in reconciling science and information needs among researchers, managers, and other stakeholders (Kemp et al. 2015, Rodriguez-Franco and Haan 2015, Timberlake and Schultz 2017).

This science review and synthesis is intended to help inform efforts to address many of these challenges. The information presented here draws from and builds on the extensive research, expertise, and experience of the Forest Service and its partners, as well as researchers from the academic and nonprofit sectors. We fully recognize that there is no one-size-fits-all approach to climate-smart forest restoration and stewardship given the wide range of forest types, natural disturbance regimes, and current forest and watershed conditions; inconsistencies in scientific assessment and management approaches; and the unique needs and concerns of local communities, industry, and other stakeholders. Nevertheless, by advancing a shared understanding of key science-based concepts and practices, we hope to provide an informational foundation for promoting multi-scale collaborations to further climate-smart restoration on our national forests.



*Widespread tree mortality due to climate change-enhanced drought, insect outbreaks, and other disturbances poses an enormous challenge for managing national forests across much of the West (Sierra National Forest, California). Photo: USFS.*

### 3.1. U.S. FOREST SERVICE RESPONSE TO CLIMATE CHANGE

As described in its *National Roadmap for Responding to Climate Change* (USFS 2011a), the Forest Service has adopted a three-pronged strategy to confront climate change, including: 1) assessment of current risks, vulnerabilities, policies, and gaps in knowledge; 2) engagement of internal and external partners in seeking solutions; and 3) management for resilience, in ecosystems as well as in human communities, through adaptation, mitigation, and sustainable consumption strategies. These activities are not mutually exclusive—they are intended to be both dynamic and synergistic. The development of a Climate Change Performance Scorecard (USFS 2011b) enabled national forest managers to begin tracking progress in each of these areas and helped the agency increase its awareness and capacity to respond to climate change. This tracking system has more recently been replaced by a Sustainability Scorecard that retains core elements from the previous scorecard (e.g., vulnerability, adaptation, monitoring, carbon, and sustainable operations), but adds watershed stewardship as an additional element (USFS 2020).

The commitment to address climate change was reinforced with formalization of the 2012 Planning Rule (USFS 2012b) and issuance of subsequent management directives (USFS 2015a). Over the past decade, the Forest Service has made considerable progress in addressing climate change through scientific assessment and partner engagement, and some national forest units have begun incorporating climate considerations into their land management plan revisions under the new planning rule. To date, climate change vulnerability assessments have been completed for nearly the entire National Forest System, using a variety of approaches and for a wide range of forest types, geographies, spatial scales and resolutions, and time frames (Brandt et al. 2017, Timberlake and Schultz 2019). There has also been progress in understanding carbon stocks on federal forest lands, with data regularly compiled, maintained, and made publicly

available under the Forest Service’s Forest Inventory and Analysis (FIA) Program (Radtke et al. 2017, Smith et al. 2019, Wurtzebach et al. 2020).

To help integrate climate change science and local management knowledge and needs, the agency has championed the development of science–management partnerships (Peterson et al. 2011, Littell et al. 2012, Halofsky and Peterson 2016, Ontl et al. 2018, Timberlake and Schultz 2019). In addition, the U.S. Department of Agriculture (USDA) has established a network of regional “Climate Hubs” that represent a collaboration across the department’s agencies to develop and deliver science-based information and technologies to agricultural and natural resource managers (USDA, n.d.). This includes efforts such as the Climate Change Response Framework, which helps forest planners and managers integrate climate considerations into their work through tools like the Adaptation Workbook (CCRF, n.d.), and the Adaptive Silviculture for Climate Change (ASCC) project, which is carrying out experimental trials designed to identify measures that might be effective in preparing forest ecosystems for climatic changes (Nagel et al. 2017). Together, these programs and activities play a critical role in advancing the Forest Service’s commitment to managing for resilience through climate adaptation and mitigation efforts, which are at the heart of climate-smart restoration and management (Solomon et al. 2009).

However, there remains a disconnect between the progress at the institutional and research levels and implementation of climate-smart forest restoration and management in practice (Laatsch and Ma 2015). As noted by Hart et al. (2015), this is due, in part, to differing perspectives between restoration scientists, restoration practitioners, and stakeholders about whether, and the degree to which, restoration represents a return to historical conditions or one that explicitly incorporates the potential for new stresses and novel forest conditions. Addressing this challenge underscores the need for a more forward-looking, collaborative approach to national forest management. Building a shared, science-based understanding of climate-smart forest restoration among Forest Service

practitioners, their partners, and the public is critical to achieving the agency's mission (Albrich et al. 2018). Toward this, knowledge of recent advances in forest and climate science, as well as the ongoing evolution in the concepts and principles of climate-smart and ecologically appropriate restoration, can provide a strong foundation for furthering both the dialogue and actions necessary for progress.

### 3.2. PROPOSED PRINCIPLES FOR CLIMATE-SMART FOREST RESTORATION

To inform further discussion and help advance a shared understanding of climate-smart restoration and management of America's national forests, we propose the following overarching principles. These principles draw from the Forest Service's climate-related work, from the principles underlying the National Wildlife Federation's climate-smart conservation adaptation planning framework (Stein et al. 2014), from the climate adaptation experience of other partners and practitioners, and from the vast and growing body of scientific literature summarized in the chapters below.

- **Look to the future while learning from the past.**

Forest planners and their partners should develop forward-looking goals for management that build on an understanding of the historical range of variability and past responses to disturbance, but account for and anticipate future climate-related changes.

- **Embrace functional restoration of ecological integrity.**

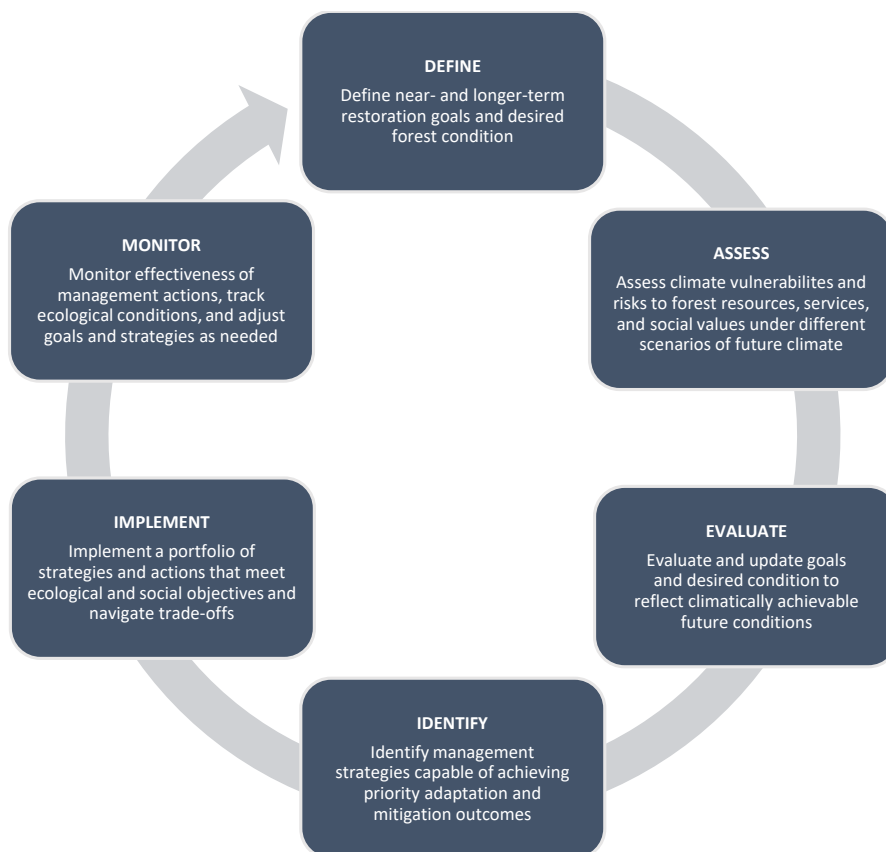
As climatic conditions continue to change, it will become increasingly difficult to restore the ecological integrity of forest systems based on historical species compositions and structures. Rather, goals for ecological integrity should emphasize the capacity of forest systems to adapt and adjust, including through enhancing functional diversity and habitat complexity.

- **Restore and manage forests in the context of larger landscapes and longer time frames.**

Climate change necessitates that planners and managers consider larger spatial scales (e.g., watersheds, landscapes, and regions) and longer time frames to ensure that localized and near-term actions do not compromise the capacity of forests to accommodate and adjust to changing conditions.



Photo: Joseph M. de Leon/USFS.



**Figure 2. Climate-smart forest restoration planning cycle.** To support application of the proposed climate-smart forest restoration principles, this iterative planning process illustrates the incorporation of future climatic conditions into forest restoration planning and implementation. Adapted from Stein et al. (2014), Swanston et al. (2016), Golladay et al. (2016), Meyer et al. (2021), and NPS (2021).

- **Adopt agile planning and management approaches that accommodate and address uncertainty.** Restoring and managing forests in the face of continuous climatic change requires decision-making under uncertainty, underscoring the importance of adaptive planning and management, including the consideration of multiple plausible scenarios of future conditions.

- **Address climate risks by linking adaptation strategies to key climate-related impacts.** Understanding climate vulnerabilities and risks to priority forest resources and values serves as the basis for developing and implementing adaptation strategies that are capable of reducing risks and sustaining the ecological, social, and economic systems associated with national forests.

- **Manage for change, not just persistence.** As climatic conditions exceed historical ranges of variability, national forest planners will need to consider how to reconcile “desired” future conditions with climatically achievable future conditions. Planners and managers increasingly will need to determine when and where it may be possible to manage for the persistence of current/historical forest conditions, and when it may be necessary to manage for change by accepting or even facilitating ecological transitions.

- **Optimize, rather than maximize, carbon sequestration opportunities.** National forests will play an increasingly important role in achieving the nation’s climate mitigation goals. Attempting to maximize carbon sequestration and storage, however, can undermine other important ecosystem services and

national forest values. Managers should instead seek to optimize sequestration opportunities by balancing carbon goals with other important forest restoration, management, and resilience outcomes.

- **Enhance collaboration to identify shared values, navigate trade-offs, and maximize synergies in the context of changing conditions.** Managing forests for multiple, sustained ecosystem services will necessarily entail trade-offs, particularly given the challenges and uncertainties associated with changing climatic conditions. Engaging local communities and diverse constituencies as early as possible in the forest planning process helps gain buy-in and identify opportunities to minimize trade-offs and maximize synergies, including acknowledgment and discussion of the potential for fundamental changes in national forest conditions.

To help forest managers put these principles into practice, we also offer a proposed climate-smart forest restoration planning cycle (see Figure 2). This adaptive planning framework emphasizes the need to clearly define and articulate restoration goals and objectives, to understand how current and future climatic conditions may affect forest resources and the services they provide, and to reevaluate and update goals that may be climate-compromised and unachievable under projected future conditions. Based on an understanding of climate vulnerabilities and risks, targeted management strategies can be identified that are designed to enhance forest resilience and reduce key climate risks (i.e., achieve adaptation outcomes) as well as reduce greenhouse gas emissions and optimize carbon sequestration and storage capacity (i.e., achieve mitigation outcomes). A portfolio of strategies and actions can then be selected for implementation that collectively help achieve key ecological and social objectives and account for trade-offs. Monitoring the effectiveness of management actions and ecological responses is essential for understanding when and how strategies and goals may need to be adjusted to reflect changing climatic, ecological, or social/economic conditions. Indeed, this framework fundamentally represents an adaptive management loop, and assumes an iterative approach to planning, implementation, and refinement.



*Changing conditions will place new strains on forest resources, such as freshwater supplies, and will require increased attention to working with diverse groups of stakeholders. Photo: Mark Doliner/Flickr.*

This proposed climate-smart forest restoration cycle draws from and builds on a variety of existing climate adaptation and adaptive management planning frameworks. This includes the climate-smart conservation adaptation planning cycle that was developed by the National Wildlife Federation in collaboration with a number of U.S. federal agencies including the Forest Service (Stein et al. 2014), as well as the Forest Service’s Climate Change Response Framework and Adaptation Workbook (Swanston et al. 2016).

The following sections offer a deeper exploration of the science underlying the proposed climate-smart forest restoration principles and planning cycle. Chapter 4 (Restoration in the Context of Future Change) examines the first four principles, providing an overview of the challenges associated with restoration as conditions move beyond historical ranges of variability. Chapter 5 (Managing National Forests for Adaptation and Resilience), delves more deeply into adaptation outcomes, discussing principles that focus on reducing climate risks and managing for ecological change. Chapter 6 (Managing National Forests to Achieve Natural Climate Solutions), in turn, looks at climate mitigation and the principle related to carbon sequestration. Finally, chapter 7 (Balancing Trade-offs Among Ecosystem Services) considers some of the trade-offs that may emerge in seeking to achieve adaptation and mitigation outcomes and explores how collaborative processes can help in navigating those trade-offs.



Forest ecosystems have always been dynamic, but the scale and intensity of disturbances, including hurricanes, wildfires, and pest infestations, increasingly are exceeding historical bounds (Great Dismal Swamp National Wildlife Refuge, North Carolina). Photo: Rob Wood/USFWS.

## 4. RESTORATION IN THE CONTEXT OF FUTURE CHANGE

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**T**raditionally, the goal of ecological restoration has been the return of degraded systems to some chosen preexisting natural condition (Clewell et al. 2002). As management of natural resources began to place greater emphasis on ecosystem health and integrity, the development of the related concepts of *historical range of variability* (sometimes referred to as *historical range of variation*) (HRV) and *natural range of variability* (NRV) offered an important advance in identifying restoration and management benchmarks.<sup>1</sup> Rather than specifying a single reference condition, HRV and NRV emphasize the

dynamic nature of ecosystems and account for periodic fluctuations (whether interannual or decadal) in the system driven by natural events over time and across landscapes (Swanson 1994, Landres et al. 1999).

Within the Forest Service, the concept of HRV played an important role in the agency's adoption of "ecosystem management" in the 1990s as an alternative to more resource extraction-oriented management approaches (Keane et al. 2009). More recently, NRV has become embedded in agency planning and practice, with the 2012 Forest Planning Rule requiring a determination

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<sup>1</sup> Because of the difficulty of defining "natural," particularly in the context of a changing climate, we find the concept of "historical range of variability" to be more appropriate for application in climate adaptation and climate-smart restoration. However, given the prominence of the term "natural range of variability" in recent U.S. Forest Service management directives, we recognize that it may be more commonly applied in the agency's work.

of NRV for a suite of ecological characteristics as a basis for measuring “ecological integrity” (USFS 2012b, Safford and Stevens 2017). Continuing and accelerating climatic changes, however, are complicating the application of these concepts in restoration, and forest practitioners increasingly will need to anticipate and manage for conditions that exceed historical ranges of variability.

## 4.1. LOOKING TO THE FUTURE WHILE LEARNING FROM THE PAST

Engaging in forward-looking and climate-smart forest restoration and management does not mean that the past is no longer relevant. Indeed, evaluation of where and how forest systems have existed and evolved in the past allows for greater understanding of their ecological processes and functions under a range of environmental conditions, as well as their potential response to natural and human disturbances and long-term change (Barrett et al. 2006, Löf et al. 2019, Beller et al. 2020). It also provides important insights into cultural connections to forests over time (Higgs et al. 2014). Thus, assessment of HRV/NRV in national forest planning remains appropriate as a tool to inform management and restoration efforts. However, the 2015 *Land Management Planning Handbook (Handbook)* acknowledges that NRV does not necessarily constitute a management target or desired condition, stating that “[f]or specific areas within an ecosystem, the Responsible Official may determine that it is not appropriate, practical, possible, or desirable to contribute to restoring conditions to the natural range of variation” (USFS 2015a). It also suggests that it may be necessary to manage an ecosystem for characteristics that were either rare or never occurred in the past in order to support its ability to withstand or recover from disturbance events caused under unique circumstances (USFS 2015a). Thus, forest restoration in a changing climate must look to the future while learning from the past.

A rapidly changing climate, in particular, is compromising the idea among many managers that restoring forest conditions to those reflected by their HRV/NRV is an optimal, or even achievable, outcome (Millar et al. 2007, Millar 2014, Hanberry et al. 2015, Golladay et al. 2016, McKelvey et al. 2021). Although the concepts recognize and reflect dynamic and cyclic conditions, the general assumption has been that these variations typically revolve around a mean value (i.e., despite fluctuations, the value remains stable over the long term). Yet, as Milly et al. (2008) and Craig (2010) acknowledge, in the context of climate change, “stationarity is dead.” Directional shifts in climatic factors (e.g., long-term warming trends) will in many places eventually exceed the bounds of the historical range of variability. Additionally, for some factors (e.g., precipitation), there is also a shift toward greater extremes (Hayhoe et al. 2018). As a result, sooner or later many forests will experience conditions that are well outside the limits of historical variability.

In fact, evidence suggests this may already be occurring in some places, as forest systems are nearing or have surpassed ecological thresholds due to extreme conditions beyond what they likely have experienced in the past (e.g., Young et al. 2019, Hansen et al. 2020). Therefore, a growing number of scientists argue that exclusively relying on a selected set of historical conditions as benchmarks for forest restoration outcomes will be increasingly untenable (Millar et al. 2007, Alagona et al. 2012, Millar 2014, Dumroese et al. 2015, Jacobs et al. 2015, Millar and Stephenson 2015,

***Look to the future while learning from the past.*** Forest planners and their partners should develop forward-looking goals for management that build on an understanding of the historical range of variability and past responses to disturbance, but account for and anticipate future climate-related changes.

1958



2001



*Directional trends in climatic factors are pushing many natural systems beyond the bounds of historical variability. Warming in Alaska, for instance, is leading to a loss of permafrost and contributing to the expansion of trees in formerly treeless tundra (Denali National Park and Preserve, Alaska). Photo: L.A. Viereck.*

Golladay et al. 2016, Gann et al. 2019, Prober et al. 2019, Beller et al. 2020, Donato et al. 2020, McKelvey et al. 2021, Meyer et al. 2021). Rather, information about historical ecological conditions and processes should be used to support a more forward-looking and anticipatory approach to restoration and management (e.g., Heller and Hobbs 2014, Stanturf et al. 2014, Golladay et al. 2016, Bradford et al. 2018, Timberlake et al. 2018). As noted by Meyer et al. (2021), the applicability of reference conditions identified by NRV analysis may be reduced under rapidly changing environmental conditions. They suggest that although historical ecological information is still important, NRV-based management targets may require modification, or treated as “waypoints” rather than “endpoints” (Meyer et al. 2021).

Given the challenges in applying HRV/NRV in management decisions under a rapidly changing climate, some scientists suggest augmenting the use of HRV/NRV with the concept of *future range of variability* (FRV) which incorporates projections for the species assemblages and disturbance regimes that will characterize future forest conditions (McComb and Duncan 2007; Duncan et al. 2010; Seidl et al. 2016; Keane et al. 2009, 2018, 2019; Loehman et al. 2020; Meyer et al. 2021). Such analyses can then be used to evaluate the potential effectiveness of various management decisions and help identify intended and unintended consequences of those options (see section 4.4) (Mozelewski and Scheller 2021).

## 4.2. EMBRACING FUNCTIONAL RESTORATION OF ECOLOGICAL INTEGRITY

The challenge of applying the concepts of HRV/NRV in national forest management in an era of climate change also requires an explicit evaluation of what restoring *ecological integrity* means (Dumroese et al. 2015, Wurtzebach and Schultz 2016, Carter et al. 2019, Rohwer and Marris 2021). Under the 2012 Planning Rule, ecological integrity is defined as “[t]he quality or condition of an ecosystem when its dominant ecological

**Embrace functional restoration of ecological integrity.** Goals for ecological integrity should emphasize the capacity of forest systems to adapt and adjust, including through enhancing functional diversity and habitat complexity.

characteristics (for example, composition, structure, function, connectivity, and species composition and diversity) occur within the natural range of variation and can withstand and recover from most perturbations imposed by natural environmental dynamics or human influence” (USFS 2012b). This reflects the traditional application of ecological integrity, which is based on selecting a pre-European-influenced reference period—an approach echoed in the 2015 *Handbook* as it describes the natural range of variation (Wagner 2000, USFS 2015a, Zellmer et al. 2018).

Importantly, the *Handbook* suggests a longer-term perspective of the historical record than has generally been considered in national forest restoration, stating that “[t]he pre-European influenced reference period considered should be sufficiently long, often several centuries, to include the full range of variation produced by dominant natural disturbance regimes such as fire and flooding and should also include short-term variation and cycles in climate” (USFS 2015a). Drawing from the field of historical ecology, this approach builds on the idea that considering longer time horizons into the past (e.g., centuries to millennia) will help forest managers better understand the dynamics and complexities that confer ecological resilience over time (see section 5.2.1) (Swetnam et al. 1999, Millar and Brubaker 2006, Littell et al. 2016, Colombaroli et al. 2017, Zellmer et al. 2018, Beller et al. 2020, D’Amato and Palik 2021). The application of such a long-term perspective in practice depends on whether sufficient data (e.g., from paleorecords) exist to support such analysis, among other challenges (Keane et al. 2009, Safford et al. 2012). Traditional ecological knowledge can also provide valuable insights into past ecological changes and current ecosystem structure (e.g., Vinyeta and Lynn 2013, McClenachan et al. 2015).



Prescribed fire is an important tool for achieving functional restoration in many fire-adapted forest ecosystems (Oconee National Forest, Georgia). Photo: USFS.

The *Handbook* also underscores the need for a more future-oriented perspective on ecological restoration, which it describes as “[t]he process of assisting the recovery of an ecosystem that has been degraded, damaged, or destroyed.” Specifically, “[e]cological restoration focuses on reestablishing the composition, structure, pattern, and ecological processes necessary to facilitate terrestrial and aquatic ecosystems’ sustainability, resilience, and health under current *and future conditions*” (emphasis added) (USFS 2015a). It also highlights the concept of *functional restoration*, which focuses on the underlying processes that may be degraded, acknowledging that the functionally restored system may look different from the reference condition in terms of the composition and structure of vegetation patches and/or their spatial pattern. The *Handbook’s* definition of ecological restoration is an important update of previous terminology, and the concepts it espouses provide a critical underpinning for effective climate adaptation and mitigation activities throughout the National Forest System.

### 4.3. RESTORING AND MANAGING FORESTS IN THE CONTEXT OF LARGER LANDSCAPES AND LONGER TIME FRAMES

Forest managers have long recognized the importance of taking a large landscape approach. The emergence of ecosystem management approaches since the 1990s has helped institutionalize thinking at scales larger than a forest stand or management unit, including efforts to incorporate regional-scale processes such as gene flow, successional pathways, and disturbance regimes (e.g., Seidl et al. 2012). Nevertheless, with climate-driven range shifts among species already underway, the need for considering and managing across broader geographic scales takes on an added significance (Stein et al. 2014). From an ecological perspective, focusing on management of individual

national forest units, while still necessary, may not be sufficient to meet conservation goals given the strong potential for dynamic range shifts, emergence of novel (also called *no-analog* or *emerging*) species assemblages, and changes in biotic interactions (Joyce et al. 2008, Stein et al. 2014, Olson et al. 2017, Dudley et al. 2020).

The 2012 Planning Rule recognizes the importance of considering conditions outside the boundaries of national forests and enhancing *functional connectivity*, which it defines as the “ecological conditions that exist at several spatial and temporal scales that provide landscape linkages that permit the exchange of flow, sediments, and nutrients; the daily and seasonal movement of animals within home ranges; the dispersal and genetic interchange between populations; and the long distance range shifts of species, such as in response to climate change” (USFS 2012a, Cannon et al. 2019).

From a planning perspective, broader landscape-based approaches also require coordination and collaboration across management jurisdictions and among diverse partners (Butler et al. 2013). Recognizing this, the Collaborative Forest Landscape Restoration Program (CFLRP) was established in 2009 to allow communities to work directly with Forest Service managers to

***Restore and manage forests in the context of larger landscapes and longer time frames.*** Climate change necessitates that planners and managers consider larger spatial scales and longer time frames to ensure that localized and near-term actions do not compromise the capacity of forests to accommodate and adjust to changing conditions.



Carrying out forest restoration at landscape scales requires broad collaboration and engagement with nearby communities and other stakeholders (Four Forest Restoration Initiative, Arizona). Photo: USFS.

plan, implement, and monitor large-scale restoration programs. Effective collaboration is also exemplified by the National Cohesive Wildland Fire Management Strategy, which was developed through a partnership between federal, state, local, and tribal governments, nongovernmental partners, and public stakeholders to achieve greater social and ecological resilience to wildfire at a national scale (DOI and USDA 2014). Yet, as described further in section 7.2, the added complexities and uncertainties associated with climate change have made it especially challenging for participants in such collaborative efforts to define desired outcomes for restoration.

Another important consideration is the relevant time frame for adaptation strategies and actions. Although forest managers are accustomed to accounting for long-term implications of decisions given the long lifespan of trees, there is often a mismatch between forest planning time frames<sup>2</sup> and the need to consider potential climatic changes and associated impacts throughout this century and beyond (Timberlake and Schultz 2019). Further, governance challenges, such as annual budgeting and turnover among leadership and local staff, can hinder the ability of forest managers to achieve long-term ecological restoration goals (Schultz et al. 2019b).

While consideration of longer-term effects and ecological and human responses to climate change does not replace short-term operational and management planning, it provides an important strategic context for near-term decisions. Ideally, near-term management decisions and actions should align with and help advance longer-term climate adaptation and mitigation outcomes. Failure to consider the longer-term climatic changes affecting the national forest might result in decisions that will be either ineffective in meeting forest management goals, translate to lost opportunities for climate adaptation and mitigation,

or, worse, compromise the capacity of forests to accommodate and adjust to changing conditions and exacerbate potential negative effects and climate risks.

#### 4.4. ADOPTING AGILE PLANNING AND MANAGEMENT APPROACHES THAT ACCOMMODATE AND ADDRESS UNCERTAINTY

Despite considerable advances in the use of climate models and projections in forest management, ongoing concerns about uncertainty in understanding future climatic conditions and associated impacts has created some reticence about applying the concept of FRV in national forest planning (e.g., Haugo et al. 2015, DeMeo et al. 2018). To be sure, articulating goals for future conditions can be challenging. In particular, climate change may require managers to aim for moving targets (Luce et al. 2012). For instance, some impacts to forests will be gradual, such as shifts in species' ranges as temperature and precipitation patterns change. Others will be rapid, including possible transformations to non-forest ecosystem types due to exceedance of key climatic or fire tolerance thresholds (Crausbay et al. 2017), or forest regeneration failures due to loss

***Adopt agile planning and management approaches that accommodate and address uncertainty.*** Restoring and managing forests in the face of continuous climatic change requires decision-making under uncertainty, underscoring the importance of adaptive planning and management, including the consideration of multiple plausible scenarios of future conditions.

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<sup>2</sup> National forest plans have an expected duration of no longer than 15 years, although project and budget planning generally are based on much shorter timelines.



The iconic coast redwood of California has been suggested as a possible candidate for assisted migration to ensure its persistence into the future (Humboldt Redwoods State Park, California). Photo: Kirt Edblom/Flickr.

of seed sources after intense fires (Coop et al. 2020). Nevertheless, climate-smart restoration necessitates managing under uncertainty.

The need to reconsider management goals, and not just strategies, is a core climate adaptation principle that acknowledges how changing climatic conditions may compromise the ability to achieve existing goals and desired future conditions (Stein et al. 2014). Recognizing this challenge, Golladay et al. (2016) suggest the adoption of risk-based approaches to incorporate uncertainty into forest management efforts, with management goals based on what they call *achievable future conditions*. Their conceptual framework involves engaging managers and scientists in collaborative, scenario-based planning and adaptive management, building on existing ecosystem management principles and best practices. The authors also emphasize that determining restoration goals is necessarily about human values. Thus, forest restoration efforts must capitalize on collaboration and development of shared goals, reflecting both desired biophysical conditions and the relevant social dimensions (Duncan et al. 2010, Stanturf et

al. 2014, McKelvey et al. 2021). While science can help inform what outcomes may be achievable, society must determine what is acceptable or even desirable. As noted by Jenkins and Jenkins (2017), a critical first step toward development and success of innovative management approaches is the “inclusion of biological considerations in tandem with both social and economic concerns, for which different stakeholders have their own multi-dimensional and complex objectives, metrics, and definitions.” Forest management decisions will invariably involve dealing with trade-offs, such as whether strategies should aim toward maintaining particular species, genetic resources, or biotic assemblages; production of particular resources such as timber, forage, or other non-timber forest products; whether greater emphasis should be placed on managing for ecosystem processes or functions; and whether truly novel forest conditions are acceptable, such as if they support valued ecosystem services like water production or carbon storage (Hayward et al. 2016, Rissman et al. 2018, van Kerkhoff et al. 2019). Such trade-offs in setting restoration goals and management targets are explored in more detail in chapter 7.



*Managing for change, including ecological transformations, will be an increasingly important part of forest adaptation and restoration planning (ghost forest in coastal South Carolina). Photo: William Conner/Clemson University.*

## 5. MANAGING NATIONAL FORESTS FOR ADAPTATION AND RESILIENCE

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**W**hile climate adaptation has been recognized as being important for North American forest management since the 1990s and early 2000s (e.g., Duinker 1990, Wall 1992, Spittlehouse 1997, Spittlehouse and Stewart 2003), an increased scientific understanding of the climate vulnerabilities of forest ecosystems, along with evidence of the accelerating pace of climate change and associated impacts, have heightened attention to the issue in recent years (e.g., Joyce et al. 2008, 2009; Blate et al. 2009; Millar and Stephenson 2015;

Halofsky et al. 2018). The Forest Service has invested in a broad range of adaptation-related activities, from developing adaptation guidance and conducting climate change vulnerability assessments at multiple scales, to engaging staff and partners in adaptation planning workshops across the country (e.g., Peterson et al. 2011, Davison et al. 2012, Furniss et al. 2013, Janowiak et al. 2014, Treasure et al. 2014, Hatfield et al. 2015, Swanston et al. 2016, Nagel et al. 2017, Halofsky et al. 2018, Timberlake and Schultz 2019).

## *Address climate risks by linking adaptation strategies to key climate-related impacts.*

Understanding climate vulnerabilities and risks to priority forest resources and values serves as the basis for developing and implementing adaptation strategies that are capable of reducing risks and sustaining the ecological, social, and economic systems associated with national forests.

### 5.1. ADDRESSING CLIMATE RISKS BY LINKING ADAPTATION STRATEGIES TO KEY CLIMATE-RELATED IMPACTS

Forest adaptation strategies are necessarily highly variable and context specific, reflecting differences in both the vulnerabilities of forest ecosystems and the institutional systems and contexts in which they are managed. No single approach will be appropriate in all cases due to a number of contextual factors, including: differences in regional and localized climate change projections; the diversity of ecosystems within and across national forests; variability in management goals and approaches across different forest units; complex and dynamic interactions among multiple forest stressors; and differences in jurisdictional settings (e.g., between the eastern and western United States and between publicly and privately owned forest lands) (Blate et al. 2009, Ontl et al. 2018, Oswalt et al. 2019, Rice 2019). Thus, it is important for scientists and forest managers to work together to identify specific science needs and resources in order to inform development of appropriate and effective adaptation strategies (Peterson et al. 2011, Littell et al. 2012, Furniss et al. 2013, Joyce and Millar 2014, Timberlake and Schultz 2017).

Effective climate adaptation relies on explicitly linking adaptation strategies and actions to climate-related impacts and vulnerabilities (Stein et al. 2014). Doing so facilitates the development of appropriate adaptation and management responses that support agreed-upon (and ideally climate-informed) goals and

objectives. Toward this, the USDA Climate Hubs, the Forest Service Office of Sustainability and Climate, and the Forest Service research stations play an essential role in fostering scientific understanding of climate impacts and promoting relevant adaptation strategies and actions across the diverse contexts of national forest management. The greater challenge is ensuring that such information is readily translatable to on-the-ground management and restoration, which underscores the importance of engaging national forest managers, partners, and other stakeholders early in the adaptation planning process (Stein et al. 2014, Olliff and Hansen 2016).

A comprehensive review and evaluation of specific adaptation strategies and tactics is beyond the scope of this review. Rather, we provide a general overview of conceptual and practical considerations for managing national forests to achieve adaptation goals, building on the Forest Service's approach of managing forests for both persistence and change (e.g., Janowiak et al. 2014, Swanston et al. 2016, Halofsky et al. 2018).

### 5.2. MANAGING FOR CHANGE, NOT JUST PERSISTENCE

As discussed previously, given the pace and magnitude of climatic changes already underway, future conditions for much of America's forests will be different from those experienced in the past, and often dramatically so. In many places, the climatic factors that helped produce the forests of the past and present have already shifted. While measuring HRV/NRV enables forest managers to understand how forests have responded to past disturbances and how much they have changed over time, ultimately, the use of historical ecosystem conditions will be less relevant—and perhaps even inappropriate—as a benchmark for

*In many places, the climatic factors that helped produce the forests of the past and present have already shifted.*

defining forest management and restoration goals (Millar 2014, USFS 2015a). Indeed, as the pace and scale of change increase, even managing for persistence of *current* conditions (as opposed to re-creating *historical* conditions) will become increasingly difficult (Jackson and Hobbs 2009). Thus, a growing emphasis on the development of forward-looking goals requires managers to acknowledge that ecological change, including possible transformation from one system type to another, may be inevitable. Accordingly, forest management and restoration in an era of rapid climate change will increasingly focus on managing for change (Millar et al. 2007, Joyce et al. 2008, Stein et al. 2013, Prober et al. 2019, St-Laurent et al. 2021b).

Managing forest systems under climate change can be envisioned as falling along a spectrum, from attempting to restore/maintain a historical or current desired state (i.e., persistence), to accepting or actively facilitating transitions to a new system state that, ideally, also support important ecological and social values. Management actions at various points along this spectrum could range from a hands-off (or passive) approach to active management approaches (St-Laurent et al. 2021b). Importantly, persistence and change are scale dependent. For example, facilitating a shift of a species to a new geography would likely be viewed as a change-oriented approach at the local scale, while at a regional scale this same action could be viewed as a persistence-oriented strategy for assuring the regional survival of the species (Stein et al. 2014).

The Forest Service’s approach to adaptation acknowledges this continuum of change and management options and has often been organized around the following high-level approaches (Peterson et al. 2011, USFS 2011a, Swanston et al. 2016):

1. Enhance *resistance* to climate-related stressors to maintain desired states and conditions;
2. Increase ecosystem *resilience* by enhancing their capacity to withstand or absorb impacts without irreversible changes in ecological processes and functions;
3. Facilitate ecological *transitions/transformation* (sometimes referred to as *response* or *realignment*); and
4. Allow forests to change on their own (i.e., *no action* or *acceptance*).<sup>3</sup>

This general framework, first described by Millar et al. (2007), has been widely cited and increasingly applied in practice (e.g., Peterson et al. 2011, Duveneck and Scheller 2016, Nagel et al. 2017, Dunn and Thompson 2018, Hagerman and Pelai 2018, James et al. 2018, Long et al. 2018, Clifford et al. 2020, Frelich et al. 2020, Kershner et al. 2020). For example, the Climate Change Response Framework, which offers a menu-based approach for adaptation planning developed by the Northern Institute of Applied Climate Science—a collaborative effort among the Forest Service, universities, conservation organizations, and the forest industry—helps forest managers connect these broad conceptual approaches to specific adaptation strategies and tactics tailored to meet specific project needs (Swanston et al. 2016, Ontl et al. 2018, Shannon et al. 2019).

### ***Manage for change, not just persistence.***

As climatic conditions exceed historical ranges of variability, National Forest planners and managers increasingly will need to determine when and where it may be possible to manage for the persistence of current/historical forest conditions, and when it may be necessary to manage for change by accepting or even facilitating ecological transitions.

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<sup>3</sup> Recent collaborative efforts have further refined and advanced the concept of managing along a continuum of ecological change. Schuurman et al. (2020), Thompson et al. (2021), and NPS (2021) have articulated a Resist-Accept-Direct (RAD) framework that describes the decision space available to managers along the persistence to transformation continuum.

As discussed below, however, there are some notable ambiguities and inconsistencies in how these general approaches are defined and measured, which some forest managers suggest has made them challenging to apply in on-the-ground decisions (DeRose and Long 2014, Bone et al. 2016, Seidl et al. 2016). Our intent here is to highlight various ways in which these adaptation-related terms and concepts have been used in the context of natural resource management, with the hope that such understanding will help clarify the important underlying principles and, in turn, enable national forest managers and their partners to more effectively apply them in practice.

## 5.2.1. Managing for Persistence: Resistance and Resilience

### 5.2.1.1. Definitions of Resistance and Resilience

In general, both resistance- and resilience-based adaptation strategies for natural resource management have tended to focus on restoring or maintaining persistence of desired conditions (i.e., an effort to limit undesirable changes) (St-Laurent et al. 2021b).

However, the nuances between the terms *resistance* and *resilience* have proved to be vexing for many resource managers, complicating their application in practice (DeRose and Long 2014, Long et al. 2018, Selles and Rissman 2020). From an ecological perspective, *resistance* is often defined as the ability of a species or ecosystem to experience stressors but not change (i.e., *ecological resistance*) (Luce et al. 2012, Parent et al. 2020). By comparison, management strategies to maintain or enhance the resistance of a focal species or system (i.e., *resistance strategies*) entail specific actions designed to forestall impacts. Such strategies may focus on enhancing ecological resistance, such as through selective thinning to increase the resistance of ponderosa pine trees to bark beetles (Kolb et al. 2007, Crotteau and Keyes 2020). They also may entail active management to prevent change, such as: maintenance of fuelbreaks around structures at risk from wildfires; use of insecticides to combat mortality from insect outbreaks; removal of juniper and other woody vegetation expanding into sagebrush or grassland systems; and maintaining or creating climate refugia (Millar et al. 2007, Morelli et al. 2016, Bottero et al. 2017, D'Amato et al. 2018, Krawchuk et al. 2020; Maestas et al. 2021).



**Resistance approaches** often focus on sustaining current ecological conditions or restoring a system to a desired historical state, as in this effort to restore grasslands by clearing encroaching junipers (Kaibab National Forest, Arizona). Photo: USFS.



Aspen declines are occurring across many parts of the West, with warming temperatures and drought conditions appearing to play a role in the onset of disease and die-backs. Photo: Jeff Sullivan/Flickr.

In its most common ecological usage, *resilience* refers to the ability of a system to return to a particular functional state after a disturbance (Holling 1973, Walker et al. 2004, Millar et al. 2007, Benson and Garmestani 2011, Swanston et al. 2016, Bryant et al. 2019, Koontz et al. 2019, Greiner et al. 2020, Gustafson et al. 2020). Based on this application, both resistance and resilience strategies are focused on sustaining persistence of the target ecological system and preventing system change or transformations. Indeed, resistance and resilience have been described by some scientists as interrelated concepts, where the capacity to resist or recover from perturbations is viewed as an essential component of system resilience (e.g., Oliver et al. 2015, Duveneck and Scheller 2016, Bryant et al. 2019, Hessburg et al. 2019, DeSoto et al. 2020). For example, techniques such as thinning and prescribed fire may enhance forest resistance to high-severity wildfire, preventing catastrophic losses. Those treatments may also facilitate forest regrowth and recovery (i.e., resilience), such as where overly dense forest cover impedes growth of shade-intolerant trees (e.g., Stevens-Rumann et al. 2013, Short et al. 2019).

One factor that makes the application of resilience challenging, however, is that the concept has also been applied where desired management outcomes suggest room for at least some amount of ecological change—a

distinction sometimes termed *adaptive resilience* (Dudney et al. 2018, Gillson et al. 2019). For example, the Forest Service's 2016 Ecosystem Restoration policy describes resilient National Forest System lands as being "self-sustaining and, if subject to disturbances or environmental change, have the ability to reorganize and renew themselves" (USFS 2016). Thus, the policy recognizes the adaptive capacity of restored ecosystems. Falk et al. (2019) describe resilience as comprising the full range of possible responses: resistance, recovery, and reorganization. Indeed, this perspective highlights a key challenge in applying HRV/NRV as a sufficient benchmark for resilience if forest conditions (e.g., composition or structure) shift outside of that range in response to climate change, yet continue to support desired ecological functions. As noted by Seidl et al. (2016), "resilience is a dynamic property, and climate change might not only change the disturbance regime ... but also impact the range of variability and recovery rates of an ecosystem." Thus, they suggest the need to also explicitly consider the FRV resulting from the combined effects of climate change to describe system resilience.

In the context of national forest management, adaptive resilience might also apply to the more societal elements of the agency's mandate. Hagerman and Pelai (2018), for instance, acknowledge that, particularly for socio-ecological systems, resilience is often seen as a function of both adaptation and transformation. Schoennagel et al. (2017) and McWethy et al. (2019) discuss the importance of adaptive resilience to wildfires, which includes implementing strategies to help human communities "live with fire." Similarly, McWethy et al. (2019) argue that communities in the West need to go beyond "basic resilience" (i.e., allowing and supporting ecosystem recovery from wildfires and helping individuals and communities manage the impacts and recover from fires) to consider both adaptive resilience (e.g., focusing on fuel management and community planning to improve fire preparedness and response), and *transformative resilience* (e.g., accept fire-catalyzed transitions in ecosystems, changing patterns and characteristics of social organization to reduce overlap between fire and human built environment) (see section 5.2.2).



**Resilience approaches** often focus on enhancing the capacity of a system to retain key functions. Beavers act as ecosystem engineers, and restoring them to a watershed can increase the availability of water for wildlife and riparian vegetation during dry periods (Uinta-Wasatch-Cache National Forest, Utah). Photo: Tom Kelly/Flickr.

### 5.2.1.2. Challenges and Opportunities in Applying Resistance and Resilience in Practice

Several studies suggest that the ambiguity in the definitions of resistance and resilience may actually be useful. For one, it provides a “common language” for resource managers to communicate adaptation-related issues across disciplines (Brand and Jax 2007, Bone et al. 2016, Selles and Rissman 2020). It also may be the case that focusing on “resilience” as an alternative to more polarizing terms, such as climate change, provides the Forest Service with “political space to act” (Bone et al. 2016, Selles and Rissman 2020).

From an on-the-ground management perspective, however, the ambiguity creates challenges (DeRose and Long 2014, Chapin and Abrams 2020, Coughlan et al. 2020). One way to more effectively apply the concepts in practice is to explicitly consider resistance or resilience *of what, to what* (Carpenter et al. 2001, Long et al. 2018, Higuera et al. 2019, Chapin and Abrams 2020). Otherwise, management efforts could be fruitless or, worse, result in maladaptive outcomes (Fischelli et al. 2016). For example, an overemphasis on thinning to increase resistance to high-severity

wildfire could leave a forest susceptible to other forms of disturbance, such as windthrow (Bone et al. 2016, Long et al. 2018). Further, if resilience is defined as having a forest rebound as quickly as possible from wildfire, then managers may choose to minimize wildfire through suppression. On the other hand, if resilience is viewed as having wildfire be “part of a healthy forest,” then management might focus on ways to enhance forest composition that allows for wildfire without significantly altering the state of the forest (Bone et al. 2016).

Charnley et al. (2017) suggest that it is also important to also ask, resilience *for what benefit*. That is, managing national forests for resilience should reflect both ecological and social values, as identified and developed through collaborative forest planning efforts. For example, managing to achieve the goal of resilience of fire-prone forest ecosystems to high-severity wildfire can be important for several purposes: protection of homes and structures; protection of timber assets and production; protection of scenic quality and recreational opportunities; and protection of certain ecological values. However, potential trade-offs exist in managing for forest resilience to wildfire

related to each of these purposes (see chapter 7). As Seidl et al. (2017) acknowledge, increased resilience can be viewed as a “positive” outcome of management (e.g., rapid recovery of clean water production in a watershed after a disturbance) or a “negative” outcome (e.g., persistence of an invasive species after disturbance).

Recognizing the operational challenges in managing national forests for resilience, Chapin and Abrams (2020) suggest the following “best practices” for forest practitioners:

- Establish a clear definition for resilience to increase understanding of the concept;
- Plan for disturbances, ecological change, and surprises;
- Embrace adaptive management and experimentation;
- Identify creative solutions for scientific and monitoring capacity; and
- Manage with (rather than against) change, including acknowledging trade-offs between short- and long-term goals.

Ultimately, improving metrics for measuring the various factors that confer forest resistance and resilience under changing climatic conditions will continue to be an important area for scientific research to better enable forest planners and managers to apply the concepts in practice (Rice 2019).

### 5.2.2. Managing for Change: Forest Transitions and Transformations

Although the Forest Service acknowledges that managing for change will be increasingly necessary for effective climate adaptation, to date, most forest adaptation strategies have focused on promoting resistance and enhancing resilience, with the intention of achieving persistence-oriented goals (e.g., maintaining forest composition, structure, or function). In a review of global literature on climate adaptation within the forest sector, Hagerman and Pelai (2018) found that recommendations for forest management tactics “overwhelmingly” focused on maintaining current ecological patterns, either through passive



*The Adaptive Silviculture for Climate Change project (ASCC) is a nationwide study designed to test different silvicultural approaches to climate adaptation, including resistance, resilience, and transition-oriented (shown here) approaches (Chippewa National Forest, Minnesota). Photo: USFS.*

approaches (no direct management intervention) or active approaches (direct management, such as removal of nonnative species). Only 10% of those recommendations considered actions to achieve transitions to new species assemblages. The authors argue that this was a surprising outcome given the growing attention among resource managers and conservation researchers on transformative approaches and the potential need to manage novel ecosystems (e.g., West et al. 2009; Stein et al. 2014; Wise et al. 2014; Kueffer 2015; Prober et al. 2015, 2019; Swanston et al. 2016; Colloff et al. 2017a, 2017b; Ontl et al. 2018; Rissman et al. 2018; van Kerkhoff et al. 2019).

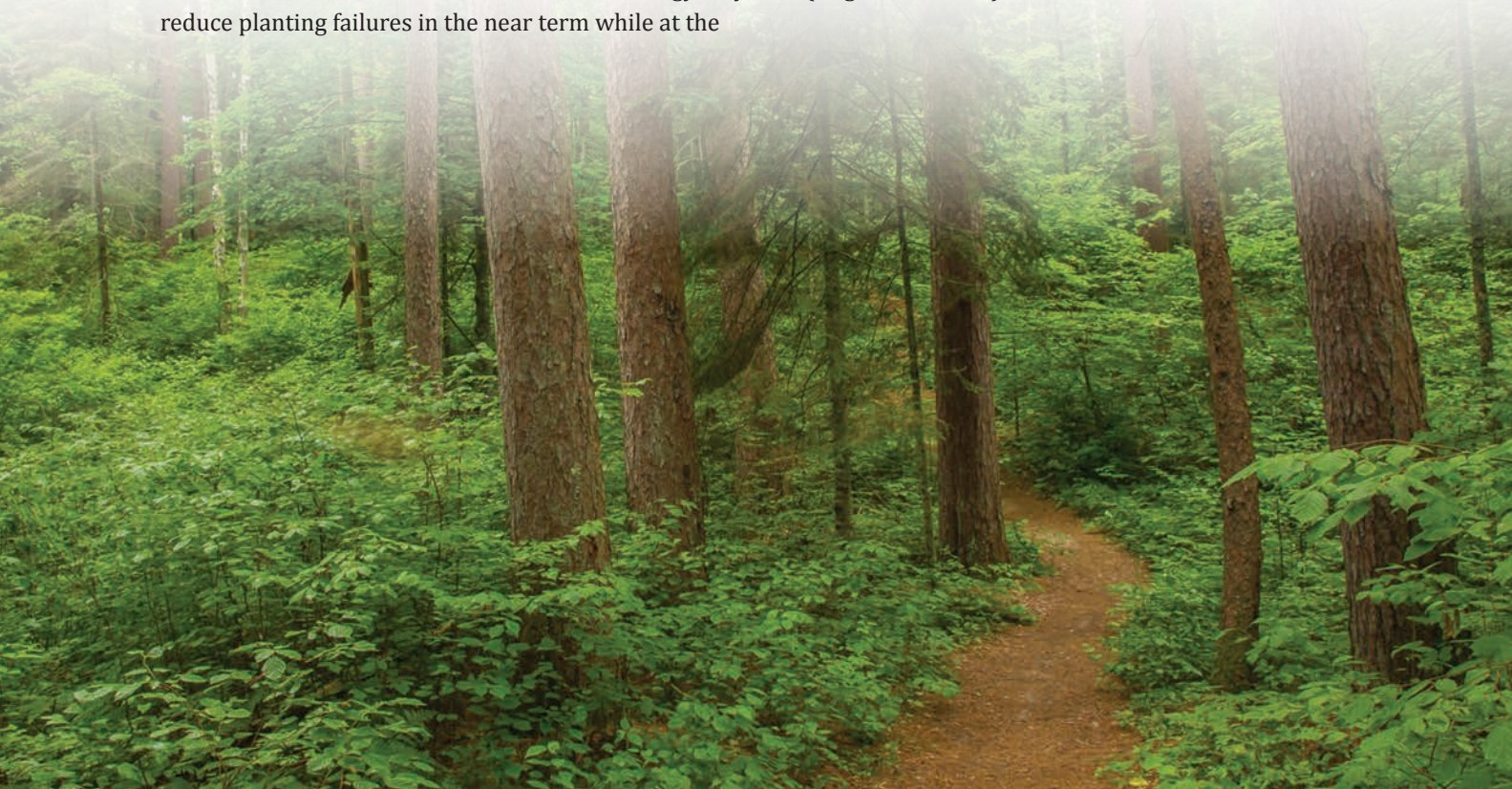
Yet, there is growing recognition that, over time, efforts to help forests resist or recover from (i.e., be resilient to) climate-related changes across the United States will become increasingly risky and costly (Millar et al. 2007, Millar and Stephenson 2015). As noted by North et al. (2009, 2019), although trying to maintain current forest structures may be futile in the face of long-term climate change, managing the transition of forests to new norms can help prevent a total collapse of an ecosystem. Ultimately, managers will need to consider whether to accommodate or accept changes, and when, where, and how they might actively direct transitions to a new desired state—strategies often referred to in Forest Service adaptation literature as *response* or *realignment* (Joyce et al. 2008; Perring et al. 2013;

Millar and Stephenson 2015; Radeloff et al. 2015; Truitt et al. 2015; Colloff et al. 2017a, 2017b; Schuurman et al. 2020; Thompson et al. 2021).

Many of the response and realignment strategies implemented by the Forest Service and its partners to date have focused on enabling forests to respond to change by seeking to work with natural adaptive processes, such as increasing redundancy and diversity of habitats and managed stands, enhancing habitat connectivity to facilitate species migration, and expanding genetic guidelines for replanting (Joyce et al. 2008, 2009; DeRose and Long 2014; Janowiak et al. 2014; Oliver et al. 2015; Spathelf et al. 2018; Isabel et al. 2020). Adaptation strategies at this level can be described as “incremental” and “anticipatory” adaptation, where incremental strategies provide benefits under current climate conditions, and anticipatory strategies apply similar techniques but are more future oriented (Stanturf 2015). Reflecting this approach, Butterfield et al. (2017) introduce the concept of “prestation”—using alternative species in restoration for which a site represents suitable habitat both now and in the future. Such a strategy may reduce planting failures in the near term while at the

same time improve the likelihood that at least some species will thrive over time. Similarly, studies across North America suggest that expanding seed sources (e.g., from zones farther south or at lower elevations) for forest restoration efforts, rather than relying on fixed zones or narrowly constrained ecotypes, may be important to facilitate adaptation over time (e.g., Aitken and Bemmels 2016, Etterson et al. 2020, Pike et al. 2020). Of note, however, sourcing seeds from a broader range of environmental conditions and ecotypes, rather than using locally adapted seeds, is contrary to what currently is considered best practices for ecological restoration (Breed et al. 2018).

In other cases, a more explicit effort to direct ecological change may be warranted—an approach that has been referred to by some as *transformational adaptation* (Kates et al. 2012, Stanturf et al. 2018). At the Chippewa National Forest in north-central Minnesota, for example, researchers are experimenting with planting ponderosa pine from the Black Hills of South Dakota, a species not native to the region but which is more drought-tolerant than the local red pine. (Nagel et al. 2017).



To prepare for expected declines of red pine (shown) under warming and drying conditions, transition-oriented experimental plantings include ponderosa pine, a more drought-tolerant species similar to red pine, but not native to the region (Chippewa National Forest, Minnesota). Photo: Brett Whaley/Flickr.

As mentioned previously, however, change is a relative term, and perspectives on what constitutes change may depend on one's values or desired outcomes. Expanding a ski area to accommodate shifting snow conditions, for instance, may be considered an undesirable change by some if it adversely affects habitat of a rare species, yet it may allow for persistence of associated recreational opportunities. This is where the concept of functional restoration is useful, in that it focuses on the "societal values and services a forest provides, rather than on the specific species compositions and structures that

formerly were present" (Dumroese et al. 2015). This necessitates understanding the societal values of forest function, not just a particular ecological condition, to determine what constitutes successful restoration (Crow 2014, Stanturf et al. 2014, Dumroese et al. 2015). Dumroese et al. (2015) acknowledge that functional restoration may well entail approaches that may not have been considered appropriate under traditional ecological restoration, such as *assisted migration* (see sidebar) and bioengineering.

## Facilitating Change Through Assisted Migration

Assisted migration, which perhaps more appropriately should be termed *managed relocation*, refers to human-assisted movement of a species in response to climate change. Assisted migration may be considered when a species is unable to disperse at a pace necessary to keep up with changing climatic conditions, when dispersal is limited by natural or anthropogenic barriers, or when the species may be subject to local extirpation or extinction without proactive management. Done thoughtfully, it also may be implemented for the benefit of the recipient system, to increase biodiversity or improve desired ecological functions (Swanston et al. 2016, Wallingford et al. 2020, Karasov-Olson et al. 2021). The Forest Service recognizes the following types of assisted migration of trees (Williams and Dumroese 2013, Handler et al. 2018):

- **Assisted population migration**, including assisted gene flow, involves moving seed sources or populations to new locations within the species' historical range;
- **Assisted range expansion** involves moving seed sources or populations from within the current range to suitable areas just beyond their historical range, facilitating natural migration;
- **Assisted species migration** involves moving seed sources or populations far outside of their historical range, beyond locations accessible by natural dispersal.

Although there are instances of assisted migration of tree species already underway, use of the strategy within the Forest Service has generally been limited to experimental research or use of "no regrets" strategies, based on relatively near-term (e.g., 20-year) climate projections (Handler et al. 2018).

To some scientists, the distinction between transformational adaptation and incremental or anticipatory adaptation lies in the degree of tolerance for novel conditions, where novelty represents “the degree of dissimilarity of a system, measured in one or more dimensions relative to a reference baseline” (Radeloff et al. 2015, Stanturf et al. 2018). For instance, Stanturf et al. (2018) suggest that incremental adaptation does not support novelty—rather, efforts are aimed at *preventing* nonnative species or emergence of native species in new combinations (i.e., “neo-native” ecosystems). The concept of anticipatory adaptation supports management of *neo-native* systems and might allow for establishment of nonnative species into a system, but only if desired ecological functions are maintained. Transformational adaptation, by contrast, is often understood to embrace and actively manage for truly novel assemblages (which may include novel combinations of native species as well as native/nonnative mixes) or different ecological systems that are considered more likely to persist under changing climatic conditions (Stanturf et al. 2018; Colloff et al. 2016, 2017a; Löff et al. 2019; Dudley et al. 2020). Here, the use of the term “transformation” can be multifold, with ecological transformation (e.g., a shift to novel assemblages or from one system type to another) leading to transformation from a societal perspective (e.g., changes in associated ecosystem services) (Colloff et al. 2017b).

### 5.2.2.1. Challenges and Opportunities in Managing for Transformation in Practice

To some forest managers, the idea of accepting or managing toward ecological transformations that differ from historical conditions may run counter to what are considered best practices for forest restoration—primary among them the focus on use of locally adapted native species (e.g., Murcia et al. 2014, Backstrom et al. 2018, Breed et al. 2018). Given the pervasive ecological and economic impacts of nonnative, invasive plant species in national forests across the country, for instance, invasive species control has understandably and importantly been a high priority for the Forest Service (D’Amato et al. 2017, Poland et al. 2021). Similarly, forest restoration has

traditionally adhered to the mantra of “local is best” when considering sources for seeds or seedlings, based on the assumption that local populations are likely to be better adapted to local conditions (Prober et al. 2015, Aitken and Bemmels 2016).

Certainly, many conventional forest restoration and management strategies will continue to be relevant and useful in an era of climate change. As some scientists suggest, it is important that we not “throw the baby out with the bathwater” (Prober and Dunlop 2011) or abandon fundamental principles of ecological restoration, such as focusing on ecosystem processes and functions (Hanberry et al. 2015). Yet, there is growing recognition that, in some areas, the rate and degree of climatic changes are going beyond the adaptive capacity of species and ecosystems to cope or adjust, which may necessitate more rigorous interventions to forestall change, or, alternatively, decisions to either accept or actively manage for change (Millar and Stephenson 2015; Dey et al. 2019; Kemp et al. 2019; North et al. 2019; Parks et al. 2019a, 2019b; Prober et al. 2019; Stevens-Rumann and Morgan 2019).

The consideration of ecological transformations as an outcome of restoration may be particularly appropriate in instances where landscapes are so altered that a return to a previously desirable set of conditions, such as those represented by HRV/NRV, is deemed unlikely or infeasible, with or without management intervention (Miller and Bestelmeyer 2016). For instance, low-elevation ponderosa pine and Douglas-fir forests in parts of the West may already have passed a threshold where regeneration is unlikely due to the combined effects of climate change and high-severity fire (Davis et al. 2019). Proponents of transformational adaptation argue that, given the pace of climate change combined with a host of additional, often synergistic stressors, novel conditions are likely to occur in many forest ecosystems, even within a single tree generation (Stanturf et al. 2018). Golladay et al. (2016) note that, in light of inevitable and possibly irreversible change, the ability of forest systems to provide valued ecosystem services will depend, in part, on the development of novel ecosystems. Perring et al. (2013) argue that the already existing pervasiveness of

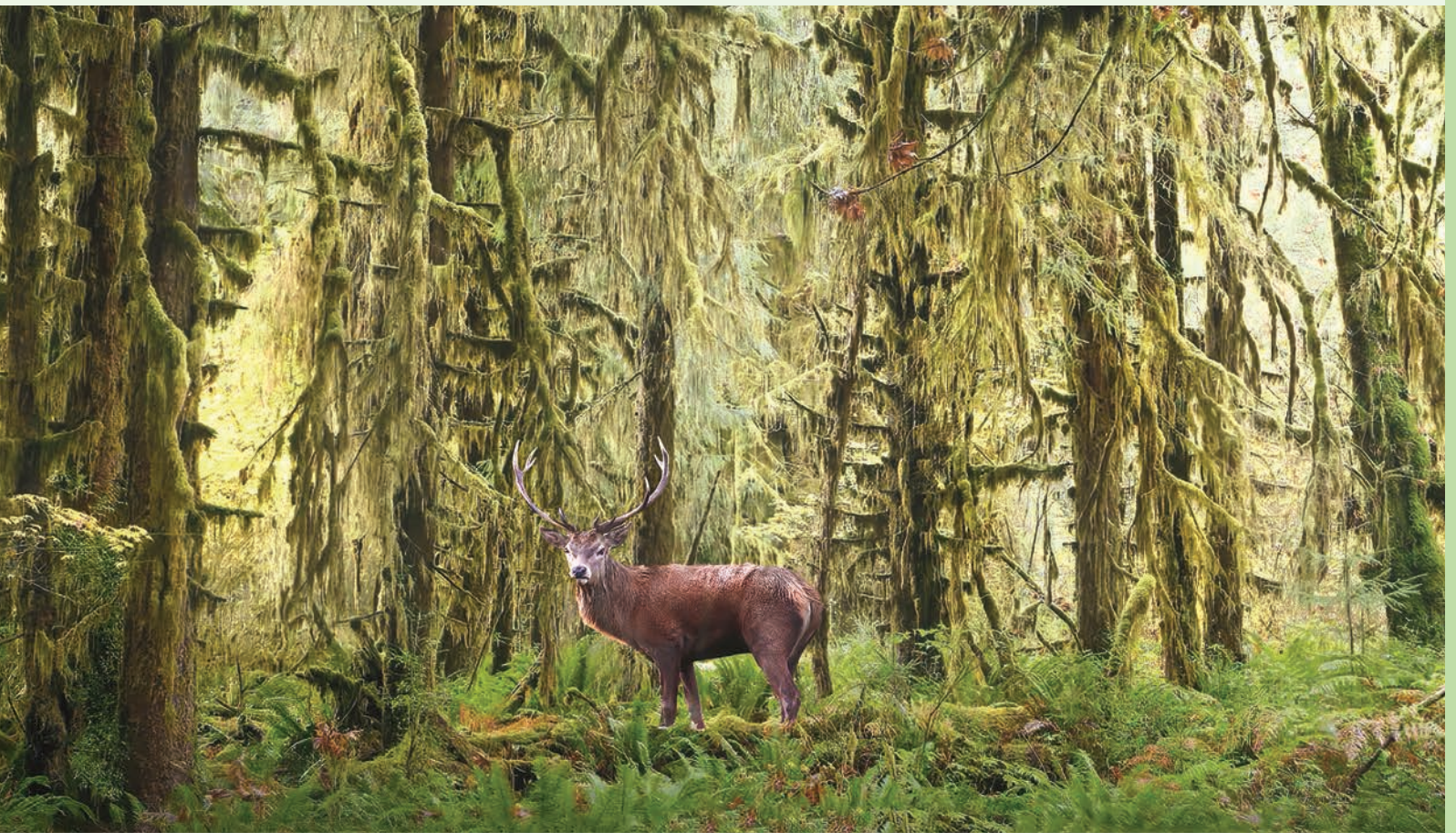
novelty and novel ecosystems underscores the need for a broader perspective for conservation and restoration. They suggest several goals for managing novel ecosystems. For example, it may be useful to focus on restoring ecosystem services through experimental and innovative approaches (i.e., “test beds”) to determine the most ecologically and ethically sound approaches. The Forest Service’s Experimental Forests and Ranges, for instance, are likely to be especially relevant for this approach. Another option entails conservation of desired species through the use of novel components, or the creation of new ecosystems to achieve desired goals, such as afforestation in highly degraded areas (e.g., on abandoned mine sites) (Evans et al. 2013).

For forest systems that are either nearing or have surpassed ecological thresholds and are unlikely to recover, proactive measures may enable managers to achieve certain desired outcomes compared to strategies (e.g., no action) that leave the system to shift spontaneously. This necessitates a better understanding and anticipation of potential ecological thresholds in terrestrial and aquatic forest systems, which can be challenging (Keenan 2015, Cantarello et al. 2017). Several studies also underscore the importance of identifying indicators to help managers

anticipate potential ecological tipping points—thresholds that, if reached or exceeded, lead to fundamental system change (e.g., Reyer et al. 2015, Scheffer et al. 2015, Johnstone et al. 2016, Miller et al. 2019, Sánchez-Pinillos et al. 2019, Turner et al. 2020). Often, tipping points are only identified once they have been crossed (Stein et al. 2014). As noted by Millar and Stephenson (2015), however, being able to anticipate an approaching forest transition can help managers identify possible management strategies to ensure that delivery of desired ecosystem services is maintained. Recent research on the response of forests to disturbances such as extreme droughts, massive wildfires, and severe insect outbreaks has offered important insights into the role of thresholds and ecological conversions (e.g., Koch et al. 2009, Adams 2013, Allen et al. 2015, Reyer et al. 2015, Scheffer et al. 2015, Teskey et al. 2015, Johnstone et al. 2016, Sánchez-Pinillos et al. 2019, Batllori et al. 2020, Turner et al. 2020). On the Kenai Peninsula, Alaska, for instance, researchers have been assessing the potential trajectories of ecological systems to identify where prospective management, including active facilitation of ecological conditions, may be possible (Magness and Morton 2018).



*High severity fire and changing climatic conditions can compromise seedling establishment and natural regeneration, leading in places to conversion from forests to non-forested systems. (Santa Fe National Forest, New Mexico) Photo: Los Alamos National Laboratory.*



Managing forest ecosystems for climate resilience, such as in the carbon-dense temperate rainforests of the Pacific Northwest, is essential to ensure that forest-related climate mitigation strategies will offer long-term benefits (Olympic National Park, Washington). Photo: Jarr1520/Flickr.

## 6. MANAGING NATIONAL FORESTS TO ACHIEVE NATURAL CLIMATE SOLUTIONS

**S**trategies that support or enhance the ability of natural systems to store and sequester carbon—often referred to as *natural climate solutions*—are considered essential to help stabilize the climate and mitigate climate change at a scale necessary to prevent the most catastrophic impacts (Griscom et al. 2017, IPCC 2018, NASEM 2019). Forests play a critical role in the global carbon cycle, with carbon flowing from the atmosphere to forests and back through processes such as photosynthesis, respiration, decomposition, and combustion (Ryan et

al. 2010). In general, about 60% of carbon in mature forests is stored in live and dead tree matter, with the other 40% in soil and forest litter (Ryan et al. 2010). Carbon gains and losses in forest systems fluctuate, with gains in carbon (i.e., sequestration) occurring largely through vegetative growth, and losses occurring through respiration, disturbances (e.g., fires and insect outbreaks), and harvest. Whether a forest is a net carbon “sink” (i.e., it gains more carbon than it loses) or “source” (i.e., it loses more than it gains) depends on the balance among each of these factors

*Strategies that support or enhance the ability of natural systems to store and sequester carbon—often referred to as natural climate solutions—are considered essential to help stabilize the climate.*

## *Forest ecosystem carbon stocks on Forest Service lands alone account for more than 26% of the nation's total forest carbon stocks—more than for any other federal land agency.*

(Ryan et al. 2010). In turn, how much net CO<sub>2</sub> enters the atmosphere over time depends on factors such as rate and extent of forest regrowth, how quickly forest matter combusts or decomposes, and how much harvested timber is stored in durable wood products.

Studies suggest that managing the nation's forests to enhance carbon sequestration has the potential to contribute significantly to achieving national and international goals for reducing atmospheric CO<sub>2</sub> (USFS 2012c, Fargione et al. 2018, Baker et al. 2019). In 2019, the forestry sector represented an estimated net uptake of 774.6 million metric tons of CO<sub>2</sub> equivalent (MMT CO<sub>2</sub> eq.), offsetting approximately 11% of total U.S. greenhouse gas emissions (U.S. EPA 2021). Based on estimates developed from current FIA database evaluations, total forest ecosystem carbon stocks on Forest Service lands alone account for more than 26% of the nation's total forest carbon stocks—more than for any other federal land agency (Smith et al. 2019).

Accordingly, maintaining forest carbon stocks and enhancing carbon sequestration have gained prominence as a goal for Forest Service management. In 2014, the USDA formally acknowledged carbon sequestration on national forests as a land-use objective (USDA 2014, McNulty et al. 2018). This policy was further clarified in the USDA's *Building Blocks for Climate Smart Agriculture and Forestry* (USDA 2016). These policies provide general guidance on management of forests for carbon; however, considerable variability in carbon sequestration potential across the country, including significant differences among ecosystems, makes it challenging to translate this broader goal to on-the-ground management actions (McNulty et al. 2018, Ontl et al. 2020).

In particular, projecting future forest carbon fluxes in the United States is uncertain and varies greatly with assumptions for future economic and political contexts, as well as physical conditions, such as forest growth trends and the influence of fire and other disturbances (Nepal et al. 2012, Coulston et al. 2015, Wear and Coulston 2015, Janowiak et al. 2017, Keeton 2018, Tian et al. 2018, Birdsey et al. 2019). For example, the amount of carbon stored in forests depends on both the area and condition of forest land and specific forest management practices, and the degree and rates of carbon sequestration and storage over time will vary considerably by region and forest type. The ability of terrestrial systems to provide for sustainable carbon sequestration also depends on the potential for significant releases of forest carbon due to more severe disturbances, land-use changes, and other factors (Seddon et al. 2019). Disturbances, for instance, can have long-lasting impacts on forest carbon stocks, as stock recovery through vegetation regrowth is far more gradual, often requiring a decades-long time horizon, when compared to the disturbance-induced carbon loss (Williams et al. 2016). In addition, there are considerable uncertainties regarding the potential for *climate change—carbon cycle feedbacks* to affect future climate change (Friedlingstein 2015, Sellers et al. 2018).<sup>4</sup> Reflecting these complexities, some studies suggest that the carbon storage capacity of U.S. forests will decline in the future (e.g., Boisvenue and Running 2010, Wear and Coulston 2015, Dugan et al. 2018), while others are more optimistic (e.g., Nepal et al. 2012, Tian et al. 2018). Thus, it will be important for forest planners and managers to consider the underlying assumptions and context of scientific assessments for carbon sequestration when developing forest climate mitigation strategies (Dugan et al. 2017, Janowiak et al. 2017).

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<sup>4</sup> Climate change—carbon cycle feedbacks may be positive or negative. Positive feedbacks occur when climate-related impacts, such as larger wildfires or thawing permafrost, lead to greater atmospheric CO<sub>2</sub>. Negative feedbacks occur when changes in climatic variables increase carbon sequestration, such as through enhanced plant growth.

## 6.1. OPTIMIZING, RATHER THAN MAXIMIZING, CARBON SEQUESTRATION OPPORTUNITIES

Enhancing the carbon sequestration potential of forests across the United States will require a concerted effort that includes scientific research, collaborative management, and supportive state and federal policies (Anderson et al. 2019). As discussed in greater detail in chapter 7, significant trade-offs may exist between carbon sequestration and other forest management goals. Accordingly, the Forest Service should strive

***Optimize, rather than maximize, carbon sequestration opportunities.*** National forests will play an increasingly important role in achieving the nation's climate mitigation goals. Managers should seek to optimize sequestration opportunities by balancing carbon goals with other important forest restoration, management, and resilience outcomes.

to optimize, rather than maximize, forest carbon, as focusing on the latter could limit the agency's ability—and mandate—to meet other important management objectives (Jackson et al. 2005, Seidl et al. 2007, Janowiak et al. 2017, Dybala et al. 2019). In particular, recent studies highlight the need for managers to consider near- and longer-term consequences of management actions in the context of both climate adaptation and mitigation (Duvencek and Scheller 2016, Hof et al. 2017, Janowiak et al. 2017, Morecroft et al. 2019, Ontl et al. 2020, St-Laurent et al. 2021a). In some forest systems, for instance, implementing strategies to restore natural patterns of fire may emit carbon in the near term, but will enhance forest health and help maintain long-term carbon pools by reducing the risk of carbon losses due to severe wildfires (Funk et al. 2019, Hurteau et al. 2019, Krofcheck et

al. 2019). Thus, the cost of such short-term emissions should be balanced against the role of disturbances in maintaining or restoring longer-term system resilience and, in return, carbon sequestration and storage (Janowiak et al. 2017, James et al. 2018, St-Laurent et al. 2021a).

Furthermore, although the carbon sequestration potential on Forest Service lands is considerable, climate mitigation strategies must be undertaken within the broader national and international policy context. For example, Fahey et al. (2010) stress that potential carbon benefits of management actions should be measurable, verifiable, and sustainable. The Forest Service has engaged in rigorous assessments of baseline carbon stocks and changes due to disturbance and management, which is essential for understanding how potential management strategies will affect forest carbon stores (e.g., Dugan et al. 2017, 2018). Yet, some scientists argue that further advances may be warranted to improve the accuracy of forest carbon accounting, such as through more effective measurement and reporting on what happens with postharvest wood (Hudiburg et al. 2019), improved understanding of post-disturbance carbon dynamics (Bartowitz et al. 2019), and a better understanding of how to measure belowground carbon (Case et al. 2021b).

### 6.1.1. Strategies to Maintain Carbon Sequestration and Storage

Maintaining and enhancing carbon sequestration and storage on forest lands can be achieved through a variety of management approaches, although the appropriateness and effectiveness of specific strategies will vary given differences in forest types, management goals, and other social and ecological contexts (Janowiak et al. 2017, McNulty et al. 2018, Domke et al. 2020). Strategies may include: avoiding conversion of forests to other land uses; increasing forest cover through reforestation and afforestation; managing forests to enhance productivity; lengthening timber harvest rotation cycles; and increasing markets for long-term wood products (Ontl et al. 2020).

*Between 1990 and 2019, “forest land remaining forest land” was the nation’s largest net carbon sink, and conversion of forest land to other uses was the largest source of land-based emissions .*

In general, both protecting existing, intact forests and restoring forest health at a large scale will be essential components of an effective, ecologically sound carbon management and climate mitigation strategy (Watson et al. 2018; Funk et al. 2019; Krofcheck et al. 2019, McCauley et al. 2019, Moomaw et al. 2019). Between 1990 and 2019, “forest land remaining forest land” was the nation’s largest net carbon sink, and conversion of forest land to other uses was the largest source of land-based emissions (U.S. EPA 2021). Nevertheless, it is important to recognize that even intact, historically unmanaged forests are at risk from changing climatic conditions. In fact, some scientists suggest that wilderness and other protected areas may be especially vulnerable to climate-enhanced wildfire and other disturbances, particularly where they are small and isolated, are sensitive to stressors from outside their boundaries, and are limited in terms of treatment options compared to more managed systems (McKenzie and Littell 2011).

As noted previously, effectively managing the nation’s national forests to provide a range of ecosystem services, including natural climate solutions,

will require a concerted effort to reduce climate vulnerabilities and enhance the capacity of forest systems to adapt to changing conditions (Janowiak et al. 2017, Ontl et al. 2020, St-Laurent et al. 2021b). The importance of restoration efforts aimed at reducing risks of catastrophic forest loss to disturbances, for example, has gained prominence in both the science and policy arenas, owing to both recent extreme events and the considerable restoration backlog within the National Forest System. Reforestation on national forests, in particular, is likely to play a significant role in efforts to enhance long-term carbon sequestration (Dumroese et al. 2019). Areas affected by wildfires, droughts, and other disturbances provide an important near-term opportunity to significantly scale up reforestation, provided conditions are sufficient for long-term growth and productivity (Sample et al. 2015, Halofsky et al. 2018, Dumroese et al. 2019, Nave et al. 2019, Meyer et al. 2021). For instance, Sample (2017) estimates that reforesting productive sites on non-stocked Forest Service–managed lands across the contiguous United States has the potential to sequester more than 16 MMT CO<sub>2</sub> eq. each year, enough to offset emissions from nearly 3.5 million cars (U.S. EPA, n.d.).



*Future climatic conditions should be a consideration in the selection of plant materials used in reforestation efforts following major disturbance events (Kaibab National Forest, Arizona). Photo: USFS.*

Where reforestation is practicable, however, scientists argue that such efforts should take broader ecological values into consideration, rather than focusing primarily on carbon. In particular, climate mitigation activities on the nation’s forest lands should be designed synergistically with adaptation outcomes. For example, North et al. (2019) suggest that efforts to enhance carbon sequestration and storage in dry forests with a history of frequent fires should promote lower and variable tree densities to lessen the risk for more high-severity fire. This multi-values perspective is consistent with the Forest Service’s emphasis on managing forests for a suite of ecosystem services. As Janowiak et al. (2017) note, a balanced, comprehensive, and science-based approach to considering carbon in forest management will help ensure that multiple ecosystem services, including carbon sequestration, are achievable.



National forests provide a wide range of ecosystem services, including clean water for downstream communities, and restoration efforts will need to account for different social values and possible trade-offs (Sumpter National Forest, South Carolina). Photo: Eric Harrison/Flickr.

## 7. BALANCING TRADE-OFFS AMONG ECOSYSTEM SERVICES

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**A**s highlighted in the 2012 Planning Rule, forest management plans are required to emphasize management of National Forest System lands so they are “ecologically sustainable and contribute to social and economic sustainability; consist of ecosystems and watersheds with ecological integrity and diverse plant and animal communities; and have the capacity to provide people and communities with ecosystem services and multiple uses that provide a range of social, economic, and ecological benefits for the present into the future” (USFS 2012b).

Framing national forest conservation from an ecosystem service perspective stems from a desire to help forest managers better articulate the broad

range of national forest benefits to the public and, in turn, build support for management actions (Ruhl and Salzman 2020). Ultimately, the agency has an interest in engaging entities who benefit from particular services, such as water resources and recreational opportunities, to assist in carrying out projects, which will enhance collaboration and help ensure more effective management outcomes at broader landscape scales given limited resources (Smith et al. 2011, Kline and Mazzotta 2012, Kline et al. 2016, Halofsky et al. 2017).

While forest planners will need to develop associated goals that can be measured and evaluated at the project level, the 2015 *Land Management Planning Handbook* specifies that forest plan revisions must focus on “key” ecosystem services—those that are important

in the broader landscape outside of the plan area and are likely to be influenced by the land management plan (USFS 2015a, Deal et al. 2017b). The *Handbook* offers flexibility for individual national forests, which is important given that “desired” conditions and ecosystem services will vary by region, forest type, and the values of stakeholders. However, given the likelihood that there will be trade-offs when managing forests for multiple ecosystem services, conflicts could ensue even among those who believe they share the same goals (Botkin 2014).

Certainly, the Forest Service has had to navigate trade-offs in land uses throughout its history. Yet, the challenges and uncertainties associated with changing climatic conditions, along with the need to achieve both climate adaptation and mitigation goals, could render managing for multiple, sustained ecosystem services difficult over the long term (Luce et al. 2012, Spies et al. 2018).

## 7.1. EXAMPLES OF TRADE-OFFS IN NATIONAL FOREST MANAGEMENT

The following examples illustrate just a few of the trade-offs that national forest managers and collaborators may face in the context of climate-smart restoration and management.

### 7.1.1. Carbon Sequestration and Other Ecosystem Services

As highlighted in chapter 6, national forests have a critical role to play in mitigating climate change through carbon sequestration and storage. Yet, some management efforts focused on enhancing carbon uptake of forests may result in outcomes that interfere with other ecosystem services, and vice versa (Strassburg et al. 2010, Creutzburg et al. 2017, James et al. 2018, Buotte et al. 2019a, Ontl et al. 2020). For example, studies suggest that strategies to maximize carbon sequestration by planting young, rapid-growing trees may lead to a reduction in water yield due to high

losses from evapotranspiration (Bennett et al. 2009, McNulty et al. 2010, Cademus et al. 2014, Duan et al. 2016). Without consideration of that potential outcome, efforts to increase carbon sequestration in some areas could lead to water shortages for human communities and make forests more susceptible to disturbances such as wildfire, particularly where higher temperatures and enhanced evaporation due to climate change are expected (Duan et al. 2016). On the other hand, there are a number of potential synergies between forest adaptation and mitigation strategies. For instance, research by Dybala et al. (2019) suggests that restoration of deforested or degraded riparian areas has the potential for rapid carbon sequestration while also enhancing floodplain connectivity and biodiversity.

Ontl et al. (2020) suggest that evaluation of potential trade-offs between carbon management and adaptation-oriented goals should account for anticipated long-term changes in carbon fluxes. To facilitate this, the authors have developed a menu to help managers align adaptation strategies and actions specifically with carbon sequestration goals (Ontl et al. 2020). The Forest Service has also developed a framework to help identify potential trade-offs among services that can be applied at national, forest unit, and project-based levels (Deal et al. 2017a, 2017b; Deal 2020). Continued engagement and investment in these and related efforts will facilitate national forest stewardship to achieve a range of ecosystem services in this era of climate change.

### 7.1.2. Wildfire Management and Carbon Sequestration

The effects of wildfire management practices such as thinning and prescribed fire on carbon stocks is likely to depend on the initial state of a forest, the types of fuel reduction treatments conducted, and the time period over which the carbon balance is assessed (Ellenwood et al. 2012). As noted previously, however, near-term carbon emissions (as well as the potential for smoke-related health impacts) associated with treatments such as prescribed burns may ultimately lead to greater forest carbon resilience (and a reduced

risk of prolonged smoky conditions associated with extreme wildfires) despite climate change and associated shifts in fire regimes. For example, a model-based assessment of a dry western forest system in the Lake Tahoe Basin of California and Nevada found that, although there was some carbon lost during simulated fuel treatments, those losses were ultimately compensated over the longer term by increased growth of residual stock due to greater available soil water and shift in species composition to more drought- and fire-tolerant Jeffrey pine at the expense of shade-tolerant, fire-susceptible white fir (Loudermilk et al. 2017). Research on the Sierra National Forest, California, suggests that, despite emitting carbon in the short term, a combination of thinning and maintenance burning would significantly stabilize forest carbon under the extreme fire conditions expected over the coming decades by reducing the risk of high-severity wildfires (Krofcheck et al. 2017).

In the Southeast, prescribed fire and thinning to enhance dominance of longleaf pine and reduce forest density to support the red-cockaded woodpecker was found to reduce total ecosystem carbon compared to untreated areas (Martin et al. 2015). Although this

represents a trade-off between restoring fire-adapted forests and increasing carbon sequestration and storage in the near term, given the long lifespan and tolerance of longleaf pine to disturbances, continued management of restored forests is expected to provide resilient, long-term carbon benefits despite changing climatic conditions (Brantley et al. 2018).

### 7.1.3. Wildfire Management and Other Ecological Goals

Achieving both ecological and risk reduction objectives on increasingly fire-prone public lands can be especially challenging, as management approaches may necessitate accepting compromises across differing goals (Barros et al. 2018, Donovan et al. 2019). For example, salvage logging (i.e., the removal of standing and down fuels following a disturbance) has often been touted as a way to lessen the risk of subsequent wildfires, despite a lack of scientific consensus on its risk reduction benefit. However, the practice can have significant adverse ecological impacts, such as through mechanical disturbance, soil compaction, and the removal of organisms, organic material, and other elements of a post-disturbance forest system that are



*Protecting surrounding communities from wildfires has always been an overarching concern for state and federal forest managers. As climate change contributes to an increase in fast moving and destructive “megafires,” wildfire risk reduction measures will continue to drive many forest restoration efforts (2020 Almeda Fire, Phoenix, Oregon). Photo: Oregon Department of Transportation/Flickr.*



*The habitat requirements for sensitive wildlife species, such as this northern spotted owl fledgling, are important to consider in setting restoration goals and designing adaptation and mitigation approaches. Photo: Zia Fukuda/BLM.*

important for forest regeneration (Lindenmayer and Noss 2006; Lindenmayer et al. 2012; Leverkus and Castro 2017; Leverkus et al. 2018, 2021; Thorn et al. 2018; Müller et al. 2019). Actions to reduce wildfire risks to communities also may run counter to achieving habitat and water quality goals (Sun et al. 2019). In parts of the Pacific Northwest, for example, maintaining roads can provide access for fuel reduction treatments and fire response, but they also fragment habitat and can lead to erosion and siltation into streams (DellaSala et al. 2011, Luce et al. 2012). Conversely, several studies suggest that limiting the use of treatments such as thinning and prescribed fire in parts of the region in an effort to protect closed-canopy habitats for species such as the northern spotted owl have led to increased risks from insect outbreaks and severe wildfire (Hessburg

et al. 2016, Spies et al. 2018). This suggests the need to carefully consider trade-offs between maintaining existing habitat conditions and promoting landscape characteristics that are more resilient to large-scale disturbances (Hessburg et al. 2016).

Schultz et al. (2019a) note that part of the reason trade-offs between wildfire risk reduction and ecological goals exist is due to ambiguous and sometimes conflicting terminology and management directives. For instance, while some documents focus on fire as an ecological process and highlight “restoration” of natural fire regimes as important for long-term forest health, others emphasize fire as a hazard that needs to be “controlled.” The authors argue that the current system of fire management leads to prioritization of

*Ecological forestry, which seeks to use knowledge of disturbance ecology and retention-based management to achieve ecological and commodity goals, has emerged as an important concept for addressing wildfire risks as well as enhancing the health of forest ecosystems.*

management for near-term risks at the expense on longer-term ecological goals. They suggest a more integrative approach that encompasses multiple problem definitions. Toward this, *ecological forestry*, which seeks to use knowledge of disturbance ecology and retention-based management to achieve ecological and commodity goals, has emerged as an important concept for addressing wildfire risks as well as enhancing the health of forest ecosystems (D'Amato et al. 2011, 2018; Palik and D'Amato 2017; Franklin et al. 2018; Spies et al. 2018; Kelsey 2019). Here, ecological forestry practices can include a combination of strategic thinning, prescribed fire, and managed wildfire (i.e., allowing some naturally ignited wildfires to burn, with selective suppression to protect property and infrastructure) to reduce the risk of high-severity wildfire and promote healthier, more resilient forests (Stephens et al. 2016, 2018; Kelsey 2019). Done thoughtfully, the approach can help balance trade-offs between short-term impacts of treatment with long-term benefits of reduced risks of large, high-severity fires (Kelsey 2019, Krofcheck et al. 2019, Stephens et al. 2021). For example, Hessburg et al. (2016) recommend a combination of managed wildfire, prescribed fire, and variable thinning in parts of the Pacific Northwest to support heterogeneous landscapes that confer resistance to severe wildfires as well as enhance habitat for old-growth-dependent species.

#### 7.1.4. Wildfire Management and Timber Production

Although the volume of timber harvested on Forest Service lands has declined considerably from its peak in the mid- to late-20th century, commercial timber production remains an important activity for the agency (USFS, n.d.). Ager et al. (2017) note that, although efforts to optimize revenue to help finance restoration can lead to a sharp reduction in the attainment of other socioeconomic objectives, such trade-offs vary considerably among different planning areas and forest units. For example, Vogler et al. (2015) assessed potential trade-offs among restoration projects on the Wallowa-Whitman National Forest in eastern Oregon, which is considered a national priority for restoration to increase fire resilience and sustain a

range of ecosystem services. One of the sharpest trade-offs they found was between net revenue from forest management activities and implementing projects that treat areas likely to contribute to fire in the surrounding wildland–urban interface. This creates a management challenge, as the costs of forest restoration projects are often offset through production of timber revenues.

Indeed, across the western United States, conflicting management objectives may occur between ecological, risk reduction, and wood production goals, which has implications for revenues to support Forest Service restoration activities as well as rural economies. As described by Ager et al. (2019), the generalization that “logging necessarily reduces wildfire risk has merit only if one considers that removing value from land (i.e., logs) reduces expected loss more than it affects fire spread and severity.” To address these challenges, scientists in northern California have been working with traditionally natural resource–dependent communities to become more engaged in a variety of

*The generalization that “logging necessarily reduces wildfire risk has merit only if one considers that removing value from land (i.e., logs) reduces expected loss more than it affects fire spread and severity.” (Ager et al. 2019).*



*Although the volume of timber harvested on Forest Service lands has declined considerably from its peak in the mid- to late-20th century, commercial timber production remains an important activity for the agency (Chattahoochee National Forest, Georgia). Photo: Cecilio Ricardo/USFS.*

post-fire restoration activities as a way to enhance local economies (Long et al. 2014). Although salvage logging for timber is among the strategies considered, there has been a concerted effort to identify approaches that are appropriate to, and less harmful of, forest ecology. As noted by Ryan and Hamin (2009), research on post-fire restoration suggests that salvage logging is likely to be more socially acceptable if done in ways that do not harm the local ecology and, in turn, provide revenues for forest restoration.

## 7.2. ENHANCING COLLABORATION TO IDENTIFY SHARED VALUES, NAVIGATE TRADE-OFFS, AND MAXIMIZE SYNERGIES

The challenge of navigating these and other trade-offs and achieving multiple benefits from climate adaptation and mitigation efforts underscores the importance of engaging a diversity of stakeholders throughout the national forest planning process (Thom and Seidl 2016, Janowiak et al. 2017, Seidl et al. 2017, Armatas et al. 2018, McNulty et al. 2018). The *Land Management Planning Handbook* requires forest planners to build a “collaborative process” to ensure that diverse needs

***Enhance collaboration to identify shared values, navigate trade-offs, and maximize synergies in the context of changing conditions.*** Engaging local communities and diverse constituencies as early as possible in the forest planning process helps gain buy-in and identify opportunities to minimize trade-offs and maximize synergies, including acknowledgment and discussion of the potential for fundamental changes in national forest conditions.

are addressed throughout a plan’s development and implementation, including engaging the public in efforts to “explore resolutions to one or more issues” (USFS 2015a). In addition, the development and use of a range of decision-support tools and approaches will be essential to help forest planners and managers achieve desired climate adaptation and mitigation outcomes across a range of values (e.g., Janowiak et al. 2014, Swanston et al. 2016, Ontl et al. 2020, Soto-Navarro et al. 2020).

As noted previously, decisions on whether, where, and how to manage national forests for change are ultimately values based (Backstrom et al. 2018, Dudley et al. 2018). While climate-driven changes are both ongoing and inevitable, the collaborative decision-making processes that have become a foundation for Forest Service adaptation efforts, such as Adaptation Partners and the Climate Change Response Framework, will be essential to ensure that strategies reflect evolving societal goals for forest management (Swanston et al. 2016, Halofsky et al. 2018). Indeed, the agency’s Research and Development arm offers an array of resources and services to facilitate consideration of managing for change among both national forest managers and their collaborators across the country. Initiatives like the Collaborative Forest Landscape Restoration Program (CFLRP) have helped leverage resources and build capacity to promote a range of ecological, economic, and social sustainability goals (Schultz et al. 2012). Through the CFLRP, collaborative groups are tasked with defining what “ecological restoration” means to them, create landscape-level goals, develop individual projects, monitor for effectiveness, and adaptively manage (Antuma et al. 2014).

Given the added complexities and uncertainties associated with climate change, however, bridging the research and management divide remains a challenge for public lands management (Brice et al. 2020, Carter et al. 2020). For example, several recent assessments of CFLRP efforts to date found that participants have found it difficult to define ecological objectives and develop management proposals when dealing with



*Engaging stakeholders in developing a vision for forest restoration and management is key to understanding differing social values and successfully navigating trade-offs (Nez Perce-Clearwater National Forests, Idaho). Photo: NatPar Collection/Alamy Stock Photo.*

complex ecological systems, variable conditions, and uncertainty (Antuma et al. 2014, Urgenson et al. 2017, Walpole et al. 2017). As noted by Walpole et al. (2017), restoration poses a unique challenge for collaboration given the need for participants to collectively determine which historical conditions are best to use as a reference, the degree of departure from those conditions necessary to warrant a management response, and how to balance historical conditions with expected future conditions. This is particularly challenging for systems that are not currently considered significantly degraded from historical conditions (Walpole et al. 2017).

Differing socioeconomic backgrounds and motivations for participation between agency and non-agency members have also limited the ability of some collaborative groups to develop and achieve mutually desired goals (Davis et al. 2017). In addition, it has been hard to maintain collaborative partnerships over the longer time frame necessary to effectively monitor

and assess effectiveness of many adaptation strategies due to participant turnover, including Forest Service staff (Coleman et al. 2021). To be sure, establishing easily agreed-upon goals in collaborative forest management efforts has helped build relationships, trust, and momentum for tackling more contentious issues (Walpole et al. 2017). Given that many management decisions made today will have significant implications for national forests for decades to come, further work is necessary to better operationalize collaborative planning approaches and techniques for building consensus around climate-smart restoration (McKelvey et al. 2021). Recent experience suggests that collaborative restoration efforts can navigate the challenges posed by disagreement over use of historical references by: 1) integrating information about past, present, and anticipated future conditions; 2) identifying reference periods that are analogs to current or projected future climate; and 3) balancing science-based targets with other social values in restoration planning (Urgenson et al. 2018).



*Forest restoration principles and practices will need to continue evolving to meet the challenges of accelerating climatic changes. Photo: David R./Alamy Stock Photo.*

## 8. A PATH FORWARD

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**A**s discussed throughout this report, a rapidly changing climate has significant implications for the effectiveness of restoration efforts on America’s national forests. There has been considerable progress in scientific assessment of climate change and its impact on forests, and parts of the Forest Service have strongly embraced the need to advance both climate adaptation and mitigation outcomes. The 2012 Forest Planning Rule and subsequent management directives have established an important foundation for national forest managers and their partners to adopt the collaborative, flexible management approaches necessary to effectively address climate change given inherent complexities and uncertainties. Nonetheless, varying perspectives on what climate change means for ecological restoration in practice continue to pose challenges in national forest planning and management.

There is a growing recognition that conventional restoration practices—including use of natural range of variability as a benchmark for management—must evolve to meet the emerging challenges of a changing climate. A science-based and shared understanding of what constitutes climate-smart restoration will be important to the Forest Service’s ability to achieve its mission of sustaining the wide range of ecosystem services provided to society by the nation’s national forests. Toward this end, we suggest the need for continued dialogue and collaboration between national forest managers and their partners in the state, tribal, nonprofit, and private sectors to adopt climate-smart and ecologically appropriate forest restoration and management, supported by the proposed, science-based principles highlighted in this paper. Such partnerships can build on existing collaborative initiatives and planning tools in working toward a sustainable and resilient future for America’s national forests.

## 9. REFERENCES

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- Adams, M.A. 2013. Mega-fires, tipping points and ecosystem services: Managing forests and woodlands in an uncertain future. *Forest Ecology and Management* 294: 250–261.
- Ager, A.A., R.M. Houtman, M.A. Day, C. Ringo, and P. Palaiologou. 2019. Trade-offs between US National Forest harvest targets and fuel management to reduce wildfire transmission to the wildland urban interface. *Forest Ecology and Management* 434: 99–109.
- Ager, A.A., K.C. Vogler, M.A. Day, and J.D. Bailey. 2017. Economic opportunities and trade-offs in collaborative forest landscape restoration. *Ecological Economics* 136: 226–239.
- AghaKouchak, A., D. Feldman, M. Hoerling, T. Huxman, and J. Lund. 2015. Water and climate: Recognize anthropogenic drought. *Nature* 524: 409–411.
- Aitken, S.N., and J.B. Bemmels. 2016. Time to get moving: Assisted gene flow of forest trees. *Evolutionary Applications* 9: 271–290.
- Alagona, P.S., J. Sandlos, and Y.F. Wiersma. 2012. Past imperfect: Using historical ecology and baseline data for conservation and restoration projects in North America. *Environmental Philosophy* 9: 49–70.
- Albrich, K., W. Rammer, D. Thom, and R. Seidl. 2018. Trade-offs between temporal stability and level of forest ecosystem services provisioning under climate change. *Ecological Applications* 28: 1884–1896.
- Allen, D.C., D.D. Breshears, and N.G. McDowell. 2015. On underestimation of global vulnerability to tree mortality and forest die-off from hotter drought in the Anthropocene. *Ecosphere* 6: 1–55.
- Anderegg, L.D.L., W.R.L. Anderegg, and J.A. Berry. 2013. Not all droughts are created equal: Translating meteorological drought into woody plant mortality. *Tree Physiology* 33: 701–712.
- Anderegg, W.R.L., A. Fling, C.-Y. Huang, et al. 2015. Tree mortality predicted from drought-induced vascular damage. *Nature Geoscience* 8: 367–371.
- Anderegg, W.R.L., L.D.L. Anderegg, K.L. Kerr, and A.T. Trugman. 2019. Widespread drought-induced tree mortality at dry range edges indicates that climate stress exceeds species' compensating mechanisms. *Global Change Biology* 25: 3793–3802.
- Anderegg, W.R.L., A.T. Trugman, G. Badgley, et al. 2020. Climate-driven risks to the climate mitigation potential of forests. *Science* 368: eaaz7005.
- Anderson, C.M., R.S. DeFries, R. Litterman, et al. 2019. Natural climate solutions are not enough. *Science* 363: 933–934.
- Anderson, T.R., E. Hawkins, and P.D. Jones. 2016. CO<sub>2</sub>, the greenhouse effect and global warming: From the pioneering work of Arrhenius and Callendar to today's Earth System Models. *Endeavour* 40: 178–187.
- Anderson-Teixeira, K.J., A.D. Miller, J.E. Mohan, et al. 2013. Altered dynamics of forest recovery under a changing climate. *Global Change Biology* 19: 2001–2021.
- Antuma, J., B. Esch, B. Hall, E. Munn, and F. Sturges. 2014. Restoring forests and communities: Lessons from the Collaborative Forest Landscape Restoration Program. Ann Arbor: School of Natural Resources and Environment, University of Michigan.
- Archie, K.M., L. Dilling, J.B. Milford, and F.C. Pampel. 2014. Unpacking the 'information barrier': Comparing perspectives on information as a barrier to climate change adaptation in the interior mountain West. *Journal of Environmental Management* 133: 397–410.
- Armatas, C.A., R.M. Campbell, A.E. Watson, et al. 2018. An integrated approach to valuation and trade-off analysis of ecosystem services for National Forest decision-making. *Ecosystem Services* 33: 1–18.
- Asbjornsen, H., J.L. Campbell, A.W. D'Amato, et al. 2019. Managing effects of drought in the Midwest and Northeast United States. p 165–190. In: J.M. Vose et al., eds. *Effects of Drought on Forests and Rangelands in the United States: Translating Science into Management Responses*. General Technical Report WO-98. Washington, DC: U.S. Department of Agriculture, Forest Service.
- ASCC (Adaptive Silviculture for Climate Change). No date. *Silviculture and climate adaptation*. Houghton, MI: U.S. Department of Agriculture, Forest Service, Northern Research Station (accessed May 20, 2021). <https://www.adaptivesilviculture.org/node/987>
- Ault, T.R., A.K. Macalady, G.T. Pederson, J.L. Betancourt, and M.D. Schwartz. 2011. Northern Hemisphere modes of variability and the timing of spring in western North America. *Journal of Climate* 24: 4003–4014.
- Backstrom, A.C., G.E. Gerrard, R.J. Hobbs, and S.A. Bekessy. 2018. Grappling with the social dimensions of novel ecosystems. *Frontiers in Ecology and the Environment* 16: 109–117.
- Baker, J.S., C.M. Wade, B.L. Sohngen, S. Ohrel, and A.A. Fawcett. 2019. Potential complementarity between forest carbon sequestration incentives and biomass energy expansion. *Energy Policy* 126: 391–401.

- Barbero, R., J.T. Abatzoglou, N.K. Larkin, C.A. Kolden, and B. Stocks. 2015. Climate change represents increased potential for very large fires in the contiguous United States. *International Journal of Wildland Fire* 24: 892–899.
- Barrett, S.W., T. DeMeo, J.L. Jones, J.D. Zeiler, and L.C. Hutter. 2006. Assessing ecological departure from reference conditions with the Fire Regime Condition Class (FRCC) mapping tool. p 575–585. In: P.L. Andrews and B.W. Butler, comps. *Fuels Management—How to Measure Success: Conference Proceedings*. 28–30 March 2006; Portland, OR. Proceedings RMRS-P-41. Fort Collins, CO: U.S. Department of Agriculture, Forest Service, Rocky Mountain Research Station.
- Barros, A.M., A.A. Ager, M.A. Day, M.A. Krawchuck, and T.A. Spies. 2018. Wildfires managed for restoration enhance ecological resilience. *Ecosphere* 9: e02161.
- Bartowitz, K.J., P.E. Higuera, B.N. Schuman, K.K. McLaughlan, and T.W. Hudiburg. 2019. Post-fire carbon dynamics in subalpine forests of the Rocky Mountains. *Fire* 2: 58.
- Batllori, E., F. Lloret, T. Aakala, et al. 2020. Forest and woodland replacement patterns following drought-related mortality. *Proceedings of the National Academy of Sciences* 117: 29720–29729. doi: 10.1073/pnas.2002314117
- Battaglia, M.A., B. Gannon, P.M. Brown, et al. 2018. Changes in forest structure since 1860 in ponderosa pine dominated forests in the Colorado and Wyoming Front Range, USA. *Forest Ecology and Management* 422: 147–160.
- Bell, D.M., J.B. Bradford, and W.K. Lauenroth. 2014. Early indicators of change: Divergent climate envelopes between tree life stages imply range shifts in the western United States. *Global Ecology and Biogeography* 23: 168–180.
- Beller, E.E., L. McClenachan, E.S. Zavaleta, and L.G. Larsen. 2020. Past forward: Recommendations from historical ecology for ecosystem management. *Global Ecology and Conservation* 21: e00836.
- Belmecheri, S., R.S. Maxwell, A.H. Taylor, et al. 2021. Precipitation alters the CO<sub>2</sub> effect on water-use efficiency of temperate forests. *Global Change Biology* 27: 1560–1571.
- Bennett, E., G.D. Peterson, and L.J. Gordon. 2009. Understanding relationships among multiple ecosystem services. *Ecology Letters* 12: 1394–1404.
- Benson, M.H., and A.S. Garmestani. 2011. Can we manage for resilience? The integration of resilience thinking into natural resource management in the United States. *Environmental Management* 48: 392–399.
- Berner, L.T., B.E. Law, A.J. Meddens, and J.A. Hicke. 2017. Tree mortality from fires, bark beetles, and timber harvest during a hot and dry decade in the western United States (2003–2012). *Environmental Research Letters* 12: 065005.
- Birdsey, R.A., A.J. Dugan, S.P. Healy, et al. 2019. Assessment of the Influence of Disturbance, Management Activities, and Environmental Factors on Carbon Stocks of U.S. National Forests. General Technical Report RMRS-GTR-402. Fort Collins, CO: U.S. Department of Agriculture, Forest Service, Rocky Mountain Research Station.
- Blate, G.M., L.A. Joyce, J.S. Littell, et al. 2009. Adapting to climate change in United States National Forests. *Unasylva* 231/232 60: 57–62.
- BLM (Bureau of Land Management). 2018. Oil and gas statistics. Washington, DC: Department of the Interior, Bureau of Land Management (accessed April 19, 2021) <https://www.blm.gov/programs-energy-and-minerals-oil-and-gas-oil-and-gas-statistics>
- Boisvenue, C., and S.W. Running. 2010. Simulations show decreasing carbon stocks and potential for carbon emissions in Rocky Mountain forests over the next century. *Ecological Applications* 20: 1302–1319.
- Boisvert-Marsh, L., C. Périé, and S. de Blois. 2019. Divergent responses to climate change and disturbance drive recruitment patterns underlying latitudinal shifts of tree species. *Journal of Ecology* 107: 1956–1969.
- Bone, C., C. Moseley, K. Vinyeta, and R.P. Bixler. 2016. Employing resilience in the United States Forest Service. *Land Use Policy* 52: 430–438.
- Botkin, D.B. 2014. Adapting forest science, practice, and policy to shifting ground: From steady-state assumptions to dynamic change. p 35–46. In: V.A. Sample and R.P. Bixler, eds. *Forest Conservation and Management in the Anthropocene: Conference Proceedings*. Proceedings RMRS-P-71. Fort Collins, CO: U.S. Department of Agriculture, Forest Service, Rocky Mountain Research Station.
- Bottero, A., A.W. D'Amato, B.J. Palik, et al. 2017. Density-dependent vulnerability of forest ecosystems to drought. *Journal of Applied Ecology* 54: 1605–1614.
- Bowerman, T.E., M.L. Keefer, and C.C. Caudhill. 2021. Elevated stream temperature, origin, and individual size influence Chinook salmon prespawning mortality across the Columbia River Basin. *Fisheries Research* 237: 105874.
- Bradford, J.B., J.L. Betancourt, B.J. Butterfield, S.M. Munson, and T.E. Wood. 2018. Anticipatory natural resource science and management for a changing future. *Frontiers in Ecology and the Environment* 16: 295–303.

- Bradley, B.A. 2009. Regional analysis of the impacts of climate change on cheatgrass invasion shows potential risk and opportunity. *Global Change Biology* 15: 196–208.
- Brand, F., and K. Jax. 2007. Focusing the meaning(s) of resilience: Resilience as a descriptive concept and a boundary object. *Ecology and Society* 12: 23.
- Brandt, L.A., P.R. Butler, S.D. Handler, et al. 2017. Integrating science and management to assess forest ecosystem vulnerability to climate change. *Journal of Forestry* 115: 212–221.
- Brantley, S.T., J.M. Vose, D.N. Wear, and L. Band. 2018. Planning for an uncertain future: Restoration to mitigate water scarcity and sustain carbon sequestration. p 291–309. In: L.K. Kirkman and S.B. Jack, eds. *Ecological Restoration and Management of Longleaf Pine Forests*. Boca Raton, FL: CRC Press.
- Breed, M.F., P.A. Harrison, A. Bischoff, et al. 2018. Priority actions to improve provenance decision-making. *BioScience* 68: 510–516.
- Breshears, D.D., N.S. Cobb, P.M. Rich, et al. 2005. Regional vegetation die-off in response to global-change-type drought. *Proceedings of the National Academy of Sciences* 102: 15144–15148.
- Brice, E.M., B.A. Miller, H. Zhang, et al. 2020. Impacts of climate change on multiple use management of Bureau of Land Management land in the Intermountain West, USA. *Ecosphere* 11: e03286.
- Brooks, R.T. 2009. Potential impacts of global climate change on the hydrology and ecology of ephemeral freshwater systems of the forests of the northeastern United States. *Climatic Change* 95: 469–483.
- Bryant, T., K.M. Waring, A. Sánchez Meador, and J. Bradford. 2019. A framework for quantifying resilience to forest disturbance. *Frontiers in Forests and Global Change* 2: 56.
- Buford, M., C. Chan, J. Crockett, E. Reinhardt, and J. Sloan. 2015. *From Accelerating Restoration to Creating and Maintaining Resilient Landscapes and Communities Across the Nation: Update on Progress from 2012*. FS-1069. Washington, DC: U.S. Department of Agriculture, Forest Service.
- Buotte, P.C., B.E. Law, W.J. Ripple, and L.T. Berner. 2019a. Carbon sequestration and biodiversity co-benefits of preserving forests in the western United States. *Ecological Applications* 30: e02039.
- Buotte, P.C., S. Levis, B.E. Law, et al. 2019b. Near-future forest vulnerability to drought and fire varies across the western United States. *Global Change Biology* 25: 290–303.
- Busby, S.U., K.B. Moffett, and A. Holz. 2020. High-severity and short-interval wildfires limit forest recovery in the Central Cascade Range. *Ecosphere* 11: e03247.
- Butler, W.H. 2013. Collaboration at arm's length: Navigating agency engagement in landscape-scale ecological restoration collaboratives. *Journal of Forestry* 111: 395–403.
- Butterfield, B.J., S.M. Copeland, S.M. Munson, C.M. Roybal, and T.E. Wood. 2017. Prestoration: Using species in restoration that will persist now and into the future. *Restoration Ecology* 25: S155–S163.
- Cadmus, R., F.J. Escobedo, D. McLaughlin, and A. Abd-Elrahman. 2014. Analyzing trade-offs, synergies, and drivers among timber production, carbon sequestration, and water yield in *Pinus elliotii* forests in southeastern USA. *Forests* 5: 1409–1431.
- Cannon, J.B., B.M. Gannon, Z. Wurtzebach, and A.S. Cheng. 2019. Report on Potential Application of Landscape-scale Analyses for Assistance with Forest Planning. Fort Collins: Colorado Forest Restoration Initiative, Colorado State University.
- Cantarello, E., A.C. Newton, P.A. Martin, et al. 2017. Quantifying resilience of multiple ecosystem services and biodiversity in a temperate forest landscape. *Ecology and Evolution* 7: 9661–9675.
- Carpenter, S., B. Walker, J.M. Anderies, and N. Abel. 2001. From metaphor to measurement: Resilience of what to what? *Ecosystems* 4: 765–781.
- Carter, S.K., E. Fleishman, I.I. Leinwand, et al. 2019. Quantifying ecological integrity to inform management of multiple-use public lands in the United States. *Environmental Management* 64: 1–19.
- Carter, S.K., D.S. Pilliod, T. Haby, et al. 2020. Bridging the research-management gap: Landscape science in practice on public lands in the western United States. *Landscape Ecology* 35: 1–16.
- Cartwright, J. 2018. Landscape topoedaphic features create refugia from drought and insect disturbance in a lodgepole and whitebark pine forest. *Forests* 9: 715. doi: 10.3390/f9110715
- Case, J.L., L.T. Wood, J.L. Blaes, et al. 2021a. Soil moisture responses associated with significant tropical cyclone rainfall events. *Journal of Operational Meteorology* 9(1): 1–17.
- Case, M.J., B.G. Johnson, K.J. Bartowitz, and T.W. Hudiburg. 2021b. Forests of the future: Climate change impacts and implications for carbon storage in the Pacific Northwest, USA. *Forest Ecology and Management* 482: 118886.
- Cattau, M.E., C. Wessman, A. Mahood, and J.K. Balch. 2020. Anthropogenic and lightning-started fires are becoming larger and more frequent over a longer season length in the USA. *Global Ecology and Biogeography* 29: 668–681.
- CCRF (Climate Change Response Framework). No date. *Adaptation Workbook: A climate change tool for land management and conservation* (accessed May 25, 2021). <https://adaptationworkbook.org/>

- Chapin, F.S. III, A.D. McGuire, R.W. Ruess, et al. 2010. Resilience of Alaska's boreal forest to climatic change. *Canadian Journal of Forest Resources* 40: 1360–1370.
- Chapin, J., and J. Abrams. 2020. *Incorporating Resilience in National Forest Planning and Management: A Quick Guide*. Eugene, OR: Ecosystem Workforce Program, Institute for a Sustainable Environment.
- Charnley, S., T.A. Spies, A.M. Barros, E.M. White, and K.A. Olsen. 2017. Diversity in forest management to reduce wildfire losses: Implications for resilience. *Ecology and Society* 22: 22.
- Chiang, F., O. Mazdiyasn, and A. AghaKouchak. 2018. Amplified warming of droughts in southern United States in observations and model simulations. *Science Advances* 4: eaat2380.
- Cho, Y., D. Lee, and S. Bae. 2017. Effects of vegetation structure and human impact on understory honey plant richness: Implications for pollinator visitation. *Journal of Ecology and Environment* 41: 2.
- Clark, J.S., A.E. Gelfand, C.W. Woodall, and K. Zhu. 2014. More than the sum of the parts: Forest climate response from joint species distribution models. *Ecological Applications* 24: 990–999.
- Clewell, A., J. Aronson, and K. Winterhalder. 2002. *The SER Primer on Ecological Restoration*. Society for Ecological Restoration Science and Policy Working Group. Washington, DC: Society for Ecological Restoration.
- Clifford, K.R., L. Yung, W.R. Travis, et al. 2020. Navigating climate adaptation on public lands: How views on ecosystem change and scale interact with management approaches. *Environmental Management* 66: 614–628.
- Coleman, K.J., W.H. Butler, M.J. Stern, and S.L. Beck. 2021. "They're constantly cycling through": Lessons about turnover and collaborative forest planning. *Journal of Forestry* 119: 1–12.
- Colloff, M.J., M.D. Doherty, S. Lavorel, et al. 2016. Adaptation services and pathways for the management of temperate montane forests under transformational climate change. *Climatic Change* 138: 267–282.
- Colloff, M.J., S. Lavorel, L.E. van Kerkhoff, et al. 2017a. Transforming conservation science and practice for a postnormal world. *Conservation Biology* 31: 1008–1017.
- Colloff, M.J., B. Martín-López, S. Lavorel, et al. 2017b. An integrative research framework for enabling transformative adaptation. *Environmental Science and Policy* 68: 87–96.
- Colombaroli, D., C. Whitlock, W. Tinner, and M. Conedera. 2017. Paleo records as a guide for ecosystem management and biodiversity conservation. *Past Global Changes Magazine* 25: 78–79.
- Contosta, A.R., N.J. Casson, S. Garlick, et al. 2019. Northern forest winters have lost cold, snowy conditions that are important for ecosystems and human communities. *Ecological Applications* 29: e01974.
- Coop, J.D., S.A. Parks, C.S. Stevens-Rumann, et al. 2020. Wildfire-driven forest conversion in western North American landscapes. *BioScience* 70: 659–673. doi: 10.1093/biosci/biaa061
- Coughlan, M.R., A. Ellison, J. Abrams, and H. Huber-Stearns. 2020. *Land Manager Experiences with Resilience in National Forest Planning and Management*. Eugene, OR: Ecosystem Workforce Program, Institute for a Sustainable Environment.
- Coulston, J.W., D.N. Wear, and J.M. Vose. 2015. Complex forest dynamics indicate potential for slowing carbon accumulation in the southeastern United States. *Scientific Reports* 5: 8002.
- Craig, R.K. 2010. Stationarity is dead—long live transformation: Five principles for climate change adaptation law. *Harvard Environmental Law Review* 34: 9–75.
- Crausbay, S.D., J. Betancourt, J. Bradford, et al. 2020. Unfamiliar territory: Emerging themes for ecological drought research and management. *One Earth* 3: 337–353.
- Crausbay, S.D., A.R. Ramirez, S.L. Carter, et al. 2017. Defining ecological drought for the twenty-first century. *Bulletin of the American Meteorological Society* 98: 2543–2550.
- Creutzburg, M.S., R.M. Scheller, M.S. Lucash, S.D. LeDuc, and M.G. Johnson. 2017. Forest management scenarios in a changing climate: Trade-offs between carbon, timber, and old forest. *Ecological Applications* 27: 503–518.
- Crimmins, S.M., S.Z. Dobrowski, J.A. Greenberg, J.T. Abatzoglou, and A.R. Mynsberge. 2011. Changes in climatic water balance drive downhill shifts in plant species' optimum elevations. *Science* 331: 324–327.
- Crotteau, J.S., and C.R. Keyes. 2020. Restoration treatments improve overstory tree resistance attributes and growth in a ponderosa pine/Douglas-fir forest. *Forests* 11: 574.
- Crow, T.R. 2014. Functional restoration: From concept to practice. *Journal of Sustainable Forestry* 33: S3–S14.
- Crozier, L.G., J.E. Siegel, L.E. Wiesebron, et al. 2020. Snake River sockeye and Chinook salmon in a changing climate: Implications for upstream migration survival during recent extreme and future climates. *PLoS One* 15: e0238886.
- D'Amato, A.W., J.B. Bradford, S. Fraver, and B.J. Palik. 2011. Forest management for mitigation and adaptation to climate change: Insights from long-term silviculture experiments. *Forest Ecology and Management* 262: 803–816.

- D'Amato, A.W., E.J. Jokela, K.L. O'Hara, and J.N. Long. 2018. Silviculture in the United States: An amazing period of change over the past 30 years. *Journal of Forestry* 116: 55–67.
- D'Amato, A.W., and B.J. Palik. 2021. Building on the last “new” thing: Exploring the compatibility of ecological and adaptation silviculture. *Canadian Journal of Forest Research* 51: 172–180.
- D'Amato, A.W., B.J. Palik, J.F. Franklin, and D.R. Foster. 2017. Exploring the origins of ecological forestry in North America. *Journal of Forestry* 115: 126–127.
- Davis, E.J., E.M. White, L.K. Cervený, et al. 2017. Comparison of USDA Forest Service and stakeholder motivations and experiences in collaborative federal forest governance in the western United States. *Environmental Management* 60: 908–921.
- Davis, J., and K. Weber. 2018. Spatio-Temporal Relationships of Historic Wildfires: Using the NASA RECOVER Historic Fires Geodatabase to Perform Long-term Analysis of Wildfire Occurrences in the Western United States. Pocatello: Idaho State University GIS Training and Research Center.
- Davis, K.T., S.Z. Dobrowski, P.E. Higuera, et al. 2019. Wildfires and climate change push low-elevation forests across a critical climate threshold for tree regeneration. *Proceedings of the National Academy of Sciences* 116: 6193–6198.
- Davison, J.E., S. Coe, D. Finch, et al. 2012. Bringing indices of species vulnerability to climate change into geographic space: An assessment across Coronado National Forest. *Biodiversity and Conservation* 21: 189–204.
- Deal, R., L. Fong, E. Phelps, et al. 2017a. Integrating Ecosystem Services into National Forest Service Policy and Operations. General Technical Report PNW-GTR-943. Portland, OR: U.S. Department of Agriculture, Forest Service, Pacific Northwest Research Station.
- Deal, R.L. 2020. Integrating ecosystem services into sustainable forest management of public lands. p 83–93. In: L.S. Pile et al., comps. *Forest Management–Research Partnerships: Proceedings of the 2019 National Silviculture Workshop*. General Technical Report NRS-P-193. Madison, WI: U.S. Department of Agriculture, Forest Service, Northern Research Station.
- Deal, R.L., N. Smith, and J. Gates. 2017b. Ecosystem services to enhance sustainable forest management in the US: Moving from forest service national programmes to local projects in the Pacific Northwest. *Forestry: An International Journal of Forest Research* 90: 632–639.
- DellaSala, D.A., J.R. Karr, and D.M. Olson. 2011. Roadless areas and clean water. *Journal of Soil and Water Conservation* 66: 78A–84A.
- DeLucia, E.H., P.D. Nabity, J.A. Zavala, and M.R. Berenbaum. 2012. Climate change: Resetting plant-insect interactions. *Plant Physiology* 160: 1677–1685.
- DeMeo, T., R. Haugo, C. Ringo, et al. 2018. Expanding our understanding of forest structural restoration needs in the Pacific Northwest. *Northwest Science* 92: 18–35.
- DeRose, R.J., and J.N. Long. 2014. Resistance and resilience: A conceptual framework for silviculture. *Forest Science* 60: 1205–1212.
- DeSoto, L., M. Cailleret, F. Sterck, et al. 2020. Low growth resilience to drought is related to future mortality risk in trees. *Nature Communications* 11: 1–9.
- Dey, D.C., B.O. Knapp, M.A. Battaglia, et al. 2019. Barriers to natural regeneration in temperate forests across the USA. *New Forests* 50: 11–40.
- Diffenbaugh, N.S., D.L. Swain, and D. Touma. 2015. Anthropogenic warming has increased drought risk in California. *Proceedings of the National Academy of Sciences* 112: 3931–3936.
- Dixit, A., T. Kolb, and O. Burney. 2020. Provenance geographical and climate characteristics influence budburst phenology of southwestern ponderosa pine seedlings. *Forests* 11: 1067.
- DOI and USDA (U.S. Department of the Interior and U.S. Department of Agriculture). 2014. National cohesive wildland fire management strategy. DOI and USDA (accessed April 1, 2021). <https://www.forestsandrangelands.gov/strategy/>
- Domke, G.M., S.N. Oswalt, B.F. Walters, and R.S. Morin. 2020. Tree planting has the potential to increase carbon sequestration capacity of forests in the United States. *Proceedings of the National Academy of Sciences* 117: 24649–24651.
- Donato, D.C., J.S. Halofsky, and M.J. Reilly. 2020. Corraling a black swan: Natural range of variation in a forest landscape driven by rare, extreme events. *Ecological Applications* 30: e02013.
- Donovan, V.M., C.P. Roberts, C.L. Wonkka, D.A. Wedin, and D. Twidwell. 2019. Ponderosa pine regeneration, wildland fuels management, and habitat conservation: Identifying trade-offs following wildfire. *Forests* 10: 286.
- Druckenbrod, D.L., D. Martin-Benito, D.A. Orwig, et al. 2019. Redefining temperate forest responses to climate and disturbance in the eastern United States: New insights at the mesoscale. *Global Ecology and Biogeography* 28: 557–575.
- Duan, K., G. Sun, S. Sun, et al. 2016. Divergence of ecosystem services in US National Forests and Grasslands under a changing climate. *Scientific Reports* 6: 24441.

- Dudley, M.P., M. Freeman, S. Wenger, C.R. Jackson, and C.M. Pringle. 2020. Rethinking foundation species in a changing world: The case for *Rhododendron maximum* as an emerging foundation species in shifting ecosystems of the southern Appalachians. *Forest Ecology and Management* 472: 118240.
- Dudney, J., R.J. Hobbs, R. Heilmayr, J.J. Battles, and K.N. Suding. 2018. Navigating novelty and risk in resilience management. *Trends in Ecology and Evolution* 33: 863–873.
- Dugan, A.J., R. Birdsey, S.P. Healey, et al. 2017. Forest sector carbon analyses support land management planning and projects: Assessing the influence of anthropogenic and natural factors. *Climatic Change* 144: 207–220.
- Dugan, A.J., R. Birdsey, V.S. Mascorro, et al. 2018. A systems approach to assess climate change mitigation options in the landscapes of the United States forest sector. *Carbon Balance and Management* 13: 13.
- Duinker, P.N. 1990. Climate change and forest management, policy, and land use. *Land Use Policy* 7: 124–137.
- Dukes, J.S., J. Pontius, D. Orwig, et al. 2009. Responses of insect pests, pathogens, and invasive plant species to climate change in forests of northeastern North America: What can we predict? *Canadian Journal of Forest Research* 39: 231–248.
- Dumroese, R.K., N. Balloffet, J.W. Crockett, J.A. Stanturf, and L.E. Nave. 2019. A national approach to leverage the benefits of tree planting on public lands. *New Forests* 50: 1–9.
- Dumroese, R.K., M.I. Williams, J.A. Stanturf, and J.B.S. Clair. 2015. Considerations for restoring temperate forests of tomorrow: Forest restoration, assisted migration, and bioengineering. *New Forests* 46: 947–964.
- Duncan, S.L., B.C. McComb, and K.N. Johnson. 2010. Integrating ecological and social ranges of variability in conservation of biodiversity: Past, present, and future. *Ecology and Society* 15: 5.
- Dunn, C., and M.P. Thompson. 2018. The fire next time: Strategies for fire and fuels management. p 359–385. In: J.F. Franklin et al., eds. *Ecological Forest Management*. Long Grove, IL: Waveland Press, Inc.
- Duveneck, M.J., and R.M. Scheller. 2016. Measuring and managing resistance and resilience under climate change in northern Great Lakes forests (USA). *Landscape Ecology* 31: 669–686.
- Dwire, K.A., S. Mellmann-Brown, and J.T. Gurrieri. 2018. Potential effects of climate change on riparian areas, wetlands, and groundwater-dependent ecosystems in the Blue Mountains, Oregon, USA. *Climate Services* 10: 44–52.
- Dyballa, K.E., K. Steger, R.G. Walsh, et al. 2019. Optimizing carbon storage and biodiversity co-benefits in reforested riparian zones. *Journal of Applied Ecology* 56: 343–353.
- Easterling, D.R., K.E. Kunkel, J.R. Arnold, et al. 2017. Precipitation change in the United States. p 207–230. In: D.J. Wuebbles et al., eds. *Climate Science Special Report: Fourth National Climate Assessment*, vol. I. Washington, DC: U.S. Global Change Research Program.
- Elias, E., D. James, S. Heimel, et al. 2021. Implications of observed changes in high mountain snow storage, snowmelt timing and melt window. *Journal of Hydrology: Regional Studies* 35: 100799.
- Ellenwood, M.S., L. Dilling, and J.B. Milford. 2012. Managing United States public lands in response to climate change: A view from the ground up. *Environmental Management* 49: 954–967.
- Estiarte, M., and J. Peñuelas. 2015. Alteration of the phenology of leaf senescence and fall in winter deciduous species by climate change: Effects on nutrient proficiency. *Global Change Biology* 21: 1005–1017.
- Etterson, J.R., M.W. Cornett, M.A. White, and L.C. Kavajecz. 2020. Assisted migration across fixed seed zones detects adaptation lags in two major North American tree species. *Ecological Applications*: e02092.
- Evans, D.M., C.E. Zipper, J.A. Burger, B.D. Strahm, and A.M. Villamagna. 2013. Reforestation practice for enhancement of ecosystem services on a compacted surface mine: Path toward ecosystem recovery. *Ecological Engineering* 51: 16–23.
- Fahey, T.J., P.B. Woodbury, J.J. Battles, et al. 2010. Forest carbon storage: Ecology, management, and policy. *Frontiers in Ecology and the Environment* 8: 245–252.
- Falk, D.A., A.C. Watts, and A.E. Thode. 2019. Scaling ecological resilience. *Frontiers in Ecology and Evolution* 7: 275.
- Fang, J., J.A. Lutz, L. Wang, H.H. Shugart, and X. Yan. 2020. Using climate-driven leaf phenology and growth to improve predictions of gross primary productivity in North American forests. *Global Change Biology* 26: 6974–6988. doi: 10.1111/gcb.15349
- Fargione, J.E., S. Bassett, T. Boucher, et al. 2018. Natural climate solutions for the United States. *Science Advances* 4: eaat1869.
- Fei, S., J.M. Desprez, K.M. Potter, et al. 2017. Divergence of species responses to climate change. *Science Advances* 3: e1603055.
- Fernández-de-Uña, L., N.G. McDowell, I. Cañellas, and G. Gea-Izquierdo. 2016. Disentangling the effect of competition, CO<sub>2</sub>, and climate on intrinsic water-use efficiency and tree growth. *Journal of Ecology* 104: 678–690.
- Fischelli, N.A., S.R. Abella, M. Peters, and F.J. Krist Jr. 2014. Climate, trees, pests, and weeds: Change, uncertainty, and biotic stressors in eastern US national park forests. *Forest Ecology and Management* 327: 31–39.

- Fisichelli, N.A., G.W. Schuurman, and C. Hawkins-Hoffman. 2016. Is 'resilience' maladaptive? Toward an accurate lexicon for climate change adaptation. *Environmental Management* 57: 753–758.
- Ford, K.R., I.K. Breckheimer, J.F. Franklin, et al. 2017. Competition alters tree growth responses to climate at individual and stand scales. *Canadian Journal of Forest Research* 47: 53–62.
- Foster, A.C., A.H. Armstrong, J.K. Shuman, et al. 2019. Importance of tree- and species-level interactions with wildfire, climate, and soils in interior Alaska: Implications for forest change under a warming climate. *Ecological Modelling* 409: 108765.
- Franklin, J.F., K.N. Johnson, and D.L. Johnson, eds. 2018. *Ecological Forest Management*. Long Grove, IL: Waveland Press, Inc.
- Frelich, L.E., K. Jögiste, J. Stanturf, A. Jansons, and F. Vodde. 2020. Are secondary forests ready for climate change? It depends on magnitude of climate change, landscape diversity and ecosystem legacies. *Forests* 11: 965.
- Friedlingstein, P. 2015. Carbon cycle feedbacks and future climate change. *Philosophical Transactions of the Royal Society A: Mathematical, Physical and Engineering Sciences* 373: p.20140421
- Friggens, M., S. Mueller, and M. Williams. 2020. Using science management partnerships to develop landscape level indicators and assessments to measure vulnerability of piñon-juniper woodlands. *Ecological Indicators* 119: 106830.
- Funk, J.M., N. Aguilar-Amuchastegui, W. Baldwin-Cantello, et al. 2019. Securing the climate benefits of stable forests. *Climate Policy* 19: 845–860.
- Furniss, M.J., K.B. Roby, D. Cendrelli, et al. 2013. *Assessing the Vulnerability of Watersheds to Climate Change: Results of National Forest Watershed Vulnerability Pilot Assessments*. General Technical Report PNW-GTR-884. Portland, OR: U.S. Department of Agriculture, Forest Service, Pacific Northwest Research Station.
- Fyfe, J.C., C. Derksen, L. Mudryk, et al. 2017. Large near-term projected snowpack loss over the western United States. *Nature Communications* 8: 1–7.
- Gandhi, K.J., F. Campbell, and J. Abrams. 2019. Current status of forest health policy in the United States. *Insects* 10: 106.
- Gang, C., S. Pan, H. Tian, et al. 2020. Satellite observations of forest resilience to hurricanes along the northern Gulf of Mexico. *Forest Ecology and Management* 472: 118243.
- Gann, G.D., T. McDonald, B. Walder, et al. 2019. *International principles and standards for the practice of ecological restoration*. Second edition. *Restoration Ecology* 27: S1–S46.
- Gergel, D.R., B. Nijssen, J.T. Abatzoglou, D.P. Lettenmaier, and M.R. Stumbaugh. 2017. Effects of climate change on snowpack and fire potential in the western USA. *Climatic Change* 141: 287–299.
- Gill, A.L., A.S. Gallinat, R. Sanders-DeMott, et al. 2015. Changes in autumn senescence in northern hemisphere deciduous trees: A meta-analysis of autumn phenology studies. *Annals of Botany* 116: 875–888.
- Gilliam, F.S. 2007. The ecological significance of the herbaceous layer in temperate forest ecosystems. *BioScience* 57: 845–858.
- Gillson, L., C. Whitlock, and G. Humphrey. 2019. Resilience and fire management in the Anthropocene. *Ecology and Society* 24: 14.
- Golladay, S.W., K.L. Martin, J.M. Vose, et al. 2016. Achievable future conditions as a framework for guiding forest conservation and management. *Forest Ecology and Management* 360: 80–96.
- Goss, M., D.L. Swain, J.T. Abatzoglou, et al. 2020. Climate change is increasing the risk of extreme autumn wildfire conditions across California. *Environmental Research Letters* 15: 094016.
- Greiner, S.M., K.E. Grimm, and A.E.M. Waltz. 2020. Managing for resilience? Examining management implications of resilience in southwestern National Forests. *Journal of Forestry*: 1–11.
- Griscom, B.W., J. Adams, P.W. Ellis, et al. 2017. Natural climate solutions. *Proceedings of the National Academy of Sciences* 114: 11645–11650.
- Gustafson, E.J., C.C. Kern, B.R. Miranda, et al. 2020. Climate adaptive silviculture strategies: How do they impact growth, yield, diversity and value in forested landscapes? *Forest Ecology and Management* 470: 118208.
- Haffey, C., T.D. Sisk, C.D. Allen, A.E. Thode, and E.Q. Margolis. 2018. Limits to ponderosa pine regeneration following large high-severity forest fires in the United States Southwest. *Fire Ecology* 14: 143–163.
- Hagerman, S.M., and R. Pelai. 2018. Responding to climate change in forest management: Two decades of recommendations. *Frontiers in Ecology and the Environment* 16: 579–587.
- Hall, J., R. Muscarella, A. Quebbeman, et al. 2020. Hurricane-induced rainfall is a stronger predictor of tropical forest damage in Puerto Rico than maximum wind speeds. *Scientific Reports* 10: 4318.
- Halofsky, J.E., S.A. Andrews-Key, J.E. Edwards, et al. 2018. Adapting forest management to climate change: The state of science and applications in Canada and the United States. *Forest Ecology and Management* 421: 84–97.

- Halofsky, J.E., and D.L. Peterson. 2016. Climate change vulnerabilities and adaptation options for forest vegetation management in the northwestern USA. *Atmosphere* 7: 46.
- Halofsky, J.E., D.L. Peterson, and B.J. Harvey. 2020. Changing wildfire, changing forests: The effects of climate change on fire regimes and vegetation in the Pacific Northwest, USA. *Fire Ecology* 16: 4.
- Halofsky, J.E., T.W. Warziniack, D.L. Peterson, and J.J. Ho. 2017. Understanding and managing the effects of climate change on ecosystem services in the Rocky Mountains. *Mountain Research and Development* 37: 340–352.
- Hanberry, B.B., R.F. Noss, H.D. Safford, S.K. Allison, and D.C. Dey. 2015. Restoration is preparation for the future. *Journal of Forestry* 113: 425–429.
- Handler, S., C. Pike, B. St. Clair, H. Abbotts, and M. Janowiak. 2018. Assisted migration. Washington, DC: U.S. Department of Agriculture, Forest Service, Climate Change Resource Center (accessed April 1, 2021). <https://www.fs.usda.gov/ccrc/topics/assisted-migration>
- Hansen, W.D., R. Fitzsimmons, J. Olnes, and A.P. Williams. 2020. An alternate vegetation type proves resilient and persists for decades following forest conversion in the North American boreal biome. *Journal of Ecology* 109: 85–98.
- Hanson, P.J., and J.F. Weltzin. 2000. Drought disturbance from climate change: Response of United States forests. *Science of the Total Environment* 262: 205–220.
- Harrington, C.A., and P.J. Gould. 2015. Tradeoffs between chilling and forcing in satisfying dormancy requirements for Pacific Northwest tree species. *Frontiers in Plant Science* 6: 120.
- Harris, L.B., and A.H. Taylor. 2020. Rain-shadow forest margins resilient to low-severity fire and climate change but not high-severity fire. *Ecosphere* 11: e03258.
- Harrison, J.L., R. Sanders-DeMott, A.B. Reinmann, et al. 2020. Growing-season warming and winter soil freeze/thaw cycles increase transpiration in a northern hardwood forest. *Ecology* 101: e03173.
- Hart, J.L., M.L. Buchanan, and L.E. Cox. 2015. Has forest restoration been freed from the bonds of history? *Journal of Forestry* 113: 429–430.
- Harvey, J.E., M. Smiljanić, T. Scharnweber, et al. 2020. Tree growth influenced by warming winter climate and summer moisture availability in northern temperate forests. *Global Change Biology* 26: 2505–2518.
- Hatfield, J., C. Swanston, M. Janowiak, et al. 2015. Midwest and Northern Forests Regional Climate Hub Assessment of Climate Change Vulnerability and Adaptation and Mitigation Strategies. Ames, IA: U.S. Department of Agriculture, Forest Service, Midwest Hub.
- Haugo, R., C. Zanger, T. DeMeo, et al. 2015. A new approach to evaluate forest structure restoration needs across Oregon and Washington, USA. *Forest Ecology and Management* 335: 37–50.
- Hayes, D.S., J.M. Brändle, C. Seliger, et al. 2018. Advancing towards functional environmental flows for temperate floodplain rivers. *Science of the Total Environment* 633: 1089–1104.
- Hayhoe, K., D.J. Wuebbles, D.R. Easterling, et al. 2018. Our changing climate. p 72–144. In: D.R. Reidmiller et al., eds. *Impacts, Risks, and Adaptation in the United States: Fourth National Climate Assessment*, vol. II. Washington, DC: U.S. Global Change Research Program.
- Hayward, G.D., C.H. Flather, M.M. Rowland, et al. 2016. Applying the 2012 Planning Rule to Conserve Species: A Practitioner’s Reference. Washington, DC: U.S. Department of Agriculture, Forest Service.
- Heberling, J.M., C. McDonough MacKenzie, J.D. Fridley, S. Kalisz, and R.B. Primack. 2019. Phenological mismatch with trees reduces wildflower carbon budgets. *Ecology Letters* 22: 616–623.
- Heller, N.E., and R.J. Hobbs. 2014. Development of a natural practice to adapt conservation goals to global change. *Conservation Biology* 28: 696–704.
- Hellmann, J.J., J.E. Byers, B.G. Bierwagen, and J.S. Dukes. 2008. Five potential consequences of climate change for invasive species. *Conservation Biology* 22: 534–543.
- Hennon, P.E., C.M. McKenzie, D. D’Amore, et al. 2016. A Climate Adaptation Strategy for Conservation and Management of Yellow-cedar in Alaska. General Technical Report PNW-GTR-917. Portland, OR: U.S. Department of Agriculture, Forest Service, Pacific Northwest Research Station.
- Herrera-Estrada, J.E., and J. Sheffield. 2017. Uncertainties in future projections of summer droughts and heat waves over the contiguous United States. *Journal of Climate* 30: 6225–6246.
- Hessburg, P.F., S. Charnley, K.L. Wendel, et al. 2020. The 1994 Eastside Screens Large-tree Harvest Limit: Review of Science Relevant to Forest Planning 25 Years Later. General Technical Report PNW-GTR-990. Portland, OR: U.S. Department of Agriculture, Forest Service, Pacific Northwest Research Station.
- Hessburg, P.F., C.L. Miller, N.A. Povak, et al. 2019. Climate, environment, and disturbance history govern resilience of western North American forests. *Frontiers in Ecology and Evolution* 7: 239.
- Hessburg, P.F., T.A. Spies, D.A. Perry, et al. 2016. Tamm Review: Management of mixed-severity fire regime forests in Oregon, Washington, and northern California. *Forest Ecology and Management* 366: 221–250.

- Heyck-Williams, S., L. Anderson, and B.A. Stein. 2017. *Megafires: The Growing Risk to America's Forests, Communities, and Wildlife*. Washington, DC: National Wildlife Federation.
- Hicke, J.A., A.J. Meddens, and C.A. Kolden. 2016. Recent tree mortality in the western United States from bark beetles and forest fires. *Forest Science* 62: 141–153.
- Higgs, E., D.A. Falk, A. Guerrini, et al. 2014. The changing role of history in restoration ecology. *Frontiers in Ecology and the Environment* 12: 499–506.
- Higuera, P.E., and J.T. Abatzoglou. 2021. Record-setting climate enabled the extraordinary 2020 fire season in the western United States. *Global Change Biology* 27: 1–2.
- Higuera, P.E., A.L. Metcalf, C. Miller, et al. 2019. Integrating subjective and objective dimensions of resilience in fire-prone landscapes. *BioScience* 69: 379–388.
- Hof, A.R., C. Dymond, and J. Mladenoff. 2017. Climate change mitigation through adaptation: The effectiveness of forest diversification by novel tree planting regimes. *Ecosphere* 8: e01981.
- Holling, C.S. 1973. Resilience and stability of ecological systems. *Annual Review of Ecology and Systematics* 4: 1–23.
- Hudiburg, T.W., B.E. Law, W.R. Moomaw, M.E. Harmon, and J.E. Stenzel. 2019. Meeting GHG reduction targets requires accounting for all forest sector emissions. *Environmental Research Letters* 14: 095005.
- Hurteau, M.D., M.P. North, G.W. Koch, and B.A. Hungate. 2019. Opinion: Managing for disturbance stabilizes forest carbon. *Proceedings of the National Academy of Sciences* 116: 10193–10195.
- Iannone, B.V. III, S. Carnevale, M.B. Main, et al. 2020. Invasive species terminology: Standardizing for stakeholder education. *Journal of Extension* 58: v58-3a3.
- IPCC (Intergovernmental Panel on Climate Change). 2014. *Climate Change 2014: Mitigation of Climate Change. Contribution of Working Group III to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change*. (O. Edenhofer et al., eds.) Cambridge, UK: Cambridge University Press.
- IPCC (Intergovernmental Panel on Climate Change). 2018. *Global Warming of 1.5°C. An IPCC Special Report on the Impacts of Global Warming of 1.5°C Above Pre-industrial Levels and Related Global Greenhouse Gas Emission Pathways in the Context of Strengthening the Global Response to the Threat of Climate Change, Sustainable Development, and Efforts to Eradicate Poverty*. (V. Masson-Delmotte et al., eds.) Geneva, Switzerland: World Meteorological Organization.
- Isabel, N., J.A. Holliday, and S.N. Aitken. 2020. Forest genomics: Advancing climate adaptation, forest health, productivity, and conservation. *Evolutionary Applications* 13: 3–10.
- Iverson, L.R., and D. McKenzie. 2013. Tree-species range shifts in a changing climate: Detecting, modeling, assisting. *Landscape Ecology* 28: 879–889.
- Iverson, L.R., A.M. Prasad, S.N. Matthews, and M. Peters. 2008. Estimating potential habitat for 134 eastern US tree species under six climate scenarios. *Forest Ecology and Management* 234: 390–406.
- Iverson, L.R., A.M. Prasad, M.P. Peters, and S.N. Matthews. 2019. Facilitating adaptive forest management under climate change: A spatially specific synthesis of 125 species for habitat changes and assisted migration over the eastern United States. *Forests* 10: 989.
- Jackson, R.B., E.G. Jobbágy, R. Avissar, et al. 2005. Trading water for carbon with biological carbon sequestration. *Science* 310: 1944–1947.
- Jackson, S.T., and R.J. Hobbs. 2009. Ecological restoration in the light of ecological history. *Science* 325: 567–569.
- Jacobs, D.F., J.A. Oliet, J. Aronson, et al. 2015. Restoring forests: What constitutes success in the twenty-first century? *New Forests* 46: 601–614.
- James, J.N., N. Kates, C.D. Kuhn, et al. 2018. The effects of forest restoration on ecosystem carbon in western North America: A systematic review. *Forest Ecology and Management* 429: 625–641.
- James, N.A., K.L. Abt, G.E. Frey, X. Han, and J.P. Prestemon. 2020. *Fire in the Southern Appalachians: Understanding Impacts, Interventions, and Future Fire Events*. General Technical Report SRS-249. Research Triangle Park, NC: U.S. Department of Agriculture, Forest Service, Southern Research Station.
- Janowiak, M., W.J. Connelly, K. Dante-Wood, et al. 2017. *Considering Forest and Grassland Carbon in Land Management*. General Technical Report WO-95. Washington, DC: U.S. Department of Agriculture, Forest Service.
- Janowiak, M., A. Mahaffey, and C. Riely. 2020. *Moving the Needle: A Review of Needs to Increase Climate Adaptation in the Forests of New England*. Santa Fe, NM: Forest Stewards Guild.
- Janowiak, M.K., C.W. Swanston, L.M. Nagel, et al. 2014. A practical approach for translating climate change adaptation principles into forest management actions. *Journal of Forestry* 112: 424–433.
- Jantarasami, L.C., J.J. Lawler, and C.W. Thomas. 2010. Institutional barriers to climate change adaptation in US national parks and forests. *Ecology and Society* 15: 33.
- Jenkins, J., and M.W. Jenkins. 2017. Managed migration of coast redwoods: Subjectivity of stakeholders in Oregon's land use planning community. *Environment and Natural Resources Research* 7(3): 1–15.

- Johnstone, J.F., C.D. Allen, J.F. Franklin, et al. 2016. Changing disturbance regimes, ecological memory, and forest resilience. *Frontiers in Ecology and the Environment* 14: 369–378.
- Johnstone, J.F., T.N. Hollingsworth, F.S. Chapin III, and M.C. Mack. 2010. Changes in fire regime break the legacy lock on successional trajectories in Alaskan boreal forest. *Global Change Biology* 16: 1281–1295.
- Jones, G.M., R.J. Gutiérrez, D.J. Tempel, et al. 2016. Megafires: An emerging threat to old-forest species. *Frontiers in Ecology and the Environment* 14: 300–306.
- Joyce, L.A., G.M. Blate, J.S. Littell, et al. 2008. National forests. p 3-1 to 3-127. In: S.H. Julius and J.M. West, eds. *Preliminary Review of Adaptation Options for Climate-sensitive Ecosystems and Resources. A Report by the U.S. Climate Change Science Program and the Subcommittee on Global Change Research*. Washington, DC: U.S. Environmental Protection Agency.
- Joyce, L.A., G.M. Blate, S.G. McNulty, et al. 2009. Managing for multiple resources under climate change: National Forests. *Environmental Management* 44: 1022–1032.
- Joyce, L.A., and D. Coulson. 2020. *Climate Scenarios and Projections: A Technical Document Supporting the USDA Forest Service 2020 RPA Assessment*. General Technical Report RMRS-GTR-413. Fort Collins, CO: U.S. Department of Agriculture, Forest Service, Rocky Mountain Research Station.
- Joyce, L.A., and C.I. Millar. 2014. Improving the role of vulnerability assessments in decision support for effective climate adaptation. p 245–271. In: V.A. Sample and R.P. Bixler, eds. *Forest Conservation and Management in the Anthropocene: Conference Proceedings*. RMRS-P-71. Fort Collins, CO: U.S. Department of Agriculture, Forest Service, Rocky Mountain Research Station.
- Joyce, L.A., S.W. Running, D.D. Breshears, et al. 2014. Forests. p 175–194. In: J.M. Melillo et al., eds. *Climate Change Impacts in the United States: The Third National Climate Assessment*. Washington, DC: U.S. Global Change Research Program.
- Kabrick, J.M., D.C. Dey, J.W. Van Sambeek, et al. 2012. Quantifying flooding effects on hardwood seedling survival and growth for bottomland restoration. *New Forests* 43: 6950710.
- Karasov-Olson, A., M.W. Schwartz, J.D. Olden, et al. 2021. *Ecological Risk Assessment of Managed Relocation as a Climate Change Adaptation Strategy*. Natural Resource Report NPS/NRSS/NRR-2021/2241. Fort Collins, CO: National Park Service.
- Kates, R.W., W.R. Travis, and T.J. Wilbanks. 2012. Transformational adaptation when incremental adaptations to climate change are insufficient. *Proceedings of the National Academy of Sciences* 109: 7156–7161.
- Keane, R.E., K. Gray, B. Davis, L.M. Holsinger, and R. Loehman. 2019. Evaluating ecological resilience across wildfire suppression levels under climate and fuel treatment scenarios using landscape simulation modelling. *International Journal of Wildland Fire* 28: 533–549.
- Keane, R.E., P.F. Hessburg, P.B. Landres, and F.J. Swanson. 2009. The use of historical range and variability (HRV) in landscape management. *Forest Ecology and Management* 258: 1025–1037.
- Keane, R.E., R.A. Loehman, L.M. Holsinger, et al. 2018. Use of landscape simulation modeling to quantify resilience for ecological applications. *Ecosphere* 9: e02414.
- Keellings, D., and J.J.H. Ayala. 2019. Extreme rainfall associated with Hurricane Maria over Puerto Rico and its connections to climate variability and change. *Geophysical Research Letters* 46: 2964–2973.
- Keenan, R.J. 2015. Climate change impacts and adaptation in forest management: A review. *Annals of Forest Science* 72: 145–167.
- Keenan, T.F., J. Gray, M.A. Friedl, et al. 2014. Net carbon uptake has increased through warming-induced changes in temperate forest phenology. *Nature Climate Change* 4: 598–604.
- Keenan, T.F., D.Y. Hollinger, G. Bohrer, et al. 2013. Increase in forest water-use efficiency as atmospheric carbon dioxide concentrations rise. *Nature* 499: 324–327.
- Keeton, W.S. 2018. Source or sink? Carbon dynamics in eastern old-growth forests and their role in climate change mitigation. p 267–288. In: A.M. Barton and W.S. Keeton, eds. *Ecology and Recovery of Eastern Old-growth Forests*. Washington, DC: Island Press.
- Kelsey, R. 2019. *Wildfires and Forest Resilience: The Case for Ecological Forestry in the Sierra Nevada*. Sacramento, CA: The Nature Conservancy.
- Kemp, K.B., J.J. Blades, P.Z. Klos, et al. 2015. Managing for climate change on federal lands of the western United States: Perceived usefulness of climate science, effectiveness of adaptation strategies, and barriers to implementation. *Ecology and Society* 20: 17.
- Kemp, K.B., P.E. Higuera, P. Morgan, and J.T. Abatzoglou. 2019. Climate will increasingly determine post-fire tree regeneration success in low-elevation forests, Northern Rockies, USA. *Ecosphere* 10: e02568.
- Kershner, J., A. Woodward, and A. Torregrosa. 2020. *Integrating Climate Change Considerations into Natural Resource Planning—An Implementation Guide*. Reston, VA: U.S. Geological Survey.
- Kim, D., D. Medvigy, C.A. Maier, K. Johnsen, and S. Palmroth. 2020. Biomass increases attributed to both faster tree growth and altered allometric relationships under long-term carbon dioxide enrichment at a temperate forest. *Global Change Biology* 26: 2519–2533.

- Kirwan, M.L., and K.B. Gedan. 2019. Sea-level driven land conversion and the formation of ghost forests. *Nature Climate Change* 9: 450–457.
- Klapwijk, M.J., M.P. Ayres, A. Battisti, and S. Larsson. 2012. Assessing the impact of climate change on outbreak potential. p 429–450. In: P. Barbosa et al., eds. *Insect Outbreaks Revisited*. Chichester, UK: John Wiley & Sons, Ltd.
- Kleinman, S.J., T.E. DeGomez, G.B. Snider, and K.E. Williams. 2012. Large-scale pinyon ips (*Ips confusus*) outbreak in southwestern United States tied with elevation and land cover. *Journal of Forestry* 110: 194–200.
- Kliejunas, J.T. 2011. A risk assessment of climate change and the impact of forest diseases on forest ecosystems in the Western United States and Canada. General Technical Report PSW-GTR-236. Albany, CA: U.S. Department of Agriculture, Forest Service, Pacific Southwest Research Station.
- Kline, J.D., M.E. Harmon, T.A. Spies, et al. 2016. Evaluating carbon storage, timber harvest, and habitat possibilities for a Western Cascades (USA) forest landscape. *Ecological Applications* 26: 2044–2059.
- Kline, J.D., and M.J. Mazzotta. 2012. Evaluating Trade-offs among Ecosystem Services in the Management of Public Lands. General Technical Report PNW-GTR-865. Portland, OR: U.S. Department of Agriculture, Forest Service, Pacific Northwest Research Station.
- Klos, R.J., G.G. Wang, W.L. Bauerle, and J.R. Rieck. 2009. Drought impact on forest growth and mortality in the southeast USA: An analysis using Forest Health and Monitoring data. *Ecological Applications* 19: 699–708.
- Knott, J.A., J.M. Desprez, C.M. Oswalt, and S. Fei. 2019. Shifts in forest composition in the eastern United States. *Forest Ecology and Management* 433: 176–183.
- Knott, J.A., M.A. Jenkins, C.M. Oswalt, and S. Fei. 2020. Community-level responses to climate change in forests of the eastern United States. *Global Ecology and Biogeography* 29: 1299–1314. doi: 10.1111/geb.13102
- Koch, F.H., D. Yemshanov, D.W. McKenney, and W.D. Smith. 2009. Evaluating critical uncertainty thresholds in a spatial model of forest pest invasion risk. *Risk Analysis: An International Journal* 29: 1227–1241.
- Kolb, T.E., J.K. Agee, P.Z. Fule, et al. 2007. Perpetuating old ponderosa pine. *Forest Ecology and Management* 249: 141–157.
- Koontz, M.J., M.P. North, C.M. Werner, S.E. Fick, and A.M. Latimer. 2019. Local forest structure variability increases resilience to wildfire in dry western US coniferous forests. *Ecology Letters* 23: 483–495.
- Kopp, R.E., K. Hayhoe, D.R. Easterling, et al. 2017. Potential surprises—Compound extremes and tipping elements. p 411–429. In: D.J. Wuebbles et al., eds. *Climate Science Special Report: Fourth National Climate Assessment*, vol. I. Washington, DC: U.S. Global Change Research Program.
- Krawchuk, M.A., G.W. Meigs, J.M. Cartwright, et al. 2020. Disturbance refugia within mosaics of forest fire, drought, and insect outbreaks. *Frontiers in Ecology and the Environment* 18: 235–244.
- Krofcheck, D.J., M.D. Hurteau, R.M. Scheller, and E.L. Loudermilk. 2017. Restoring surface fire stabilizes forest carbon under extreme fire weather in the Sierra Nevada. *Ecosphere* 8: e01663.
- Krofcheck, D.J., E.L. Loudermilk, J.K. Hiers, R.M. Scheller, and M.D. Hurteau. 2019. The effects of management on long-term carbon stability in a southeastern US forest matrix under extreme fire weather. *Ecosphere* 10: e02631.
- Kueffer, C. 2015. Ecological novelty: Toward an interdisciplinary understanding of ecological change in the Anthropocene. p 19–37. In: H. Greschke and J. Tischler, eds. *Grounding Global Climate Change*. Berlin: Springer.
- Kupfer, J.A., A.J. Terando, P. Gao, C. Teske, and J.K. Hiers. 2020. Climate change is projected to reduce prescribed burning opportunities in the south-eastern United States. *International Journal of Wildland Fire* 29: 764–778.
- Laatsch, J., and Z. Ma. 2015. Strategies for incorporating climate change into public forest management. *Journal of Forestry* 113: 335–342.
- Landres, P.B., P. Morgan, and F.J. Swanson. 1999. Overview of the use of natural variability concepts in managing ecological systems. *Ecological Applications* 9: 1179–1188.
- Lee, E.H., P.A. Beedlow, R.S. Waschmann, et al. 2017. Regional patterns of increasing Swiss needle cast impacts on Douglas-fir growth with warming temperatures. *Ecology and Evolution* 7: 11167–11196.
- Lenoir, J., and J.C. Svenning. 2015. Climate-related range shifts—A global multidimensional synthesis and new research directions. *Ecography* 38: 15–28.
- Leppanen, C., and D. Simberloff. 2017. Implications of early production in an invasive pest. *Agricultural and Forest Entomology* 19: 217–224.
- Leverkus, A.B., B. Buma, J. Wagenbrenner, et al. 2021. Tamm Review: Does salvage logging mitigate subsequent forest disturbances? *Forest Ecology and Management* 481: 118721.
- Leverkus, A.B., and J. Castro. 2017. An ecosystem services approach to the ecological effects of salvage logging: Valuation of seed dispersal. *Ecological Applications* 27: 1057–1063.
- Leverkus, A.B., J.M. Rey Benayas, J. Castro, et al. 2018. Salvage logging effects on regulating and supporting ecosystem services—A systematic map. *Canadian Journal of Forest Research* 48: 983–1000.

- Liang, Y., M.J. Duveneck, E.J. Gustafson, J.M. Serra-Diaz, and J.R. Thompson. 2018. How disturbance, competition, and dispersal interact to prevent tree range boundaries from keeping pace with climate change. *Global Change Biology* 24: e335–e351.
- Lindenmayer, D.B., P.J. Burton, and J.F. Franklin. 2012. *Salvage Logging and Its Ecological Consequences*. Washington, DC: Island Press.
- Lindenmayer, D.B., and R.F. Noss. 2006. Salvage logging, ecosystem processes, and biodiversity conservation. *Conservation Biology* 20: 949–958.
- Litschert, S.F., T.C. Brown, and D.M. Theobald. 2012. Historic and future extent of wildfires in the Southern Rockies ecoregion, USA. *Forest Ecology and Management* 269: 124–133.
- Littell, J.S., D.L. Peterson, C.I. Millar, and K.A. O'Halloran. 2012. U.S. National Forests adapt to climate change through Science–Management partnerships. *Climatic Change* 110: 269–296.
- Littell, J.S., D.L. Peterson, K.L. Riley, Y. Liu, and C.H. Luce. 2016. A review of the relationships between drought and forest fire in the United States. *Global Change Biology* 22: 2353–2369.
- Liu, Y., S.L. Goodrick, and J.A. Stanturf. 2013. Future US wildfire potential trends projected using a dynamically downscaled climate change scenario. *Forest Ecology and Management* 294: 120–135.
- Loehman, R.A., R.E. Keane, and L.M. Holsinger. 2020. Simulation modeling of complex climate, wildfire, and vegetation dynamics to address wicked problems in land management. *Frontiers in Forests and Global Change* 3: 3.
- Loehman, R.A., R.E. Keane, L.M. Holsinger, and Z. Wu. 2017. Interactions of landscape disturbances and climate change dictate ecological pattern and process: Spatial modeling of wildfire, insect, and disease dynamics under future climates. *Landscape Ecology* 32: 1447–1459.
- Löf, M., P. Madsen, M. Metslaid, J. Witzell, and D.F. Jacobs. 2019. Restoring forests: Regeneration and ecosystem function for the future. *New Forests* 50: 139–151.
- Long, J., M. Windmuller-Campione, and R. DeRose. 2018. Building resistance and resilience: Regeneration should not be left to chance. *Forests* 9: 270.
- Long, J.W., L. Quinn-Davidson, and C.N. Skinner, eds. 2014. *Science Synthesis to Support Socioecological Resilience in the Sierra Nevada and Southern Cascade Range*. General Technical Report PSW-GTR-247. Albany, CA: U.S. Department of Agriculture, Forest Service, Pacific Southwest Research Station.
- Loudermilk, E.L., R.M. Scheller, P.J. Weisberg, and A. Kretchun. 2017. Bending the carbon curve: Fire management for carbon resilience under climate change. *Landscape Ecology* 32: 1461–1472.
- Luce, C., P. Morgan, K. Dwire, et al. D. 2012. *Climate Change, Forests, Fire, Water, and Fish: Building Resilient Landscapes, Streams, and Managers*. General Technical Report RMRS-GTR-290. Fort Collins, CO: U.S. Department of Agriculture, Forest Service, Rocky Mountain Research Station.
- Lüthi, D., M. Le Floch, B. Bereiter, et al. 2008. High-resolution carbon dioxide concentration record 650,000–800,000 years before present. *Nature* 453: 379–382.
- MacKenzie, W.H., and C.R. Mahony. 2021. An ecological approach to climate change-informed tree species selection for reforestation. *Forest Ecology and Management* 481: 118705.
- Maestas, J.D., D.E. Naugle, J.C. Chambers, et al. 2021. Conifer expansion. p 139–152. In: T.E. Remington et al., eds. *Sagebrush Conservation Strategy—Challenges to Sagebrush Conservation*. Open-File Report 2020-1125. Reston, VA: U.S. Geological Survey.
- Magness, D.R., and J.M. Morton. 2018. Using climate envelope models to identify potential ecological trajectories on the Kenai Peninsula, Alaska. *PLoS One* 13: e0208883.
- Mamet, S.D., C.D. Brown, A.J. Trant, and C.P. Laroque. 2019. Shifting global *Larix* distributions: Northern expansion and southern retraction as species respond to changing climate. *Journal of Biogeography* 46: 30–44.
- Martin, K.L., M.D. Hurteau, B.A. Hungate, G.W. Koch, and M.P. North. 2015. Carbon trade-offs of restoration and provision of endangered species habitat in a fire-maintained forest. *Ecosystems* 18: 76–88.
- Mathias, J.M., and R.B. Thomas. 2018. Disentangling the effects of acidic air pollution, atmospheric CO<sub>2</sub>, and climate change on recent growth of red spruce trees in the Central Appalachian Mountains. *Global Change Biology* 24: 3938–3953.
- Matthews, S.N., L.R. Iverson, M.P. Peters, A.M. Prasad, and S. Subburayalu. 2014. Assessing and comparing risk to climate changes among forested locations: Implications for ecosystem services. *Landscape Ecology* 29: 213–228.
- Mazdiyasi, O., and A. AghaKouchak. 2015. Substantial increase in concurrent droughts and heatwaves in the United States. *Proceedings of the National Academy of Sciences* 112: 11484–11489.
- McCauley, L.A., M.D. Robles, T. Woolley, et al. 2019. Large-scale forest restoration stabilizes carbon under climate change in Southwest United States. *Ecological Applications* 29: e01979.
- McClenahan, L., A.B. Cooper, M.G. McKenzie, and J.A. Drew. 2015. The importance of surprising results and best practices in historical ecology. *BioScience* 65: 932–939.

- McComb, B., and S. Duncan. 2007. *Biodiversity Conservation in Contemporary Landscapes, Stressors, and Ranges of Variability: Scientific and Social Views*. Washington, DC: National Commission on Science for Sustainable Forestry.
- McCurry, J.R., M.J. Gray, and D.C. Mercker. 2010. Early growing season flooding influence on seedlings of three common bottomland hardwood species in western Tennessee. *Journal of Fish and Wildlife Management* 1: 11–18.
- McKelvey, K.S., W.M. Block, T.B. Jain, et al. 2021. Adapting research, management, and governance to confront socioecological uncertainties in novel ecosystems. *Frontiers in Forests and Global Change* 4: 14.
- McKenzie, D., Z.E. Gedalof, D.L. Peterson, and P. Mote. 2004. Climatic change, wildfire, and conservation. *Conservation Biology* 18: 890–902.
- McKenzie, D., and J.S. Littell. 2011. Climate change and wilderness fire regimes. *International Journal of Wilderness* 17: 22–27.
- McNulty, S., E. Treasure, L. Jennings, et al. 2018. Translating national level forest service goals to local level land management: Carbon sequestration. *Climatic Change* 146: 133–144.
- McNulty, S.G., G. Sun, J.A.M. Myers, E.C. Cohen, and P. Caldwell. 2010. Robbing Peter to pay Paul: Trade-offs between ecosystem carbon sequestration and water yield. p 103–114. In: K.W. Potter and D.K. Frevert, eds. *Watershed Management 2010: Innovations in Watershed Management under Land Use and Climate Change*. Washington, DC: Environmental and Water Resources Institute.
- McWethy, D.B., T. Schoennagel, P.E. Higuera, et al. 2019. Rethinking resilience to wildfire. *Nature Sustainability* 2: 797–804.
- Melaas, E.K., D. Sulla-Menashe, and M.A. Friedl. 2018. Multidecadal changes and interannual variation in springtime phenology of North American temperate and boreal deciduous forests. *Geophysical Research Letters* 45: 2679–2687.
- Meyer, M.D., J.W. Long, and H.D. Safford. 2021. *Postfire Restoration Framework for National Forests in California*. General Technical Report PSW-GTR-270. Albany, CA: U.S. Department of Agriculture, Forest Service, Pacific Southwest Research Station.
- Millar, C.I. 2014. Historic variability: Informing restoration strategies, not prescribing targets. *Journal of Sustainable Forestry* 33: S28–S42.
- Millar, C.I., and L.B. Brubaker. 2006. Climate change and paleoecology: New contexts for restoration ecology. p 315–340 in: D.A. Falk et al., eds. *Foundations of Restoration Ecology*. Washington, DC: Island Press.
- Millar, C.I., and N.L. Stephenson. 2015. Temperate forest health in an era of emerging megadisturbance. *Science* 349: 823–826.
- Millar, C.I., N.L. Stephenson, and S.L. Stephens. 2007. Climate change and forests of the future: Managing in the face of uncertainty. *Ecological Applications* 17: 2145–2151.
- Miller, A.D., J.R. Thompson, A.J. Tepley, and K.J. Anderson-Teixeira. 2019. Alternative stable equilibria and critical thresholds created by fire regimes and plant responses in a fire-prone community. *Ecography* 42: 55–66.
- Miller, J.D., and H. Safford. 2012. Trends in wildfire severity: 1984 to 2010 in the Sierra Nevada, Modoc Plateau, and southern Cascades, California, USA. *Fire Ecology* 8: 41–57.
- Miller, J.R., and B.T. Bestelmeyer. 2016. What’s wrong with novel ecosystems, really? *Restoration Ecology* 24: 577–582.
- Milly, P.C.D., J. Betancourt, M. Falkenmark, et al. 2008. Stationarity is dead: Whither water management? *Earth* 4: 20.
- Mitchell, R.J., Y. Liu, J.J. O’Brien, et al. 2014. Future climate and fire interactions in the southeast region of the United States. *Forest Ecology and Management* 327: 316–326.
- Mitton, J.B., and S.M. Ferrenberg. 2012. Mountain pine beetle develops an unprecedented summer generation in response to climate warming. *American Naturalist* 179: E163–E171.
- Monahan, W.B., A. Rosemartin, K.L. Gerst, et al. 2016. Climate change is advancing spring onset across the U.S. national park system. *Ecosphere* 7: e01465.
- Moomaw, W.R., S.A. Masino, and E.K. Faison. 2019. Intact forests in the United States: Proforestation mitigates climate change and serves the greatest good. *Frontiers in Forests and Global Change* 2: 27.
- Morecroft, M.D., S. Duffield, M. Harley, et al. 2019. Measuring the success of climate change adaptation and mitigation in terrestrial ecosystems. *Science* 366: 6471.
- Morelli, T.L., C. Daly, S.Z. Dobrowski, et al. 2016. Managing climate refugia for climate adaptation. *PLoS One* 11: e0159909.
- Morin, X., L. Fahse, H. Jactel, et al. 2018. Long-term response of forest productivity to climate change is mostly driven by change in tree species composition. *Scientific Reports* 8: 1–12.
- Mote, P.W., S. Li, D.P. Lettenmaier, M. Xiao, and R. Engel. 2018. Dramatic declines in snowpack in the western US. *npj Climate and Atmospheric Science* 1: 1–6.
- Mote, P.W., D.E. Rupp, S. Li, et al. 2016. Perspectives on the causes of exceptionally low 2015 snowpack in the western United States. *Geophysical Research Letters* 43: 10980–10988.
- Mozelewski, T.G., and R.M. Scheller. 2021. Forecasting for intended consequences. *Conservation Science and Practice*: e370. doi: 10.1111/csp.2.370

- Müller, J., R.F. Noss, S. Thorn, et al. 2019. Increasing disturbance demands new policies to conserve intact forest. *Conservation Letters* 12: e12449.
- Murcia, C., J. Aronson, G.H. Kattan, et al. 2014. A critique of the 'novel ecosystem' concept. *Trends in Ecology and Evolution* 29: 548–553.
- MUSYA (Multiple-Use Sustained-Yield Act of 1960). 1960. Public Law 86-517, 86th Congress (June 12, 1960), 16 U.S. Code §528 et seq.
- Naficy, C., A. Sala, E.G. Keeling, J. Graham, and T.H. DeLuca. 2010. Interactive effects of historical logging and fire 'exclusion on ponderosa pine forest structure in the northern Rockies. *Ecological Applications* 20: 1851–1864.
- Nagel, L.M., B.J. Pakik, M.A. Battaglia, et al. 2017. Adaptive silviculture for climate change: A national experiment in manager-scientist partnerships to apply an adaptation framework. *Journal of Forestry* 115: 167–178.
- NASEM (National Academies of Sciences, Engineering, and Medicine). 2019. *Negative Emissions Technologies and Reliable Sequestration: A Research Agenda*. Washington, DC: National Academies Press.
- Nave, L.E., B.F. Walters, K.L. Hofmeister, et al. 2019. The role of reforestation in carbon sequestration. *New Forests* 50: 115–137.
- Nepal, P., P.J. Ince, K.E. Skog, and S.J. Chang. 2012. Projection of US forest sector carbon sequestration under US and global timber market and wood energy consumption scenarios, 2010–2060. *Biomass and Bioenergy* 45: 251–264.
- NOAA (National Oceanic and Atmospheric Administration). 2020. *Climate change: Atmospheric carbon dioxide*. Washington, DC: NOAA (accessed May 19, 2021). <https://www.climate.gov/news-features/understanding-climate/climate-change-atmospheric-carbon-dioxide>
- NOAA (National Oceanic and Atmospheric Administration). 2021. *2020 was Earth's 2nd-hottest year, just behind 2016*. Washington, DC: NOAA (accessed March 5, 2021). <https://www.noaa.gov/news/2020-was-earth-s-2nd-hottest-year-just-behind-2016>
- Norby, R.J., E.H. DeLucia, B. Gielen, et al. 2005. Forest response to elevated CO<sub>2</sub> is conserved across a broad range of productivity. *Proceedings of the National Academy of Sciences* 102: 18052–18056.
- North, M., P. Stine, K. O'Hara, W. Zileinski, and S. Stephens. 2009. *An Ecosystem Management Strategy for Sierran Mixed Conifer Forests*. General Technical Report PSW-GTR-220. Albany, CA: U.S. Department of Agriculture, Forest Service, Pacific Southwest Research Station.
- North, M.P., J.T. Stevens, D.F. Greene, et al. 2019. Tamm Review: Reforestation for resilience in dry western US forests. *Forest Ecology and Management* 432: 209–224.
- NPS (National Park Service). 2021. *Planning for a Changing Climate: Climate-Smart Planning and Management in the National Park Service*. Fort Collins, CO: NPS Climate Change Response Program.
- Nunez-Mir, G.C., B.V. Iannone III, K. Curtis, and S. Fei. 2015. Evaluating the evolution of forest restoration research in a changing world: A "big literature" review. *New Forests* 46: 669–682.
- Nydick, K., and C. Sydorik. 2011. Alternative futures for fire management under a changing climate. *Park Science* 28: 44–47.
- Ojima, D.S., I.R. Iverson, B.I. Sohngen, et al. 2014. Risk assessment. p 223–244. In: D.L. Peterson et al., eds. *Climate Change and United States Forests*. Dordrecht, Netherlands: Springer.
- Olander, L., K. Warnell, T. Warziniack, et al. 2021. Exploring the use of ecosystem services conceptual models to account for the benefits of public lands: An example from national forest planning in the United States. *Forests* 12: 267.
- Oliver, T.H., M.S. Heard, N.J. Isaac, et al. 2015. Biodiversity and resilience of ecosystem functions. *Trends in Ecology and Evolution* 30: 673–684.
- Olliff, S.T., and A.J. Hansen. 2016. Challenges and approaches for integrating climate science into federal land management. p 33–54. In: A.J. Hansen et al., eds. *Climate Change in Wildlands: Pioneering Approaches to Science and Management*. Washington, DC: Island Press.
- Olson, M.G., B.O. Knapp, and J.M. Kabrick. 2017. Dynamics of a temperate deciduous forest under landscape-scale management: Implications for adaptability to climate change. *Forest Ecology and Management* 387: 73–85.
- Ontl, T.A., M.K. Janowiak, C.W. Swanston, et al. 2020. Forest management for carbon sequestration and climate adaptation. *Journal of Forestry* 118: 86–101.
- Ontl, T.A., C. Swanston, L.A. Brandt, et al. 2018. Adaptation pathways: Ecoregion and land ownership influences on climate adaptation decision-making in forest management. *Climatic Change* 146: 75–88.
- Oswalt, S.N., W.B. Smith, P.D. Miles, and S.A. Pugh. 2019. *Forest Resources of the United States, 2017. A Technical Document Supporting the Forest Service 2020 RPA Assessment*. General Technical Report WO-97. Washington, DC: U.S. Department of Agriculture, Forest Service.
- Palik, B.J., and A.W. D'Amato. 2017. Ecological forestry: Much more than retention harvesting. *Journal of Forestry* 115: 51–53.

- Panchen, Z.A., R.B. Primack, A.S. Gallinat, et al. 2015. Substantial variation in leaf senescence times among 1360 temperate woody plant species: Implications for phenology and ecosystem processes. *Annals of Botany* 116: 865–873.
- Parent, G.J., C. Méndez-Espinoza, I. Giguère, et al. 2020. Hydroxyacetophenone defenses in white spruce against spruce budworm. *Evolutionary Applications* 13: 62–75.
- Park, A., K. Puettmann, E. Wilson, et al. 2014. Can boreal and temperate forest management be adapted to the uncertainties of 21st century climate change? *Critical Reviews in Plant Sciences* 33: 251–285.
- Parks, S.A., and J.T. Abatzoglou. 2020. Warmer and drier fire seasons contribute to increases in area burned at high severity in western US forests from 1985 to 2017. *Geophysical Research Letters* 47: e2020GL089858.
- Parks, S.A., S.Z. Dobrowski, J.D. Shaw, and C. Miller. 2019a. Living on the edge: Trailing edge forests at risk of fire-facilitated conversion to non-forest. *Ecosphere* 10: e02651.
- Parks, S.A., S.Z. Dobrowski, J.D. Shaw, and C. Miller. 2019b. Quantifying the risk of fire-facilitated transition to non-forest in California and the Southwest. Final Report. JFSP Project ID: 15-1-03-20. Boise, ID: Joint Fire Science Program.
- Patricola, C.M., and M.F. Wehner. 2018. Anthropogenic influences on major tropical cyclone events. *Nature* 563: 339–346.
- Perring, D.L., R.J. Standish, and R.J. Hobbs. 2013. Incorporating novelty and novel ecosystems into restoration planning and practice in the 21st century. *Ecological Processes* 2: 18.
- Peterson, D.L., J.E. Halofsky, and L.S. Pile. 2019. Managing effects of drought in the Interior West. p 123–129. In: J.M. Vose et al., eds. *Effects of Drought on Forests and Rangelands in the United States: Translating Science into Management Responses*. General Technical Report WO-98. Washington, DC: U.S. Department of Agriculture, Forest Service, Washington Office.
- Peterson, D.L., C.I. Millar, L.A. Joyce, et al. 2011. Responding to Climate Change in National Forests: A Guidebook for Developing Adaptation Actions. General Technical Report PNW-GTR-855. Portland, OR: U.S. Department of Agriculture, Forest Service, Pacific Northwest Research Station.
- Pike, C., K.M. Potter, P. Berrang, et al. 2020. New seed-collection zones for the eastern United States: The Eastern Seed Zone Forum. *Journal of Forestry* 118: 444–451.
- Pinchot, G. 1905. *A Primer of Forestry Part II: Practical Forestry*. New York: Harcourt Brace.
- Poland, T.M., T. Patel-Weyand, D.M. Finch, et al. 2021. *Invasive Species in Forests and Rangelands of the United States: A Comprehensive Science Synthesis for the United States Forest Sector*. Cham, Switzerland: Springer Verlag.
- Prasad, A., J. Pedlar, M. Peters, et al. 2020. Combining US and Canadian forest inventories to assess habitat suitability and migration potential of 25 tree species under climate change. *Diversity and Distributions* 26: 1142–1159.
- Prober, S.M., M. Byrne, E.H. McLean, et al. 2015. Climate-adjusted provenancing: A strategy for climate-resilient ecological restoration. *Frontiers in Ecology and Evolution* 3: 65.
- Prober, S.M., V.A. Doerr, L.M. Broadhurst, K.J. Williams, and F. Dickinson. 2019. Shifting the conservation paradigm: A synthesis of options for renovating nature under climate change. *Ecological Monographs* 89: e01333.
- Prober, S.M., and M. Dunlop. 2011. Climate change: A cause for new biodiversity conservation objectives, but let's not throw the baby out with the bathwater. *Ecological Management and Restoration* 12: 2.
- Pureswaran, D.S., A. Roques, and A. Battisti. 2018. Forest insects and climate change. *Current Forestry Reports* 4: 35–50.
- Radeloff, V.C., S.I. Stewart, T.J. Hawbaker, et al. 2010. Housing growth in and near United States protected areas limits their conservation value. *Proceedings of the National Academy of Sciences* 107: 940–945.
- Radeloff, V.C., J.W. Williams, B.L. Bateman, et al. 2015. The rise of novelty in ecosystems. *Ecological Applications* 25: 2051–2068.
- Radtke, P., D. Walker, J. Frank, et al. 2017. Improved accuracy of aboveground biomass and carbon estimates for live trees in forests of the eastern United States. *Forestry: An International Journal of Forest Research* 90: 32–46.
- Raymond, C.L., D.L. Peterson, and R.M. Rochefort. 2014. *Climate Change Vulnerability and Adaptation in the North Cascades Region, Washington*. General Technical Report PNW-GTR-892. Portland, OR: U.S. Department of Agriculture, Forest Service, Pacific Northwest Research Station.
- Régnière, J., R. St-Amant, and P. Duval. 2012. Predicting insect distributions under climate change from physiological responses: Spruce budworm as an example. *Biological Invasions* 14: 1571–1586.
- Rehfeldt, G.E., M.V. Warwell, and R.A. Monserud. 2020. Species, climatypes, climate change, and forest health: A conversion of science to practice for inland northwest (USA) forests. *Forests* 11: 1237.
- Restaino, C.M., D.L. Peterson, and J. Littell. 2016. Increased water deficit decreases Douglas-fir growth throughout Western U.S. forests. *Proceedings of the National Academy of Sciences* 113: 9557–9562.

- Reyer, C.P., N. Brouwers, A. Rammig, et al. 2015. Forest resilience and tipping points at different spatio-temporal scales: Approaches and challenges. *Journal of Ecology* 103: 5–15.
- Rice, A. 2019. Reexamining the utility of existing climate adaptation frameworks through application on a northern forest. Master of Science thesis, Michigan Technological University.
- Richardson, A.D., T.F. Keenan, M. Migliavacca, et al. 2013. Climate change, phenology, and phenological control of vegetation feedbacks to the climate system. *Agricultural and Forest Meteorology* 169: 156–173.
- Risser, M.D., and M.F. Wehner. 2017. Attributable human-induced changes in the likelihood and magnitude of the observed extreme precipitation during Hurricane Harvey. *Geophysical Research Letters* 44: 12457–12464.
- Rissman, A.R., K.D. Burke, H.A.C. Kramer, et al. 2018. Forest management for novelty, persistence, and restoration influenced by policy and society. *Frontiers in Ecology and the Environment* 16: 454–462.
- Rodman, K.C., T.T. Veblen, M.A. Battaglia, et al. 2020. A changing climate is snuffing out post-fire recovery in montane forests. *Global Ecology and Biogeography* 29: 2039–2051.
- Rodriguez-Franco, C., and T.J. Haan. 2015. Understanding climate change perceptions, attitudes, and needs of Forest Service resource managers. *Journal of Sustainable Forestry* 34: 423–444.
- Rohwer, Y., and E. Marris. 2021. Ecosystem integrity is neither real nor valuable. *Conservation Science and Practice* 3: e411.
- Ruhl, J.B., and J.E. Salzman. 2020. Ecosystem services and federal lands: A quiet revolution in public lands management. *University of Colorado Law Review* 91: 677–708.
- Running, S.W. 2006. Is global warming causing more, larger wildfires? *Science* 313: 927–928.
- Rustad, L., J. Campbell, J.S. Dukes, et al. 2012. Changing Climate, Changing Forests: The Impacts of Climate Change on Forests of the Northeastern United States and Eastern Canada. General Technical Report NRS-99. Newtown Square, PA: U.S. Department of Agriculture, Forest Service, Northern Research Station.
- Ryan, M.G., M.E. Harmon, R.A. Birdsey, et al. 2010. A Synthesis of the Science on Forests and Carbon for U.S. Forests. *Issues in Ecology* 13. Washington, DC: Ecological Society of America.
- Ryan, R.L., and E. Hamin. 2009. Wildland–urban interface communities’ response to post-fire salvage logging. *Western Journal of Applied Forestry* 24: 36–41.
- Safford, H.D., and J.T. Stevens. 2017. Natural Range of Variation for Yellow Pine and Mixed-conifer Forests in the Sierra Nevada, Southern Cascades, and Medoc and Inyo National Forests, California, USA. General Technical Report PSW-GTR-256. Albany, CA: U.S. Department of Agriculture, Forest Service, Pacific Southwest Research Station.
- Safford, H.D., J.A. Wiens, and A.G.D. Hayward. 2012. The growing importance of the past in managing ecosystems of the future. p 319–327 in: J.A. Weins et al., eds. *Historical Environmental Variation in Conservation and Natural Resource Management*. Oxford, UK: Wiley-Blackwell.
- Sample, V.A. 2017. Potential for additional carbon sequestration through regeneration of nonstocked forest land in the United States. *Journal of Forestry* 115: 309–318.
- Sample, V.A., R. Birdsey, R.A. Houghton, et al. 2015. *Forest Carbon Conservation and Management: Integration with Sustainable Forest Management for Multiple Resource Values and Ecosystem Services*. Washington, DC: Pinchot Institute for Conservation.
- Sánchez-Pinillos, M., A. Leduc, A. Ameztegui, et al. 2019. Resistance, resilience, and change: Post-disturbance dynamics of boreal forests after insect outbreaks. *Ecosystems* 22: 1886–1901.
- Sanford, T., R. Wang, and A. Kenward. 2015. *The Age of Alaskan Wildfires*. Princeton, NJ: Climate Central.
- Scheffer, M., S.R. Carpenter, V. Dakos, and E.H. van Nes. 2015. Generic indicators of ecological resilience: Inferring the chance of a critical transition. *Annual Review of Ecology, Evolution, and Systematics* 46: 145–167.
- Scheller, R.M., and R. Parajuli. 2018. Forest management for climate change in New England and the Klamath ecoregions: Motivations, practices, and barriers. *Forests* 9: 626.
- Schoennagel, T., J.K. Balch, H. Brenkert-Smith, et al. 2017. Adapt to more wildfire in western North American forests as climate changes. *Proceedings of the National Academy of Sciences* 114: 4582–4590.
- Schultz, C.A., T. Jedd, and R.D. Beam. 2012. The Collaborative Forest Landscape Restoration Program: A history and overview of the first projects. *Journal of Forestry* 110: 381–391.
- Schultz, C.A., M.P. Thompson, and S.M. McCaffrey. 2019a. Forest Service fire management and the elusiveness of change. *Fire Ecology* 15: 13.
- Schultz, C.A., J.T. Timberlake, Z. Wurtzebach, et al. 2019b. Policy tools to address scale mismatches. *Ecology and Society* 24: 21.
- Schuurman, G.W., C.H. Hoffman, C.H. Cole, et al. 2020. Resist-Accept-Direct (RAD)—A Framework for the 21st-century Natural Resource Manager. Natural Resource Report NPS/NRSS/CCRP/NRR-2020/2213. Fort Collins, CO: National Park Service.

- Seager, R., A. Tzanova, and J. Nakamura. 2009. Drought in the southeastern United States: Causes, variability over the last millennium, and the potential for future hydroclimate change. *Journal of Climate* 22: 5021–5045.
- Seddon, N., B. Turner, P. Berry, A. Chausson, and C.A.J. Girardin. 2019. Grounding nature-based climate solutions in sound biodiversity science. *Nature Climate Change* 9: 84–87.
- Seidl, R., D.C. Donato, K.F. Raffa, and M.G. Turner. 2016. Spatial variability in tree regeneration after wildfire delays and dampens future bark beetle outbreaks. *Proceedings of the National Academy of Sciences* 113: 13075–13080.
- Seidl, R., W. Rammer, D. Jäger, W.S. Currie, and M.J. Lexer. 2007. Assessing trade-offs between carbon sequestration and timber production within a framework of multi-purpose forestry in Austria. *Forest Ecology and Management* 248: 64–79.
- Seidl, R., W. Rammer, R.M. Scheller, and T.A. Spies. 2012. An individual-based process model to simulate landscape-scale forest ecosystem dynamics. *Ecological Modelling* 231: 87–100.
- Seidl, R., D. Thom, M. Kautz, et al. 2017. Forest disturbances under climate change. *Nature Climate Change* 7: 395–401.
- Sellers, P.J., D.S. Schimel, B. Moore, J. Liu, and A. Eldering. 2018. Observing carbon cycle-climate feedbacks from space. *Proceedings of the National Academy of Sciences* 115: 7860–7868.
- Selles, O.A., and A.R. Rissman. 2020. Content analysis of resilience in forest fire science and management. *Land Use Policy* 94: 104483.
- Shannon, P.D., C.W. Swanston, M.K. Janowiak, et al. 2019. Adaptation strategies and approaches for forested watersheds. *Climate Services* 13: 51–64.
- Short, M.F., M.C. Stambaugh, and D.C. Dey. 2019. Prescribed fire effects on oak woodland advance regeneration at the prairie-forest border in Kansas, USA. *Canadian Journal of Forest Research* 49: 1570–1579.
- Singleton, M.P., A.E. Thode, A.J. Sánchez Meador, and J.M. Iniguez. 2019. Increasing trends in high-severity fire in the southwestern USA from 1984 to 2015. *Forest Ecology and Management* 433: 709–719.
- Sittaro, F., A. Paquette, C. Messier, and C.A. Nock. 2017. Tree range expansion in eastern North America fails to keep pace with climate warming at northern range limits. *Global Change Biology* 23: 3292–3301.
- Smith, J.E., G.M. Domke, M.C. Nichols, and B.F. Walters. 2019. Carbon stocks and stock change on federal forest lands of the United States. *Ecosphere* 10: e02367.
- Smith, N., R. Deal, J.D. Kline, et al. 2011. Using Ecosystem Services as a Framework for Forest Stewardship: Deschutes National Forest Overview. General Technical Report PNW-GTR-852. Portland, OR: U.S. Department of Agriculture, Forest Service, Pacific Northwest Research Station.
- Solarik, K.A., C. Messier, R. Ouimet, Y. Bergeron, and D. Gravel. 2018. Local adaptation of trees at the range margins impacts range shifts in the face of climate change. *Global Ecology and Biogeography* 27: 1507–1519.
- Solomon, A., R. Birdsey, L.A. Joyce, and J. Hayes. 2009. Forest Service Global Change Research Strategy, 2009–2019. FS-917a. Washington, DC: U.S. Department of Agriculture, Forest Service.
- Sommers, W.T., R.A. Loehman, and C.C. Hardy. 2014. Wildland fire emissions, carbon, and climate: Science overview and knowledge needs. *Forest Ecology and Management* 317: 1–8.
- Sorensen, P.O., A.C. Finzi, M.A. Giasson, et al. 2018. Winter soil freeze-thaw cycles lead to reductions in soil microbial biomass and activity not compensated for by soil warming. *Soil Biology and Biochemistry* 116: 39–47.
- Soto-Navarro, C., C. Ravilious, A. Arnell, et al. 2020. Mapping co-benefits for carbon storage and biodiversity to inform conservation policy and action. *Philosophical Transactions of the Royal Society B* 375: 20190128.
- Spathelf, P., J. Stanturf, M. Kleine, et al. 2018. Adaptive measures: Integrating adaptive forest management and forest landscape restoration. *Annals of Forest Science* 75: 55.
- Spies, T.A., P.A. Stine, R.A. Gravenmier, J.W. Long, and M.J. Reilly. 2018. Synthesis of Science to Inform Land Management within the Northwest Forest Plan Area. General Technical Report PNW-GTR-966. Portland, OR: U.S. Department of Agriculture, Forest Service, Pacific Northwest Research Station.
- Spittlehouse, D.L. 1997. Forest management and climate change. p 24-1 to 24-8. In: E. Taylor and B. Taylor, eds. *Responding to Global Climate Change in British Columbia and Yukon*. Vancouver, BC: Environment Canada.
- Spittlehouse, D.L., and R.B. Stewart. 2003. Adaptation to climate change in forest management. *BC Journal of Ecosystems and Management* 4: 1–11.
- Stanturf, J.A. 2015. Future landscapes: Opportunities and challenges. *New Forests* 46: 615–644.
- Stanturf, J.A., P. Madsen, K. Sagheb-Talebi, and O.K. Hansen. 2018. Transformational restoration: Novel ecosystems in Denmark. *Plant Biosystems – An International Journal Dealing with all Aspects of Plant Biology* 152: 536–546.
- Stanturf, J.A., B.J. Palik, M.I. Williams, R.K. Dumroese, and P. Madsen. 2014. Forest restoration paradigms. *Journal of Sustainable Forestry* 33: S161–S194.

- Stein, B.A., P. Glick, N. Edelson, and A. Staudt, eds. 2014. *Climate-Smart Conservation: Putting Adaptation Principles into Practice*. Washington, DC: National Wildlife Federation.
- Stein, B.A., A. Staudt, M.S. Cross, et al. 2013. Preparing for and managing change: Climate adaptation for biodiversity and ecosystems. *Frontiers in Ecology and the Environment* 11: 502–510.
- Stephanson, C.A., and N. Ribarik Coe. 2017. Impacts of beech bark disease and climate change on American beech. *Forests* 8: 155.
- Stephens, S.L., M.A. Battaglia, D.J. Churchill, et al. 2021. Forest restoration and fuels reduction: Convergent or divergent? *BioScience* 71: 85–101.
- Stephens, S.L., B.M. Collins, E. Biber, and P.Z. Fulé. 2016. US federal fire and forest policy: Emphasizing resilience in dry forests. *Ecosphere* 7: e01584.
- Stephens, S.L., B.M. Collins, C.J. Fettig, et al. 2018. Drought, tree mortality, and wildfire in forests adapted to frequent fire. *BioScience* 68: 77–88.
- Stevens-Rumann, C., K. Shive, P. Fulé, and C.H. Sieg. 2013. Pre-wildfire fuel reduction treatments result in more resilient forest structure a decade after wildfire. *International Journal of Wildland Fire* 22: 1108–1117.
- Stevens-Rumann, C.S., and P. Morgan. 2019. Tree regeneration following wildfires in the western US: A review. *Fire Ecology* 15: 15.
- St-Laurent, G.P., B. Locatelli, G. Hoberg, V. Gukova, and S. Hagerman. 2021a. Models for integrating climate objectives in forest policy: Towards adaptation-first? *Land Use Policy* 104: p.105357.
- St-Laurent, G.P., L.E. Oakes, M. Cross, and S. Hagerman. 2021b. R-RT (resistance-resilience-transformation) typology reveals differential conservation approaches across ecosystems and time. *Communications Biology* 4: 39. doi: 10.1038/s42003-020-01556-2
- Strassburg, B.B., A. Kelly, A. Balmford, et al. 2010. Global congruence of carbon storage and biodiversity in terrestrial ecosystems. *Conservation Letters* 3: 98–105.
- Strzepek, K., G. Yohe, J. Neumann, and B. Boehlert. 2010. Characterizing changes in drought risk for the United States from climate change. *Environmental Research Letters* 5: 044012.
- Sturrock, R.N., S.J. Frankel, A.V. Brown, et al. 2011. Climate change and forest diseases. *Plant Pathology* 60: 133–149.
- Sun, G., D.W. Hallema, E.C. Cohen, et al. 2019. Effects of Wildfires and Fuel Treatment Strategies on Watershed Water Quantity across the Contiguous United States. JFSP PROJECT ID: 14-1-06-18. Research Triangle Park, NC: Eastern Forest Environmental Threat Assessment Center.
- Swanson, F.J., J.A. Jones, D.O. Wallin, and J.H. Cissel. 1994. Natural variability—Implications for ecosystem management. p 80–94. In: M.E. Jensen and P.S. Bourgeron, eds. Volume II: Eastside Forest Ecosystem Health Assessment. General Technical Report GTR-PNW-318. Portland, OR: U.S. Department of Agriculture, Forest Service, Pacific Northwest Research Station.
- Swanston, C., L.A. Brandt, M.K. Janowiak, et al. 2018. Vulnerability of forests of the Midwest and Northeast United States to climate change. *Climatic Change* 146: 103–116.
- Swanston, C.W., M.K. Janowiak, L.A. Brandt, et al. 2016. *Forest Adaptation Resources: Climate Change Tools and Approaches for Land Managers*, 2nd ed. General Technical Report NRS-87-2. Newtown Square, PA: U.S. Department of Agriculture, Forest Service, Northern Research Station.
- Swetnam, T.W., C.D. Allen, and J.L. Betancourt. 1999. Applied historical ecology: Using the past to manage for the future. *Ecological Applications* 9: 1189–1206.
- Taylor, P.B. 2019. The good, the bad, and the unnecessary: Forest fire suppression funding and forest management provisions of the Consolidated Appropriations Act of 2018. *Public Land and Resources Law Review* 41: 79–103.
- Terando, A.J., B. Reich, K. Pacifici, et al. 2016. Uncertainty quantification and propagation for projections of extremes in monthly area burned under climate change: A case study in the coastal plain of Georgia, USA. p 245–256. In: K. Riley et al., eds. *Natural Hazard Uncertainty Assessment: Modeling and Decision Support*. Washington, DC: American Geophysical Union.
- Terskaia, A., R.J. Dial, and P.F. Sullivan. 2020. Pathways of tundra encroachment by trees and tall shrubs in the western Brooks Range of Alaska. *Ecography* 43: 769–778.
- Teskey, R., T. Wertin, I. Bauweraerts, et al. 2015. Responses of tree species to heat waves and extreme heat events. *Plant, Cell and Environment* 38: 1699–1712.
- Thom, D., and R. Seidl. 2016. Natural disturbance impacts on ecosystem services and biodiversity in temperate and boreal forests. *Biological Reviews* 91: 760–781.
- Thompson, L.M., A.J. Lynch, E.A. Beever, et al. 2021. Responding to ecosystem transformation: Resist, accept, or direct? *Fisheries* 46: 8–21.
- Thorn, S., C. Bäessler, R. Brandl, et al. 2018. Impacts of salvage logging on biodiversity: A meta-analysis. *Journal of Applied Ecology* 55: 279–289.
- Tian, X., B. Sohngen, J.S. Baker, S.B. Ohrel, and A.A. Fawcett. 2018. Will U.S. forests continue to be a carbon sink? *Land Economics* 94: 97–113.
- Timberlake, T., L.A. Joyce, C. Schultz, and G. Lampman. 2018. *Design of a Workshop Process to Support Consideration of Natural Range of Variation and Climate Change*

- for Land Management Planning under the 2012 Planning Rule. Resource Note RMRS-RN-82. Fort Collins, CO: U.S. Department of Agriculture, Forest Service, Rocky Mountain Research Station.
- Timberlake, T.J., and C.A. Schultz. 2017. Policy, practice, and partnerships for climate change adaptation on US National Forests. *Climatic Change* 114: 257–269.
- Timberlake, T.J., and C.A. Schultz. 2019. Climate change vulnerability assessment for forest management: The case of the U.S. Forest Service. *Forests* 10: 1030.
- Tohver, I.M., A.F. Hamlet, and S.Y. Lee. 2014. Impacts of 21st-century climate change on hydrologic extremes in the Pacific Northwest region of North America. *Journal of the American Water Resources Association* 50: 1461–1476.
- Toot, R., L.E. Frelich, E.E. Butler, and P.B. Reich. 2020. Climate-biome envelope shifts create enormous challenges and novel opportunities for conservation. *Forests* 11: 1015.
- Trần, J.K., T. Ylioja, R.F. Billings, J. Régnière, and M.P. Ayres. 2007. Impact of minimum winter temperatures on the population dynamics of *Dendroctonus frontalis*. *Ecological Applications* 17: 882–899.
- Treasure, E., S. McNulty, J.M. Myers, and L.N. Jennings. 2014. Template for Assessing Climate Change Impacts and Management Options: TACCIMO User Guide Version 2.2. General Technical Report SRS-GTR-186. Asheville, NC: U.S. Department of Agriculture, Forest Service, Southern Research Station.
- Truitt, A.M., E.F. Granek, M.J. Duveneck, et al. 2015. What is novel about novel ecosystems: Managing change in an ever-changing world. *Environmental Management* 55: 1217–1226.
- Turner, M.G., W.J. Calder, G.S. Cumming, et al. 2020. Climate change, ecosystems and abrupt change: Science priorities. *Philosophical Transactions of the Royal Society B* 375: 20190105.
- Urbanski, S. 2014. Wildland fire emissions, carbon, and climate: Emission factors. *Forest Ecology and Management* 317: 51–60.
- Urgenson, L.S., C.R. Nelson, R.D. Haugo, et al. 2018. Social perspectives on the use of reference conditions in restoration of fire-adapted forest landscapes. *Restoration Ecology* 26: 987–996.
- Urgenson, L.S., C.M. Ryan, C.B. Halpern, et al. 2017. Visions of restoration in fire-adapted forest landscapes: Lessons from the Collaborative Forest Landscape Restoration Program. *Environmental Management* 59: 338–353.
- Ury, E.A., X. Yang, J.P. Wright, and E.S. Bernhardt. 2021. Rapid deforestation of a coastal landscape driven by sea level rise and extreme events. *Ecological Applications*: e2339.
- USDA (U.S. Department of Agriculture). No date. USDA Climate Hubs. Washington, DC: USDA (accessed July 20, 2020). <https://www.climatehubs.usda.gov/about-us>
- USDA (U.S. Department of Agriculture). 2014. Strategic Plan FY 2014–2018. Washington, DC: USDA.
- USDA (U.S. Department of Agriculture). 2016. Building Blocks for Climate Smart Agriculture and Forest. Washington, DC: USDA.
- U.S. EPA (U.S. Environmental Protection Agency). No date. Greenhouse gas equivalencies calculator. Washington, DC: U.S. EPA (accessed January 15, 2021). <https://www.epa.gov/energy/greenhouse-gas-equivalencies-calculator>
- U.S. EPA (U.S. Environmental Protection Agency). 2021 Inventory of U.S. Greenhouse Gas Emissions and Sinks: 1990–2019. EPA 430-R-21-005. Washington, DC: U.S. EPA (accessed May 19, 2021). <https://www.epa.gov/ghgemissions/inventory-us-greenhouse-gas-emissions-and-sinks-1990-2019>
- USFS (U.S. Forest Service). No date. Forest Products Cut and Sold from the National Forests and Grasslands. Washington, DC: U.S. Department of Agriculture, Forest Service (accessed May 21, 2021). <https://www.fs.fed.us/forestmanagement/products/cut-sold/index.shtml>
- USFS (U.S. Forest Service). 2008. Forest Service Strategic Framework for Responding to Climate Change. Washington, DC: U.S. Department of Agriculture, Forest Service.
- USFS (U.S. Forest Service). 2011a. National Roadmap for Responding to Climate Change. Washington, DC U.S. Department of Agriculture, Forest Service.
- USFS (U.S. Forest Service). 2011b. Navigating the Forest Service Climate Change Performance Scorecard. Washington, DC: U.S. Department of Agriculture, Forest Service.
- USFS (U.S. Forest Service). 2012a. Increasing the Pace of Restoration and Job Creation on Our National Forests. Washington, DC: U.S. Department of Agriculture, Forest Service.
- USFS (U.S. Forest Service). 2012b. Code of Federal Regulations Title 36, Part 219, Subpart A: National Forest System Land Management Planning. *Federal Register* 77: 21162–21275.
- USFS (U.S. Forest Service). 2012c. Future of America’s Forest and Rangelands: Forest Service 2010 Resources Planning Act Assessment. General Technical Report WO-87. Washington, DC: U.S. Department of Agriculture, Forest Service.
- USFS (U.S. Forest Service). 2015a. FSH 1909.12—Land Management Planning Handbook. Washington, DC: U.S. Department of Agriculture, Forest Service.
- USFS (U.S. Forest Service). 2015b. The Rising Cost of Wildfire Operations: Effects on the Forest Service’s Non-fire Work. Washington, DC: U.S. Department of Agriculture, Forest Service.
- USFS (U.S. Forest Service). 2016. Ecosystem Restoration Policy. *Federal Register* 81: 24785–24793.

- USFS (U.S. Forest Service). 2017. Fiscal Year 2018 Budget Justification. Washington, DC: U.S. Department of Agriculture, Forest Service.
- USFS (U.S. Forest Service). 2018. Flathead National Forest Land Management Plan. Kalispell, MT: Flathead National Forest, U.S. Department of Agriculture, Forest Service.
- USFS (U.S. Forest Service). 2019. U.S. Forest Service Pacific Southwest Region forest health protection aerial detection survey. Redding, CA: U.S. Department of Agriculture, Forest Service (accessed January 27, 2021). [https://www.fs.usda.gov/detail/r5/forest-grasslandhealth/?cid=fsbdev3\\_046696](https://www.fs.usda.gov/detail/r5/forest-grasslandhealth/?cid=fsbdev3_046696)
- USFS (U.S. Forest Service). 2020. Sustainability Scorecard. Ogden, UT: U.S. Department of Agriculture, Forest Service, Intermountain Region (accessed March 1, 2021). <https://storymaps.arcgis.com/stories/b5412bfb7159446cb0a9da0b76dd64ef>
- Van Beusekom, A.E., N.L. Álvarez-Berrios, W.A. Gould, et al. 2018. Hurricane Maria in the US Caribbean: Disturbance forces, variation of effects, and implications for future storms. *Remote Sensing* 10: 1386.
- Vander Naald, B. 2020. Examining tourist preferences to slow glacier loss: Evidence from Alaska. *Tourism Recreation Research* 45: 107–117.
- van Kerkhoff, L., C. Munera, N. Dudley, et al. 2019. Toward future-oriented conservation: Managing protected areas in an era of climate change. *Ambio* 48: 699–713.
- Van Loon, A.F., T. Gleeson, J. Clark, et al. 2016. Drought in the Anthropocene. *Nature Geoscience* 9: 89–91.
- van Mantgem, P.J., N.L. Stephenson, J.C. Byrne, et al. 2009. Widespread increase in tree mortality rates in the western United States. *Science* 323: 521–524.
- van Mantgem, P.J., J.C. Nasmith, M. Keifer, et al. 2013. Climatic stress increases forest fire severity across the western United States. *Ecology Letters* 16: 1151–1156.
- VerWey, B.J., M.J. Taylor, T.S. Garcia, and D.R. Warren. 2018. Effects of a severe drought on summer abundance, growth, and movement of cutthroat trout in a western Oregon headwater stream. *Northwestern Naturalist* 99: 209–221.
- Vinyeta, K., and K. Lynn. 2013. Exploring the role of traditional ecological knowledge in climate change initiatives. General Technical Report PNW-GTR-879. Portland, OR: U.S. Department of Agriculture, Forest Service, Pacific Northwest Research Station.
- Vogler, K.C., A.A. Ager, M.A. Day, M. Jennings, and J.D. Bailey. 2015. Prioritization of forest restoration projects: Trade-offs between wildfire protection, ecological restoration and economic objectives. *Forests* 6: 4403–4420.
- Vose, J.M., C.F. Miniati, C.H. Luce, et al. 2016. Ecohydrological implications of drought for forests in the United States. *Forest Ecology and Management* 380: 335–345.
- Vose, J.M., D.L. Peterson, G.M. Domke, et al. 2018. Forests. p 232–267. In: D.R. Reidmiller et al., eds. *Impacts, Risks, and Adaptation in the United States: Fourth National Climate Assessment*, vol. II. Washington, DC: U.S. Global Change Research Program (accessed April 7, 2021). <https://nca2018.globalchange.gov/chapter/6/>
- Vose, J.M., D.L. Peterson, and T. Patel-Weynand. 2012. Effects of Climatic Variability and Change on Forest Ecosystems: A Comprehensive Science Synthesis for the US Forest Sector. General Technical Report PNW-GTR-870. Portland, OR: U.S. Department of Agriculture, Forest Service, Pacific Northwest Research Station.
- Vose, R.S., D.R. Easterling, K.E. Kunkel, A.N. LeGrande, and M.F. Wehner. 2017. Temperature changes in the United States. p 185–206. In: D.J. Wuebbles et al., eds. *Climate Science Special Report: Fourth National Climate Assessment*, vol. I. Washington, DC: U.S. Global Change Research Program (accessed April 7, 2021). <https://science2017.globalchange.gov/chapter/6/>
- Wagner, M.R., W.M. Block, B.W. Geils, and K.F. Wenger. 2000. Restoration ecology. *Journal of Forestry* 98: 22–27.
- Walker, B., C.S. Holling, S.R. Carpenter, and A.P. Kinzig. 2004. Resilience, adaptability, and transformability in social-ecological systems. *Ecology and Society* 9: 5.
- Walker, R.B., J.D. Coop, S.A. Parks, and L. Trader. 2018. Fire regimes approaching historic norms reduce wildfire-facilitated conversion from forest to non-forest. *Ecosphere* 9: e02182.
- Wall, G., ed. 1992. *Implications of Climate Change for Pacific Northwest Forest Management*. Waterloo, ON: Department of Geography, University of Waterloo.
- Wallingford, P.D., T.L. Morelli, J.M. Allen, et al. 2020. Adjusting the lens of invasion biology to focus on the impacts of climate-driven range shifts. *Nature Climate Change* 10: 398–405.
- Walpole, E.H., E. Toman, R.S. Wilson, and M. Stidham. 2017. Shared visions, future challenges: A case study of three Collaborative Forest Landscape Restoration Program locations. *Ecology and Society* 22: 35.
- Watson, J.E., T. Evans, O. Venter, et al. 2018. The exceptional value of intact forest ecosystems. *Nature Ecology and Evolution* 2: 599–610.
- Way, D.A., and R.A. Montgomery. 2015. Photoperiod constraints on tree phenology, performance, and migration in a warming world. *Plant, Cell and Environment* 38: 1725–1736.
- Wear, D.N., and J.W. Coulston. 2015. From sink to source: Regional variation in US forest carbon futures. *Scientific Reports* 5: 16518.

- Weed, A.S., M.P. Ayres, and J.A. Hicke. 2013. Consequences of climate change for biotic disturbances in North American forests. *Ecological Monographs* 83: 441–470.
- West, J.M., S.H. Julius, P. Kareiva, et al. 2009. U.S. natural resources and climate change: Concepts and approaches for management adaptation. *Environmental Management* 44: 1001–1021.
- Westerling, A.L. 2016. Increasing western US forest wildfire activity: Sensitivity to changes in the timing of spring. *Philosophical Transactions of the Royal Society B* 371: 20150178.
- Westerling, A.L., H.G. Hidalgo, D.R. Cayan, and T.W. Swetnam. 2006. Warming and earlier spring increase western US forest wildfire activity. *Science* 313: 940–943.
- Westley, P.A. 2020. Documentation of en route mortality of summer chum salmon in the Koyukuk River, Alaska, and its potential linkage to the heatwave of 2019. *Ecology and Evolution* 10: 10296–10304.
- Williams, C.A., H. Gu, R. MacLean, J.G. Masek, and G.J. Collatz. 2016. Disturbance and the carbon balance of US forests: A quantitative review of impacts from harvests, fires, insects, and droughts. *Global and Planetary Change* 143: 66–80.
- Williams, M.I., and R.K. Dumroese. 2013. Preparing for climate change: Forestry and assisted migration. *Journal of Forestry* 111: 287–297.
- Wilson, G., M. Green, J. Brown, et al. 2020. Snowpack affects soil microclimate throughout the year. *Climatic Change* 163: 705–722.
- Wise, R.M., I. Fazey, M.S. Smith, et al. 2014. Reconceptualizing adaptation to climate change as part of pathways of change and response. *Global Environmental Change* 28: 325–336.
- Wolken, J.M., T.N. Hollingsworth, T.S. Rupp, et al. 2011. Evidence and implications of recent and projected climate change in Alaska's forest ecosystems. *Ecosphere* 2: 1–35.
- Woodall, C.W., C.M. Oswalt, J.A. Westfall, et al. 2009. An indicator of tree migration in forests of the eastern United States. *Forest Ecology and Management* 257: 1434–1444.
- Wurtzebach, Z., R.J. DeRose, R.R. Bush, et al. 2020. Supporting national forest system planning with forest inventory and analysis data. *Journal of Forestry* 118: 289–306.
- Wurtzebach, Z., and C. Schultz. 2016. Measuring ecological integrity: History, practical applications, and research opportunities. *BioScience* 66: 446–457.
- Young, D.J., J.T. Stevens, J.M. Earles, et al. 2017. Long-term climate and competition explain forest mortality patterns under extreme drought. *Ecology Letters* 20: 78–86.
- Young, D.J., C.M. Werner, K.R. Welch, et al. 2019. Post-fire forest regeneration shows limited climate tracking and potential for drought-induced type conversion. *Ecology* 100: e02571.
- Zellmer, S.B., S.F. Bates, and J. Brown. 2018. Restoring beavers to enhance ecological integrity in National Forest planning. *Natural Resources and Environment* 33: 1–7.
- Zhu, K., C.W. Woodall, and J.S. Clark. 2012. Failure to migrate: Lack of tree range expansion in response to climate change. *Global Change Biology* 18: 1042–1052.



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